Feeding ecology of a stream fish assemblage in an Atlantic Forest remnant (Serra do Japi, SP, Brazil)

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This study aimed to characterize the trophic structure of the fish assemblage in streams of the Serra Japi, an ecotonal area between the Atlantic Forest and inland forests of São Paulo State, Southeastern Brazil. Fish were collected with electrofishing equipment in 15 sites covering different regions, substrate types and riparian vegetation, distributed throughout the Caguaçú River, Caxambú, Piraí and Guapeva River micro-basins, during the rainy (January/February) and dry season (June) of 2007. The 589 specimens analyzed from 22 species, were assigned to different trophic groups, discriminated through a matrix of similarity, based on the food index (IA_i). The results show the formation of seven groups with a predominance of insectivorous and omnivorous species, followed by detritivores, piscivores, omnivore-carnivores and herbivores, which consumed mostly items of autochthonous origin, where algae and young insects were dominant. Canonical correspondence analysis (CCA), correlating the biomass of trophic groups to environmental variables, showed that omnivores, insectivores and omnivore-carnivores displayed a wide distribution, while detritivores, herbivores and piscivores were restricted to specific locations, related to different physical and chemical variables as total nitrogen, conductivity and temperature. Despite the increase in total biomass at the most urbanized sites, the results indicate that the streams maintain a diverse community, suggesting that most of them are in preserved conditions.

O presente trabalho teve por objetivo caracterizar a estrutura trófica da comunidade de peixes de riachos da Serra do Japi, uma área de transição entre a Mata Atlântica e o interior paulista, no sudeste brasileiro. Os peixes foram coletados com equipamento de pesca elétrica em 15 pontos localizados em regiões com diferentes tipos de substrato e vegetação ripária, distribuídos pelas microbacias do Ribeirão Caguaçú, Caxambú, Piraí e Rio Guapeva, durante as estações chuvosa (janeiro/fevereiro) e seca (junho) de 2007. Os 589 indivíduos analisados, pertencentes a 22 espécies, foram classificados em diferentes grupos tróficos, discriminados através de uma matriz de similaridade baseada no Índice Alimentar (IA_i). Os resultados mostraram a formação de sete grupos com predomínio de espécies insetívoras e onívoras, seguidas pelas detritívoras, piscívoras, onívoro-carnívoras e herbívoras, que consumiram principalmente itens de origem autóctone com predomínio de algas e formas imaturas de insetos aquáticos. A Análise de Correspondência Canônica, ao correlacionar a biomassa dos grupos tróficos às variáveis ambientais, mostrou que as espécies onívoras, insetívoras e onívoro-carnívoras apresentaram ampla distribuição, enquanto que as detritívoras, herbívoras e piscívoras estiveram restritas a determinados locais, relacionados a diferentes variáveis físicas e químicas tais como nitrogênio total, condutividade e temperatura. Embora o aumento da biomassa de alguns grupos tróficos tenha sido verificado nos locais com maior influência antrópica, os resultados indicaram que os diferentes riachos mantêm uma fauna diversificada, sugerindo que estes se encontram em condições relativamente preservadas.

Key words: Tropical streams, Freshwater fish, Trophic ecology, Distribution.

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Introduction

The Atlantic Forest is one of the richest and most diverse ecosystems in the world, where it is considered one of the 25 hotspots of global biodiversity (Tabarelli et al., 2005). It is, however, one of the most threatened ecosystems (Myers et al., 2000), especially in the state of São Paulo, Southeastern Brazil, where forest cover has been reduced to about 10% (Castro & Menezes, 1998). One of the most prominent threats in this ecosystem is related to the degradation of natural rivers and streams, which has occurred mainly due the loss of riparian vegetation, organic chemical pollution and the introduction of exotic species, causing the destruction of the habitat and reducing the food supply (Menezes et al., 1990). Additionally, little information in tropical environmental areas exists, with inventories for one or more groups of organisms restricted to approximately 5% of the tropical protected areas (Hawksworth, 1995).

Streams in the Atlantic rainforest harbor a rich fish fauna, with a high degree of endemism, and intimate association with the riparian vegetation, considered important for food, shelter and reproduction (Menezes *et al.*, 1990; Menezes, 1994). Thus, several species represent an important link between the terrestrial and aquatic ecosystems (Gelwick & Matthews, 1996), as some of the food consumed originates in the riparian forest (Sabino & Castro, 1990).

Studies on fish communities in Atlantic Forest streams have been conducted in recent decades (Sabino & Castro, 1990; Buck, 1994; Uieda, 1995; Esteves, 1996; Uieda & Barreto, 1999; Vilella *et al.*, 2002), focusing on the composition, distribution and feeding habits of species. Although studies regarding the trophic structure of food chains in streams are common (Luiz *et al.*, 1998; Esteves & Lobon-Cerviá, 2001; Motta & Uieda, 2005; Uieda & Motta, 2007; Esteves *et al.*, 2008), they are still scarce in southeastern Brazil as compared to information related to rivers and reservoirs (Esteves & Aranha, 1999).

Considering that large scale alterations in forest structure may have serious impacts on stream biodiversity, affecting shading, discharge, siltation and availability of food types (Angermeier & Karr, 1983), an understanding of the trophic organization and food web structure can reveal fundamental properties of the stream ecosystem and help evaluate the influence of physical disturbance on the community and develop conservation strategies (Barreto & Aranha, 2006).

The Serra do Japi is part of some state environmental protection areas (APAs) for sustainable use, characterized by a high diversity of life forms, and was put under governmental trust in 1983. It represents one of the last areas of continuous forest of São Paulo State (Morellato, 1992), being considered an ecotonal region, i.e., in transition between two forest formations - the Atlantic Forest of Serra do Mar and the inland forests of São Paulo (Instituto Serra do Japi, 1998). Because of its size and significant degree of conservation, it is one of the best examples of semideciduous

forests of São Paulo, containing all the physiognomic components (Joly, 1992).

Little is known about the fish communities of the Serra do Japi streams, in spite of the growing impacts on these environments, resulting from increased removal of riparian vegetation, urban sprawl, and small dam construction. Focusing on streams under different conditions, this study aimed to examine the trophic ecology of the fish fauna, determining (1) the food spectrum of the different species, (2) the diversity of the trophic groups, and (3) the distribution of the trophic groups in relation to different environmental conditions.

Materials and Methods

Study area. The Serra do Japi is located to the west of the Atlantic Plateau, between the cities of Jundiaí, Cabreúva, Pirapora do Bom Jesus and Cajamar, in the state of São Paulo, in southeast Brazil (Fig. 1). It has an area of approximately 354 km², characterized by a quartzitic soil (Ab 'Saber, 1992), relatively flattened hilltops and very steep slopes, with altitudes ranging from 700 to 1219 meters (Instituto Serra do Japi, 1998). Vegetation is composed by semideciduous altitudinal forests, semideciduous mesophyllic forests, gallery forests and rocky substrates (Joly, 1992). The average annual temperature varies between 15.7 ° C and 19.2 ° C with an annual rainfall of about 1400 mm (Instituto Serra do Japi, 1998).

Preservation of the vegetated areas has occurred partly due to the establishment of the mountain ridges under governmental trust (Serra do Japi, Guaxinduva and Jaguacoara), located in the environmental protection areas (APAs) of Jundiaí, Cabreúva and Cajamar. However, since the early XVIII century the Serra do Japi has suffered several human interferences such as timber exploitation, agricultural and urban uses, real estate speculation and deforestation. Part of the area is presently occupied by *Pinus* spp. and *Eucalyptus* spp. reforestation, pastures and small agricultural properties (Instituto Serra do Japi, 1998).

The streams in the area are predominantly clear water, originating in mountain escarpments and comprising waterfalls, rapids and slow flowing stretches. The main microbasins considered for this study are part of the Sorocaba/Middle Tietê and Piracicaba-Capivari-Jundiaí sub-basins, both belonging to the Tietê River Basin, which is part of the Paraná River basin. Selection of sampling sites was based on the diversity of environmental conditions, represented both by preserved and deforested areas, as well as the existence of stream stretches with different types of substrate.

Fifteen sites distributed over different microbasins were sampled. In the Jundiaí APA, stretches of the Caguaçú River micro-basin, formed by São Jerônimo Stream (sampling site P1), its tributaries (P2, P9), Caguaçú River (P8 and P14), and the stream called here "Paiol Velho" (P7) - a small tributary of Cachoeira River - were sampled. Within this same APA, in the Guapeva River micro-basin, sites 4, 5 and 6 located in Pedreiras

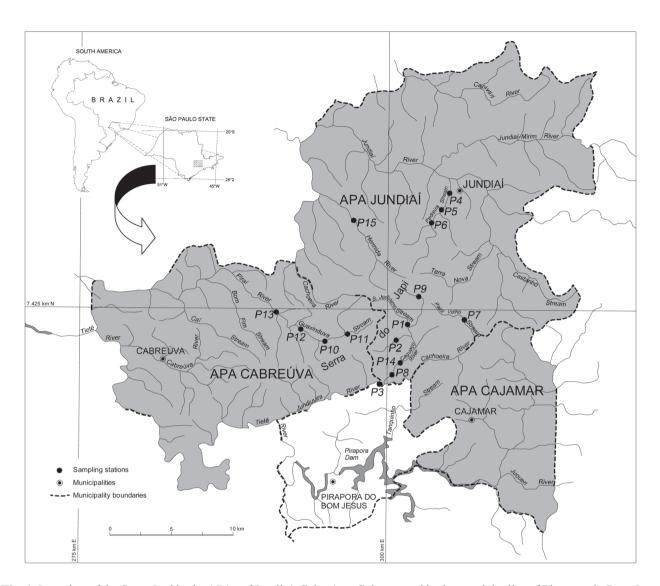


Fig. 1. Location of the Serra Japi in the APAs of Jundiaí, Cabreúva, Cajamar and in the municipality of Pirapora do Bom Jesus (SP), showing sampling sites in the different streams. Adapted from "Atlas das Unidades de Conservação Ambiental do Estado de São Paulo- Parte II-Interior, SMA, 1998".

Stream, and Hermida River in the Caxambu River micro-basin (P15) were selected. In the Cabreúva APA, sites P10, P11, P12, and P13 were located in Guaxinduva Stream (Piraí River micro-basin) and site P3 in Jundiuvira River, which belongs to the Caguaçú River micro-basin (Fig. 1).

Most of the stations were located in relatively preserved areas, with a predominance of native forest (P1, P2, P8, P10, P11, P12, and P14). Sites P9 and P15 were located in pasture areas and P3, P4, P5, P6, P7 and P13 in areas subject to anthropogenic influences, where riparian vegetation is disturbed and there is discharge of domestic sewage.

Data collection. Fish samples were obtained during the rainy (January/February) and dry season (June) of 2007 using stationary electrofishing equipment (HONDA EX 1000 generator, 120V, 60Hz, < 2.5 Amp., AC). This configuration was similar to that used by Mazzoni *et al.* (2000), who found

that sampling efficiency was not affected in an Atlantic Forest stream which presented varying conductivities. At each site three successive catches were conducted over a 50-m stretch, following the 3-catch removal method (Zippin, 1958), resulting in a constant fishing effort at each locality. Borders were blocked with nets of 1.3-mm mesh size to prevent fish from escaping. Samples were fixed in 10% formalin solution and preserved in 70% alcohol, and specimens were identified according to Britski (1972) and Britski *et al.* (1986), which was later confirmed by a specialist of the Museum of Zoology of the University of São Paulo, where a voucher was deposited.

The following environmental characteristics were recorded for each sampling site: altitude (m), depth (m), width (m), water speed (m.s⁻¹) - according to the float method of Marques & Argento (1988); flow (m³.s⁻¹), using the formulae Q= W x A x S, where Q = flow, W = width, A= area and S = water speed; riparian vegetation cover (%), recorded as the

percentage shading of the stream channel; proportion of rapids and pools (%) – relative proportion of each type in a 50m long stretch; predominant substrate, visually classified as the relative proportion of sand ($\simeq 3$ mm), fine gravel ($\simeq 3$ to 49 mm), coarse gravel (50 to 149 mm), small boulders (150 to 300 mm) or large boulders (> 300 mm), and presence of impoundments in the surroundings. Physicochemical parameters for each site, such as water temperature (°C), dissolved oxygen (OD) (mg. L⁻¹), pH, conductivity (µS.cm⁻¹) and total dissolved solids (mg.L-1) were measured in the field with a multi-parameter water quality monitoring system (HORIBA U-22). Other analyses included ammonia (mg.L⁻¹), dissolved phosphorus (µg.L⁻¹), total suspended solids (mg.L⁻¹), alkalinity (mg.L⁻¹) (APHA, 1989), nitrate (mg.L⁻¹) (CETESB, 1978), nitrite (µg.L⁻¹) (FEEMA, 1981), total nitrogen (mg.L⁻¹) and total phosphorus (ig.L⁻¹) (Strickland & Parsons, 1968).

Diet analysis. Sub-samples of 20 individuals for the most abundant species for each season were taken, while for the remaining ones, the total number was analyzed. Gut contents were analyzed according to the method of frequency of occurrence and points method (Hynes, 1950), and counted in a counting chamber under a light microscope and stereomicroscope (25X). For this latter method, the stomach was opened and given a score that varied from 0 to 20, related to the total food volume. Items were then grouped into different categories, and a number of points assigned to each one, according to the volume occupied in relation to the total volume. The volumetric proportion of each category was calculated, considering 100% as the sum of points assigned to each one.

Frequency of occurrence of each item was correlated with volumetric data using the feeding index (IA) proposed by

$$IA_i = \frac{F_i \times V_i}{\sum_{x=1}^{n} (F_i \times V_i)} \times 100$$

Kawakami & Vazzoler (1980) and adapted by Hahn *et al.* (1997b), and is described by the equation:

where i=1,2,...., n food items; F_i = frequency of occurrence (%) of a given food item; V_i = volume (%) of a given food item.

Diet similarity of fish species, based on the IA₁ was analyzed through a triangular similarity matrix between pairs of species, after square-root transformation. Based on this matrix, a dendrogram using Bray-Curtis distance and Ward's clustering method and ordinations by non-metric multidimensional scaling (NMDS) were obtained. NMDS is a nonparametric procedure that uses ranks of the similarities between samples to construct a map in which the positions of the samples reflect as closely as possible the dissimilarities among them. The stress value is a measure of the accuracy of this representation, with a value below 0.1 indicating a reliable representation (Clarke & Warwick, 1994).

After assigning fish assemblage to different trophic groups, the biomass for each one, considering both seasons was totaled, and a canonical correspondence analysis (CCA) was used to determine the relationships between the biomass of each trophic group and environmental variables including habitat characteristics and physicochemical water parameters. The environmental matrix was based on the average values between the dry and rainy season, and was standardized, while biological data were transformed into log (x+1). Some species may be classified into different trophic groups according to the resolution level adopted, and therefore insectivores and benthic insectivores were grouped for the biomass calculations, since we opted to place fish in feeding guilds that resembled them most, as has been adopted in other studies (Jeppesen *et al.*, 2000).

For this analysis, a stepwise backward selection procedure was performed to produce a parsimonious model. Starting with the full model including all explanatory variables, Akaike's Information Criterion (AIC) was used to assess the change in model performance resulting from excluding individual explanatory variables. At each step in this procedure, the change in AIC compared to the current model was evaluated and the model that resulted in the lowest AIC was selected. A correlation between the weighted average scores and linear combination scores was used and ANOVA was applied to test the significance of the model. All statistical analyses were carried out using the statistical computing software R (R Development Core Team, 2007) and the package vegan (Oksanen *et al.*, 2007).

Results

Environmental variables. General environmental characteristics of the different sampling zones are summarized in Table 1. The results indicated few differences among sampling sites, with most of them being characterized by the presence of rapids, high concentration of dissolved oxygen, small depth and width, and a predominance of sand and coarse gravel in the substrate. Percentage vegetation cover also varied, with open pasture areas (P9 and P15) and closed canopy sites (P1, P2, P8, P10, P11 and P14). Higher values of conductivity occurred at sites under anthropogenic influences (P4, P5, P6 and P7) and higher order streams (P3, P8), while discharge showed the highest values at P3 and P12.

Trophic group assignment. From a total of 2474 specimens belonging to 10 families and 29 species obtained during the dry and rainy seasons, 589 stomachs from 22 species were considered for the analysis. Major food types, grouped into 12 broad categories included algae, adult insects, young insects (larvae, nymphs and pupae), insect fragments (with predominance of exoskeleton), vegetal matter (seeds and leaf parts), organic matter (amorphous material of animal or plant origin), detritus, crustaceans, fish, nematodes, annelids and other (low occurrence items such as cladocera, copepods, ostracods, bivalves, fish scales, arachnids, and mites). The

importance of these categories in fish diet is shown through the results of the IA₂ (Table 2).

Analysis of the NMDS resulted in seven trophic groups, with a stress value of 0.05, indicating good ordination and confidence in the representation of the results (Fig. 2). These groups were characterized as follows:

Group 1: Benthic insectivores – This group was represented by *Trichomycterus* sp. (Trsp), *Characidium gomesi* (Chgo) and *Characidium oiticicai* (Choi), which consumed predominantly young insects of benthic origin (IA_i: 40-65%) (Table 2), with high occurrence of chironomidae, simuliidae, trichoptera and ephemeroptera (Table 3). Insect fragments also contributed with significant proportions in the diet of these species.

Group II: Insectivores – These included species such as *Astyanax paranae* (Aspa), *Astyanax altiparanae* (Asal), *Astyanax bockmanni* (Asbo), *Gymnotus carapo* (Gyca), *Bryconamericus stramineus* (Brstr), *Cetopsorhamdia iheringi* (Ceih), *Imparfinis mirini* (Immi) and *Trichomycterus brasiliensis* (Trbr), which represented 36.3% of all species. Insect fragments were the preferential food item (IA_i: 39-62%), followed by young insects (IA_i: 12–27%) composed of chironomidae, trichoptera, simuliidae and unidentified diptera larvae (Table 3).

Group III: Detritivores – This group was formed by three species (13.6% of species analyzed), composed of *Pareiorhina* sp. (Pasp), *Hisonotus depressicauda* (Hide) and

Hypostomus ancistroides (Hyan). These species exploited the bottom, ingesting large amounts of detritus (IA_i: 65-78%), followed by organic matter and algae, which were also important, as they were probably deposited on the sediment (IA_i: 5-16%).

Group IV: Herbivores – This group contained only *Tilapia* rendalli (Tire) (4.5% of species analyzed), which fed preferentially on plant fragments (IA_i: 72%), followed by algae (mainly diatoms) and organic matter. Insects, detritus and fish showed low occurrence.

Group V: Omnivores – These fish included species such as *Corydoras aeneus* (Coae), *Geophagus brasiliensis* (Gebr), *Neoplecostomus paranensis* (Nepa), *Phalloceros* sp. (Phsp) and *Poecilia reticulata* (Pore) (22.8% of species analyzed). Food habits included a variety of items such as algae, plant pieces, insect fragments, young insects, detritus, nematodes, oligochaeta and organic matter (Table 3). *Phalloceros* sp. and *P. reticulata* displayed the highest values of detritus (IA=40%) and organic matter (IA=36.3%).

Group VI: Piscivores – The group was composed of *Hoplias malabaricus* (Homa) (4.5% of species analyzed), where this species had a diet consisting exclusively of fish (IA_i=99%). Other items such as insect fragments, plant matter and organic matter were rare in the diet.

Group VII: Omnivores with a tendency to carnivory - Formed by *R. quelen*, which represented 4.5% of species that had

Table 1. Main characteristic and physicochemical water parameters (average values - rainy and dry season) in 15 sites of the Serra do Japi streams (APA Cabreúva and Jundiaí). DO= Dissolved oxygen; TDS= Total Dissolved Solids; s=sand; fg= fine gravel; cg= coarse gravel; sb= small boulder; lb= large boulder; nf= native forest; sf= secondary forest; Zyn= Zyngiberaceae (+) presence; (-) absence.

Parameters	P1	P2	P3	P4	P5	P6	P7	P8	P9	P10	P11	P12	P13	P14	P15
	São Jerônimo stream	Jerônimo	vira River	Pedreira stream	stream	Pedreira stream	Paiol Velho stream	Caguaçú River	Jerônimo	duva stream	duva stream	Guaxin- duva stream	Guaxin- duva stream	River	Hermida River
Coordinates	23°17'01.2" 46°56'21.4"					23°15'11.3" 46°54'27.9"								' 23°18'34.8'' ' 46°56'38.8''	
Dominant substrate Altitude (m)	890	s/fg/cg 830	s/cg 807	s/cg 858	s/cg 802	lb/sb 805	s/silt 804	s/cg 694	s/cg 927	s/cg 930	cg/sb 924	s/sb 854	s/cg 861	s/fg 798	s/fg 740
Order Riparian Vegetation	4 ^a nf	4 ^a nf Zyn	5 ^a sf Zyn	4 ^a sf, Zyn bamboo	4 ^a grass, sf Zyn	4 ^a Eucalyptus Zyn	3 ^a sf bamboo	5 ^a nf	4 ^a pasture	4 ^a nf	4 ^a nf	5 ^a Zyn	5 ^a grass	5 ^a nf Zyn	4 ^a pasture
PH	6.84	6.65	6.58	6.79	6.67	6.53	6.53	6.58	6.6	6.85	6.72	6.81	6.73	6.93	6.63
DO (mg.L ⁻¹)	13.31	13.77	11.72	13.16	11.08	11.08	11.09	12.01	12.11	12.02	12.09	12.66	12.66	12.17	12.82
Temp. (°C)	17.07	17.86	17.58	19.70	19.6	19.95	19.05	17.68	17.82	17.14	17.38	17.88	19.38	18.14	19.55
TDS (mg.L ⁻¹)	0.02	0.02	0.04	0.04	0.04	0.03	0.05	0.03	0.02	0.02	0.02	0.01	0.02	0.02	0.02
Conductivity (µS.cm ⁻¹)	36.00	29.00	61.50	56.5	56.5	43.00	79.5	42.5	35.5	28.00	29.50	19.00	26.00	27.50	22.00
Width (m)	3.23	1.76	7.93	3.40	2.89	2.89	3.06	4.95	1.35	3.12	2.79	5.50	3.73	4.39	2.14
Depth (m)	0.29	0.17	0.50	0.31	0.39	0.26	0.25	0.37	0.33	0.42	0.16	0.31	0.69	0.49	0.54
Canopy cover (%)	85.00	90.00	50.00	65.00	40.00	75.00	75.00	75.00	0.00	85.00	98.00	25.00	25.00	80.00	0.00
Water speed (m.s ⁻¹)	0.41	0.47	0.56	0.48	0.31	0.41	0.29	0.58	0.59	0.80	0.48	1.04	0.34	0.49	0.40
Discharge (m³s⁻¹)	1.29	0.19	13.20	1.20	0.99	0.39	0.51	4.80	0.17	2.86	0.57	6.41	2.89	3.98	1.40
Runs (%)	30	-	-	-	-	-	-	-	-	-	-	5	-	-	-
Riffles (%)	60	90	60	100	90	80	65	70	85	85	80	95	20	75	90
Shallow pools (%)	10	10	40		10	20	45	30	15	15	20	-	80	25	10
Trunks/roots (%)	2.5	9.0	8.0	5.0	5.0	2.5	2.0	15.0	0.0	8.0	4.0	4.0	3.0	10.0	3.0
Impoundment	-	-	-	+	+	+	-	-	+	+	+	-	-	-	+

their diet analyzed. This species consumed similar proportions of insect fragments (IAi = 23.5%), crustaceans (IAi = 23.3%) and organic matter (IAi = 24.5%). Other items such as fish, immature insects and annelids were also important in the diet of this species.

Insect fragments, detritus, plant pieces, organic matter, aquatic insects and algae (Bacyllariophyceae and Chlorophyceae) were the food items consumed most often by the fish assemblage (Table 3). Terrestrial resources were not commonly represented in the diet of most of the species, while plant pieces and organic matter occurred more frequently.

Relationships between trophic groups and environmental variables. Biomass of the different trophic groups indicated a predominance of omnivores (37%) and insectivores (28%), followed by piscivores (16%), omnivores with a tendency to carnivory (8%), detritivores (7%) and herbivores (4%). Distribution of total biomass among the sites indicated that P4, P5, P6 and P7 had the highest values of total biomass (Fig. 3).

Stepwise procedure selected the variables Total Nitrogen (AIC = 28.320); water velocity (AIC = 24.726); temperature (AIC = 23.482) and conductivity (AIC = 22.073). The canonical correspondence analysis (CCA) between the trophic groups and selected environmental variables (p>0.005), explained 68% of the total variance (62% on the first two axes), indicating that total nitrogen, temperature and conductivity in axis 1 and total nitrogen in axis two, were the most important factors in the distribution of the trophic groups (Table 4). Figure 4 shows that omnivores, insectivores and omnivores with a tendency to carnivory were widely distributed in the study area, while detritivores, piscivores and herbivores were associated with specific variables. Detritivore biomass was nega-

tively correlated with temperature and conductivity, and positively with water velocity. Herbivores showed a strong correlation to total nitrogen and low water velocity at an open area pasture site (P9), while piscivores were associated with higher temperature and conductivity values (P3, P4, P6 and P7).

Discussion

The biotic composition of streams is strongly influenced by physical habitat. Important aspects of habitat structure include channel morphology, water depth, water velocity and flow, cover and substratum composition (Gorman & Karr, 1978; Schlosser, 1990). Riparian buffer zones also have an important role in linking a stream to its watershed, once they may control within stream communities by influencing solar radiation, water temperature, nutrient budgets, stabilization of stream banks, also providing a source of organic energy (Schiemer & Zalewski, 1992).

Habitats of Serra do Japi streams presented a high variety of environmental conditions with clear water where dense native riparian forest is found, as well as degraded sites. Human settlement has probably impacted these areas, altering local conditions through impoundment construction, sewage discharge and removal of riparian forest. Several studies show that fish communities respond to the effects of human activities, which may affect diversity, richness, abundance, growth and trophic organization (Karr, 1981; Rabeni & Smale, 1995; Paul & Mayer, 2001).

The streams were numerically dominated by fishes with relatively generalized diets (insectivores and omnivores), followed by more specialized fish (detritivores, herbivores and piscivores). The predominance of insectivorous species is

Table 2. Feeding Index (IA₁), which combines frequency of occurrence (%) and volume (%) of food items consumed by the different fish species of Serra do Japi streams (dry and rainy season). Preferential food items are marked in bold.

Species	Code	Voucher	Algae		Young insects	Insect frag- ments	Vegetal	Organic matter	Detritus	Crusta- cea	Fish	Nemato- da	Anneli- da	Other
Astyanax bockmanni	Asbo	98806	0.55	_	10.24	41.97	6.67	38.90	1.34	_	0.17	0.14	_	0.02
Bryconamericus stramineus	Brst	98809	2.97	_	26.39	50.08	7.08	12.51	0.97	_	-	-	_	-
Cetopsorhamdia iheringi	Ceih	98818	0.01	0.10	25.10	56.78	0.58	17.28	0.15	_	_	0.01	_	_
Characidium gomesi	Chgo	98815	0.00	0.47	65.67	28.46	0.63	4.39	0.35	_	_	-	_	0.01
Characidium oiticicai	Choi	98814	-	0.01	47.87	46.44	0.12	5.25	0.08	-	-	_	-	0.22
Corydoras aeneus	Coae	98828	2.67	-	9.12	14.06	3.10	56.90	9.24	-	-	1.39	-	3.51
Geophagus brasiliensis	Gebr	98829	0.93	-	27.88	8.73	7.73	48.32	4.85	0.00	0.14	0.01	0.21	1.20
Gymnotus carapo	Gyca	98833	0.04	0.01	13.11	39.43	10.33	34.54	1.20	0.44	0.20	0.56	0.07	0.07
Imparfinis mirini	Immi	98817	0.04	0.03	19.33	53.02	3.67	23.18	0.01	0.13	-	-	0.24	0.35
Trichomycterus brasiliensis	Trbr	98826	0.01	0.04	12.59	62.01	3.51	16.21	3.37	-	-	0.68	1.53	0.05
Trychomytcerus sp.	Trsp	98824	-	-	39.54	43.89	0.03	14.63	1.66	-	0.06	0.20	-	0.00
Astyanax altiparanae	Asal	98805	0.14	1.98	31.29	15.99	21.52	26.25	0.35	0.05	-	1.42	0.35	0.66
Astyanax paranae	Aspa	98807	0.54	0.25	10.10	26.70	36.95	24.45	0.03	-	-	0.32	-	0.66
Neoplecostomus paranensis	Nepa	98821	20.14	-	27.96	2.27	4.02	20.97	24.63	-	-	-	-	-
Hisonotus depressicauda	Hide	98819	5.49	-	-	-	3.95	25.22	65.44	-	-	-	-	-
Hypostomus ancistroides	Hyan	98820	5.42	-	0.08	0.01	6.31	20.33	67.84	-	-	0.00	-	0.02
Phalloceros sp.	Phsp	98831	1.41	0.01	0.79	2.90	8.41	49.92	36.45	0.05	-	-	-	0.05
Poecillia reticulata	Pore	98832	9.71	-	2.68	9.89	1.24	40.11	36.35	-	-	-	-	-
Pareiorhina sp.	Pasp	98822	15.82	-	0.23	0.00	2.63	3.04	78.29	-	-	-	-	0.00
Hoplias malabaricus	Homa	98816	-	-	-	0.13	0.33	0.09	0.14	-	99.32	-	-	-
Rhamdia quelen	Rhqu	98827	0.01	0.05	10.66	23.47	1.41	24.56	0.50	23.34	6.80	-	9.07	0.14
Tilapia rendalli	Tire	98830	8.67	-	0.22	1.12	72.29	11,93	2.37	-	2.47	-	-	0.93

Table 3. Frequency of occurrence (%) of food items of 22 fish species of serra do Japi streams (rainy and dry season). Species codes in Table 2. undet. = undetermined.

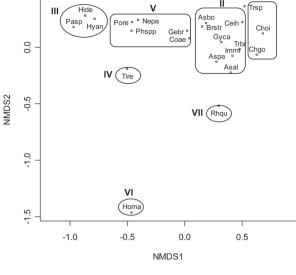
Species	Asal	Asbo	Aspa	Brstr	Ceih	Chgo	Choi	Coae	Gebr	Gyca	Hide	Homa	Hyan	Immi	Nepa	Pasp	Phsp	Pore	Rhqu	Tire	Trbr	Trsp	Tota (%)
AQUATIC ORIGIN													• •										
Chironomidae	10.2	8.2	4.5	16.7	18.6	18.2	12.2		16.0	12.6			2.8	13.6	7.8	2.9	2.4	3.3	5.0		12.2	19.7	9.0
Ceratopogonidae					1.4			1.3	0.5	3.6				2.9					2.5			0.7	0.6
Culicidae						0.6	2.0												2.5				0.2
Diptera (undet.)	1.6	2.0	1.9	6.7		1.3		1.3		4.3				1.9	1.4			1.3	2.5		2.7	2.2	1.4
Simuliidae	8.7	6.1	6.5	3.3	10.0	19.5	8.8	2.1	1.9	4.3				2.9	9.2	1.2	1.2	3.3	2.5	2.1	5.4	5.1	4.7
Ephemeroptera	1.6		0.6	3.3	2.9	6.5	9.5								2.8		0.6				1.4	2.2	1.4
Plecoptera						3.2	1.4								0.7							0.7	0.3
Coleoptera	1.6		0.6				0.7												2.5				0.2
Trichoptera	5.5		5.2	3.3	14.3	6.5	15.5	0.8	6.3	7.9				10.7	2.1		1.2		2.5		4.1	10.2	4.4
Odonata		2.0	0.6			0.6	1.4		1.0	2.4				1.0					2.5				0.5
Cladocera								0.4	0.5							0.6	0.6						0.1
Copepoda								1.7	1.5				1.1									0.7	0.2
Ostracoda	0.8							5.0	3.4					2.9			0.6				1.4	0.7	0.7
Aeglidae										0.8				1.0					7.5				0.4
Isopoda	1.6								0.5	1.6				1.0			1.2						0.3
Decapoda														1.0					5.0				0.3
Nematoda	5.5	4.1	3.2	3.3	2.9			3.8		2.8			0.6								4.1	3.6	1.5
Oligochaeta	1.6								1.0	0.4				1.0					5.0		2.7		0.5
Acarina						1.3	3.4	0.8	1.0														0.3
Tecameba								4.2	1.0	2.8			0.6								2.7		0.5
Osteichthyes		2.0								0.8		62.5							7.5	4.2		0.7	3.5
Scale	0.8		1.9				0.7		0.5	1.2										4.2			0.4
Bacillariophyceae	3.9	2.0	2.6	3.3				12.1	7.3	1.2	20.0		18.8	1.9	17.0	21.8	11.4	16.7		18.8	1.4		7.2
Chlorophyceae	3.1	4.1	3.2	3.3	1.4	0.6		7.1	5.8	2.0	6.7		11.6	1.9	12.1	8.2	7.2	10.0		12.5			4.6
Cyanophyceae			1.9	6.7				2.9	0.5	0.4	13.3		9.4		4.3	15.3	3.6	2.0		12.5			3.3
Bivalvia																			2.5				0.1
TERRESTRIAL ORI	GIN																						
Coleoptera (adult)					1.4	0.6	0.7			0.4				1.0					2.6				0.3
Diptera	0.8					0.6	•••																0.1
Hymenoptera	4.7		1.3														0.6						0.3
Lepidoptera	,		1.0														0.0						
(lmature)																							0.0
Arachnidae	0.8		0.6											1.0									0.1
INDETERMINATE																							
Insects Fragments	13.4	24.5	19.5	20.0	21.4	20.9	20.8	13.4	10.2	15.4		12.5	1.1	22.3	5.0	1.2	9.6	11.3	17.5	6.3	25.7	24.1	14.3
Vegetal Matter	13.4	12.2	18.8	10.0	4.3	5.2	2.7	7.9	11.2	13.0	20.0	12.5	16.6	8.7	12.1	14.1	14.4	8.0	7.5	18.8	4.1	0.7	10.7
Seeds	3.1		1.9	3.3				0.4	0.5	0.4				1.0									0.5
Organic matter	14.2	24.5	22.1	13.3	18.6	11.0	13.5	10.0	18.9	15.4	20.0	12.5	17.7	21.4	12.8	11.8	22.8	22.0	20.0	12.5	18.9	19.0	16.9
Detritus	3.1	6.1	1.3	3.3	2.9	3.2	2.0	12.1	9.2	5.9	20.0	12.5	19.3	1.0					5.0		13.5		9.9
Eggs		2.0	1.3				0.7	0.4	1.0								0.6						0.3
Stomaches analysed	22	12	40	6	17	39	40	40	40	40	6	6	39	25	25	40	40	32	16	10	22	22	589

often observed in streams of tropical regions (Uieda *et al.*, 1997; Castro & Casatti, 1997; Luiz *et al.*, 1998; Buck, 2000; Casatti *et al.*, 2001). Aquatic invertebrates, mainly insects in various stages of development, are also common in the diet of species of freshwater fish (Russo *et al.*, 2002). The insectivorous species of Serra do Japi streams used large proportions of aquatic insects, including groups which may be indicative of the good state of preservation of some sampling sites (trichoptera and ephemeroptera nymphae), given the basic requirements needed for their occurrence in highly oxygenated waters (McCfferty, 1981).

Observed diet differences between the group of insectivores composed by *C. gomesi*, *C. oiticicai* and *Trichomycterus* sp. which fed predominantly on benthic insect larvae and the group composed of I. *mirini*, *C. iheringi* and *T. brasiliensis* may be probably related to the use of different micro-habitats, periods of activity and feeding tactics,

commonly observed in tropical regions (Moyle & Senanayake, 1984; Sabino & Castro, 1990; Esteves & Galetti Jr., 1995). Some morphological characteristics of *T. brasiliensis* such as the presence of a cephalic intestine, adapted to a diversified diet and the presence of gill rakers as well, which may retain larger prey and allow the species to ingest animal food including fish (Ribeiro & Menin, 1996), may also explain the diet differences in relation to *Trichomycterus* sp.

Among the Loricariidae, three species were considered detritivores, occurring in habitats with a predominance of gravel and sand (*H. ancistroides* and *H. depressicauda*) and boulders and rapids (*Pareiorhina* sp.). The presence of algal tufts in the diet of this species suggests the use of grazing tactics similar to other armored catfishes such as *Kronichthys heylandii*, *Harttia kronei* and *Schizolecis guentheri* which occupy the rocky and sandy bottom of Atlantic Forest streams and graze basically on algae (Buck & Sazima, 1995). As ob-



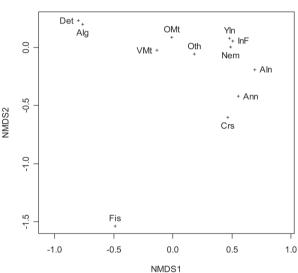


Fig. 2 a-b. Scores of NMDS for the fish species (a) and food items (b) along the axes 1 and 2. Circles and rectangles (a) indicate trophic groups formed by the similarity array. Benthic insectivores (I), insectivores (II), detritivores (III), herbivores (IV), omnivores (V), piscivores (VI), omnivore-carnivores (VII). (b) Alg = algae; Det = detritus; OMt = organic matter; VMt = vegetal matter; Oth = others, YIn = young insects; Fis=fish; AIn = adult insects; InF = insect fragments, Nem= nematodes; Ann = Annelidae, Crs = Crustacea. Codes of species are shown in Table 2.

served by Garavello & Santana (1998) for streams of the Paraiba do Sul and Tietê basins, *Pareiorhina* sp. seems to be associated with clear water environments, with sediments composed of sand and large boulders, feeding on algae of the periphyton.

Analysis of the origin of the food items revealed that aquatic-origin resources, especially immature insects and algae were the most important food resources for the fish assemblage, similar to findings of other studies (Sabino & Castro, 1990; Uieda *et al.*, 1997; Casatti, 2002; Deus & Petrere

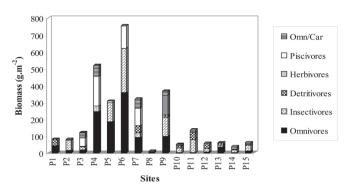


Fig. 3. Biomass (g.m⁻²) of the different trophic groups of fish at each collecting site in Serra do Japi (SP) streams.

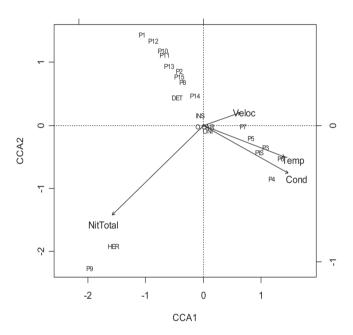


Fig. 4. Canonical Correspondence Analysis (CCA) showing the relationship between the biomass of insectivores (INS), omnivores (ONI), herbivores (HER), detritivores (DET), piscivores (PIS), omnivores-carnivores (O.CAR) and selected environmental variables. Temp = temperature; Veloc = Water Velocity; T.Nit = total nitrogen, Cond= Conductivity.

Jr., 2003) in southeastern brazilian streams. In other tropical rainforest streams, however, the importance of allochthonous food items has been highlighted, as these represent an important input of organic matter into the system (Henry *et al.*, 1994), being considered the basis of the food chain in some stream ecosystems (Gregory *et al.*, 1991). These apparently contradictory results can be explained by the relative importance of food of terrestrial and aquatic origin, the composition of the fish assemblage, the specific adaptations for feeding on aquatic organisms, as well as by the environmental

Table 4. Results of Canonical Correspondence Analysis (CCA) on fish guild biomass and selected environmental variable. Variable/ Axis = Biplot scores for constraining variables.

Variable/Axis	1	2	3	4
Temperature	0.65	-0.23	0.58	0.23
Total Nitrogen	-0.73	-0.65	0.07	0.19
Water Velocity	0.30	0.09	0.03	0.95
Conductivity	0.67	-0.35	-0.47	0.44
Axis				
Eigenvalues	0.21	0.16	0.02	0.01
Percentual cumulative variation	0.35	0.62	0.67	0.68
Species/Variable correlation	0.95	0.93	0.53	0.39

conditions of the streams studied (Melo et al., 2004; Dias et al., 2005). In our study, many of the species seemed to be morphologically adapted to bottom feeding. This behavior, associated with a more intense growth of algae in several sites were forest canopy was less dense (allowing sunlight to reach the water), could explain the predominance of authorhthonous items in the fishes diet.

The analysis of the trophic structure of the fish assemblage of the Serra do Japi indicated the existence of seven trophic groups, similar to those described by Esteves & Lobón-Cerviá (2001) in Atlantic Forest streams in southeastern Brazil and higher than in other stream studies (Sabino & Castro,1990; Casatti, 2002). In the Paraná River basin, the number of trophic guilds varied from 7 to 10 (Hahn *et al.*, 1997a; Luz-Agostinho, 2006). The comparison of the number of trophic groups within communities is often hampered by several factors, and must consider the diversity based on resources available, the number and morphological diversity of species present, the availability of prey in the environment (Pusey *et al.*, 1995), and the use of different methods to separate the guilds as well, which suggests the need to standardize these methods in order to make comparisons possible.

Trophic organization seems to be related to habitat variables, as shown by Chipps et al. (1994), Rabeni & Smale (1995) and Zambrano et al. (2006). Generally, stream flow has been a major influence on fish communities like many other physical variables as substratum, depth, presence of debris, habitat heterogeneity and physical-chemical factors such as temperature and oxygen (Schlosser, 1990). In Serra do Japi streams, according to the canonical correspondence analysis (CCA), the most important variables in the distribution of the different trophic groups were total nitrogen, temperature, conductivity and temperature. Although a direct correlation between selected variables and the presence of certain trophic groups cannot be directly established, the presence of the herbivore group (T. rendalli) was probably related to the transportation process from local fish ponds rather than to the high total nitrogen values (P9), caused by the input of animal excrements.

Several studies show that some fish species are particularly associated with good physical habitat conditions (Casatti *et al.*, 2006b), as was observed for the detritivore group formed by *H.ancistroides*, *Pareiorhina* sp. and

H.depressicauda, which showed a preference for preserved conditions. Power (1990) and Antoniassi et al. (1998) observed the environmental preferences of Hypostomus species for clean and running waters of brazilian rivers, while Garavello & Santana (1998) observed similar requirements for Pareiorhina sp. These species are probably negatively affected by alterations such as removal of the vegetation or siltation due to the embeddedness of the rocky substrate.

On the other hand our study showed that a higher piscivore biomass was associated to the more impacted sites (P4, P5, P6, P7), which can be attributed to the presence of large piscivores such as *H.malabaricus*. Since a higher biomass of other trophic groups as insectivores and omnivores was recorded at these sites and considering that other habitat features such as the presence of hiding places including trunks and roots was similar to the other areas, one can suggest that piscivores occurred in the most impacted sites because of greater food availability. The combination of higher temperatures, conductivity and total phosphorus at these sites (Rolla, 2008) may have contributed to an increase in productivity, as also suggested by Treviño et al. (1969) and Paul & Meyer (2001) which have shown that nutrient loads can stimulate fish production. In North American streams Scott et al. (1986) found that fish grew more rapidly and to larger sizes in urban sites, increasing fish production when compared to forested sites, presumably as a result of warmer temperatures and greater invertebrate biomass in the urban stream.

Considering that a loss of habitat quality results in a change in trophic structure, with dominance of generalist species (Casatti *et al.*, 2006a), the diversity of feeding groups found in this study seems to indicate that the fish fauna in the Serra do Japi streams is maintained due to the physical integrity of most of the sites studied. However, fish assemblage in the most disturbed sites showed a different structure, suggesting that loss of stream integrity with the spread of low-density levels of urbanization may have occurred. As changes in fish assemblages do not occur immediately following impacts such as urbanization, but may take many years with gradual replacement of species (Fitzgerald *et al.*, 1998), our results reinforce the importance of promoting the conservation of riparian vegetation and the control of sewage in the region, in order to maintain the integrity of stream habitats and its fish fauna.

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