Revista Brasileira de Ciência do Solo

Division - Soil Use and Management | Commission - Soil and Water Management and Conservation

Forest harvest management systems and residual phytomass affecting physical properties of a sandy soil

Karla Nascimento Sena^{(1)*} (D), Kátia Luciene Maltoni⁽²⁾ (D), Maria Júlia Betiolo Troleis⁽¹⁾ (D) and Glaucia Amorim Faria⁽²⁾ (D)

⁽¹⁾ Universidade Estadual Paulista "Júlio de Mesquita Filho", Escola de Engenharia, Programa de Pós-Graduação em Agronomia, Ilha Solteira, São Paulo, Brasil.

⁽²⁾ Universidade Estadual Paulista "Júlio de Mesquita Filho", Escola de Engenharia, Ilha Solteira, São Paulo, Brasil.

ABSTRACT: Organic carbon introduced in soils, mainly through organic matter, has a relevant role in various soil properties and is particularly important in sandy soils. In these soils, the input of organic material is necessary to ensure the sustainability of production systems. This study aimed to investigate the changes in total organic carbon content and its effect on physical properties in areas under different harvest management systems (HMS) after the harvest of eucalyptus. The study was performed in December 2017 in a Eucalyptus urograndis (clone E13) commercial plantation, in the municipality of Agua Clara, Mato Grosso do Sul State, Brazil. The soil of this area was classified as a sandy-textured Neossolo quartzarênico, which corresponds to Quartzipsamments. Soil samples were taken from the 0.00-0.05, 0.05-0.10 and 0.10-0.20 m layers for determinations of aggregate stability, soil bulk density (BD), macroporosity (Macro), microporosity (Micro), total porosity (TP) and total organic carbon (TOC); and for calculation of carbon stock (CS). Total organic carbon and CS continued down into the 0.20-0.40, 0.40-0.60, 0.60-0.80, and 0.80-1.00 m layers. Soil mechanical penetration resistance (PR) was determined to the 0.40 m depth in 0.10 m intervals. Carbon content was evaluated in the aggregates of the 0.00-0.05 m layer after wet sieving in 2000, 1000, 250 and 53 µm diameter sieves. Statistical evaluation consisted of analysis of variance, the Tukey test, and regression for the sources of variation that showed significance at 5 %. The data suggest that keeping the residual phytomass on the soil surface can positively impact total organic carbon, with a smaller reduction under the cut-to-length harvest management system. However, carbon stock is greater at the layer of 0.20-0.60 m; as the soil has a sandy texture, carbon moves through the soil profile, which has lower soil mechanical penetration resistance at the surface layers (0.00-0.10 m), once more under the cut-to-length system. Maintaining crop residual phytomass on the soil surface in the cut-to-length harvest management system provides better soil physical conditions, with greater macroporosity (0.00-0.05 m), aggregates with more carbon, and lower soil mechanical penetration resistance compared to systems that maintain only part of the harvest residual phytomass or no residual phytomass on the surface.

* Corresponding author: E-mail: karlla_senna@hotmail.com

Received: January 04, 2021 **Approved:** June 09, 2021

How to cite: Sena KN, Maltoni KL, Troleis MJB, Faria GA. Forest harvest management systems and residual phytomass affecting physical properties of a sandy soil. Rev Bras Cienc Solo. 2021;45:e0200190. https://doi.org/10.36783/18009657rbcs20200190

Editors: José Miguel Reichert () and Luciano da Silva Souza ().

Copyright: This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided that the original author and source are credited.



Keywords: carbon stock, aggregates, physical quality, *Eucalyptus* sp., organic matter.

Sena et al. Forest harvest management systems and residual phytomass affecting physical...



INTRODUCTION

Brazil has 9.0 million hectares of planted forests, 6.97 million hectares of which are of eucalyptus, and the state of Mato Grosso do Sul has 1.12 million hectares of eucalyptus growing on sandy soils, which predominate in that region (IBÁ, 2020). Sandy soils often have weak structure, low water retention, high permeability and sensitivity to compaction, low field capacity, and low organic carbon content and cation exchange capacity (Laclau et al., 2010; Huang and Hartemink, 2020). For sustainable use of sandy soils, their physical and chemical properties must be preserved through reduced tillage or no-till systems, crop residue management, carbon inputs, decreases in losses of soil nutrients (carbon, nitrogen, phosphorus, etc), and increase in the amount of macroaggregates and total porosity (Laclau et al., 2010; Six and Paustian, 2014; Du et al., 2015).

Forest areas are important in the economic sector and play a key role in the global carbon cycle. Eucalyptus plantations have been indicated as a valuable practice for carbon sequestration when associated with conservationist silvicultural practices in which soil turnover is reduced, residual phytomass is maintained, and nutrient cycling is promoted (Du et al., 2015; Rocha et al., 2018). Carbon plays an essential function in soil formation and maintenance of soil properties when incorporated in the soil. Soil contains more carbon than the total amount present in plants and the atmosphere, thus constituting a significant carbon reserve and effective stabilizer (Schmidt et al., 2011; Guan et al., 2015), which contributes to the mitigation of the greenhouse gas effect.

The impacts of soil use and management on soil physical quality have been quantified through physical properties related to soil structural stability, such as, aggregate stability. This property is dependent on soil mineralogy, soil particle size, and the presence of organic matter. At the same time, aggregate stability performs a significant function in the physical protection of organic carbon by enabling its occlusion within structural units and reducing its microbiological decomposition or its biodegradation, thus preserving organic carbon in the soil (Six and Paustian, 2014; Rocha et al., 2018; Vicente et al., 2019).

In the West Central region of Brazil, eucalyptus is normally grown on areas of degraded pasture, where it contributes by providing organic material during cultivation and harvest, when plant residual phytomass is left on the soil surface. Harvesting methods may lead to considerable differences in the amounts of organic matter deposited on the soil surface. The methods most used in this region are tree-length and cut-to-length harvest systems (Machado, 2008). The tree-length (TL) system generates tree-length logs; that is, the tree is semi-processed (delimbing and topping) in the area where the tree is felled, and the trunk plus branches, of more than 7-meter length, is taken to the roadside, where the processing of the logs is concluded (Machado, 2008).

The cut-to-length (CTL) method is characterized by processing the tree at the location where it is felled; delimbing, bark removal, and cutting trees into logs of pre-defined size are carried out at that location, leaving plant residual phytomass on the area (Machado, 2008). When this residual phytomass remains on the surface, it ensures a certain soil cover and maintains a large amount of plant material on the surface, initially forming a mulch and then plant litter.

The nutrients in the forest are returned to the soil through plant litter, which will be transformed and incorporated into the soil in the form of mineralized organic matter; the nutrients are then released for plant uptake (Krishna and Mohan, 2017; Rocha et al., 2018). The increase in organic matter input promotes soil aggregation and reduces exposure of carbon, which remains within soil aggregates, contributing to the improvement of soil structural stability, increase in total porosity, and microbial activity, among other benefits (Six and Paustian, 2014; Rocha et al., 2018; Vicente et al., 2019).



The CTL harvest management system (HMS) leads to greater input of organic matter than other practices, suggesting that it may promote conservation of the physical properties of soils of low structural stability, such as sandy soils, through the cycling of nutrients by residual phytomass left on the soil after harvest. This hypothesis supposes that the CTL, with a greater amount of material left in the area, may reduce losses of carbon/organic matter and preserve soil physical properties due to the residual harvest phytomass maintained in the area.

This study aimed to evaluate changes in organic carbon content and stock, and the changes in some physical properties of a sandy soil in areas under different harvest management practices of eucalyptus in the east of the state of Mato Grosso do Sul, Brazil.

MATERIALS AND METHODS

Study area

The study was performed in a *Eucalyptus urograndis* (clone E13) commercial plantation in the municipality of Água Clara, Mato Grosso do Sul State, Brazil (Figure 1). The climate in the region is tropical with dry winter (Aw), according to the Köppen classification system (Alvares et al., 2013), with mean annual rainfall and temperature of 1370 mm and 24.4 °C. The site was originally *Cerrado stricto sensu* (Brazilian tropical savanna), with pasture introduced in the 1960s and the first cycle of eucalyptus initiated in 2011, which was harvested in 2017.

Initial soil characterization

The soil of this area was classified as a *Neossolo quartzarênico* (Santos et al., 2018), which corresponds to a *Quartzipsamments* (Soil Survey Staff, 2014) or *Arenosol* (IUSS working group WRB, 2015), with loamy sand textural class. The area under study was initially characterized (120 days after harvest of the eucalyptus of the first cycle and before soil preparation for the next eucalyptus cycle) in terms of soil particle size and chemical properties. Disturbed soil samples were collected randomly with three







replications per plot from each HMS to make up a composite sample for the 0.00-0.20, 0.20-0.40, and 0.40-0.60 m soil layers.

This soil has loamy sandy texture, with sand, silt and clay content ranging from 843 to 878 g kg⁻¹, 91 to 54 g kg⁻¹ and 83 to 61 g kg⁻¹, respectively; low fertility (with averages of P=1.86 mg dm⁻³, MO=7.87 g dm⁻³, pH(CaCl₂) = 4.2, K⁺ = 0.17 mmol_c dm⁻³, Ca²⁺ = 2.97 mmol_c dm⁻³, Mg²⁺ = 2.23 mmol_c dm⁻³, H+AI = 14.70 mmol_c dm⁻³, AI³⁺ = 6.22 mmol_c dm⁻³, total organic carbon (TOC) 14 g kg⁻¹, sum of basis 5.37 mmol_c dm⁻³, cation exchange capacity 20.07 mmol_c dm⁻³, base saturation 26.77 % and aluminum saturation 30.97 %.

Particle size was determined by the pipette method (Gee and Bauder, 1986). Total organic carbon (TOC) was determined by the weight loss on ignition method (Ben-Dor and Banin, 1989). Soil chemical analysis was determined at the beginning of the study, as described by Teixeira et al. (2017): exchangeable calcium (Ca), magnesium (Mg), and aluminum (Al) were extracted with ion exchange resin; available phosphorous (P) and potassium (K) determined through an anion exchange resin extractor; organic matter (OM) determined by the digestion method; pH determined in $CaCl_2 0.01 \text{ mol L}^{-1}$ at a soil/solution ratio of 1:2.5; and potential acidity (H+Al) determined indirectly through SMP solution.

Experimental design

Harvest occurred at the beginning of June 2017, after six years of eucalyptus growth, according to the cut to length system. After 120 days (September, 2017) the plots were prepared to simulate the three different harvest management systems (HMS): (i) cut-to-length (CTL) – all residual phytomass was maintained (100 % of leaves, branches, bark, and litter) in the area, simulating the cut-to-length harvest system; (ii) bare (B) – removal of 100 % of the residual phytomass (leaves, branches, bark, and litter through the installation of Sombrite[®] protective netting to impede deposition of plant biomass from the current cycle on the soil surface; and (iii) tree-length (TL) – removal of the bark but maintaining the residual phytomass (leaves, branches, and litter were maintained in the area, but not the bark), simulating the tree-length harvest system.

These three harvest management systems (HMS) were established 120 days after harvest of the eucalyptus due to the need to reduce humidity; then, the logs were kept in place as in the commercial area. Field studies were conducted in a completely randomized experimental design, with four replications, and the soil layers (0.00-0.20, 0.20-0.40, 0.40-0.60, 0.60-0.80, and 0.80-1.00 m) were analyzed in a split-plot experimental arrangement.

The management systems were arranged in rectangular areas of 40.8×27.6 m, containing 12 plant rows at a spacing of 3.4 m between rows and 2.3 m between plants, for an area of 1126 m² (Figure 2). Within this area, the 8 central rows were considered for evaluations, thus constituting an area of 500.8 m², avoiding effects of field access roads and interference from neighboring areas.

Single soil samples were taken from the 0.00-0.05, 0.05-0.10, 0.10-0.20, 0.20-0.40, 0.40-0.60, 0.60-0.80 and 0.80-1.00 m soil layers with a spiral type auger from between the rows, with three replications per soil depths, to determine total organic carbon (TOC) by the weight loss on ignition method (Ben-Dor and Banin, 1989). The organic carbon stock (CS) was calculated from the organic carbon accumulated at each depth using equation 1 (Xie et al., 2007):

$$CS = \frac{(TOC \times BD \times e)}{10}$$

Eq. 1

4





Figure 2. Schematic representation of the work area and respective management systems and replications.

in which CS is the organic carbon stock at a determined depth (Mg ha⁻¹); TOC is the total organic carbon content (g kg⁻¹); BD is the mean soil bulk density of the layer (Mg m⁻³), determined from undisturbed samples; and *e* is the thickness of each layer (0.00-0.10, 0.10-0.20, 0.20-0.40, 0.40-0.60, 0.60-0.80, and 0.80-1.00 m) considered (cm).

Stability of aggregates in water (water-stable aggregates) was determined in undisturbed soil core samples taken in the row from each plot, according to the method of Nimmo and Perkins (2002) and using 2000, 1000, 250, and 53 μ m sieves. These analyses were performed in each one of the four replications of each HMS in the 0.00-0.05, 0.05-0.10, and 0.10-0.20 m soil layers. The aggregates retained in the 2000, 1000, 250, and 53 μ m sieves for the 0.00-0.05 m layers were selected for evaluation of TOC (Ben-Dor and Banin, 1989) because this soil layer directly received the residual phytomass.

Morphology of these aggregates in each HMS for the 0.00-0.05 m layer was analyzed in a scanning electron microscope, EVO-LS15- ZEIS[®] (White, 2008). The aggregates selected had diameters ranging from 4000 to 2000 μ m, 2000 to 1000 μ m, 1000 to 250 μ m, and 250 to 53 μ m. Semi-quantitative microanalyses were carried out using the energy dispersive X-ray analysis (EDX) system, for recognition of the chemical composition of the crystalline materials present in the aggregates smaller than 53 μ m through the electron backscattered diffraction system.

Bulk density (BD), macroporosity (Macro), microporosity (Micro), and total porosity (TP) were evaluated in two undisturbed samples collected in rows from each plot with a Kopecky ring (height: 40 mm; diameter: 55 mm; volume: 95.03 cm³) in the 0.00-0.05, 0.05-0.10, and 0.10-0.20 m soil layers. Bulk density was determined via the volumetric cylinder method, and macroporosity and microposity were determined using the tension table method (Grossman and Reinsch, 2002). Total porosity (TP) was calculated from the sum of macro- and micropores (Flint and Flint, 2002).

Mechanical penetration resistance (PR) was evaluated with an electronic penetrometer (Penetrolog – Falker - cone diameter 12.83 mm). The penetration resistance (PR) recordings were taken in the row, with three replications in each plot, for the 0.00-0.10, 0.10-0.20, 0.20-0.30, and 0.30-0.40 m layers. The data were automatically recorded for each 0.01 m increment of penetration to a maximum of 0.4 m, and the results for each 0.10 m layer is the average of 10 increments of the 0.01 m. The measurements were conducted at a moisture content of around 30 % of field capacity in summer 2017.

Statistical analysis

Analysis of variance was performed on the results obtained for soil properties (TOC, CS, aggregate stability, organic carbon of aggregates, bulk density, macroporosity,

microporosity, total porosity, mechanical penetration resistance) and the F test was applied (p<0.05) when presuppositions (homogeneity of variance and data normality) were met.

When significant difference was found among the sources of variation, the means of the HMS were compared by the Tukey test (p<0.05) and polynomial regression for the soil layers. Since the model was checked based on the p value of the standard deviation of regression, the polynomial regression models were selected from the superior correlation coefficients (r^2) among the regressions significant by the F test, respecting the characteristics of each factor. The SISVAR program (Ferreira, 2019) was used for data analysis.

RESULTS

The TOC and CS content (Table 1) had significant differences for HMS; CTL and B had the highest values in a quadratic response for soil layers ($\hat{y}_{TOC}^{**} = 10.0705 + 0.1486 \text{ x} - 0.0020 \text{ x}^2$, $r^2 = 0.5706$, maximum point = 0.37 m; $\hat{y}_{CS}^{**} = 8.3235 + 1.2688 \text{ x} - 0.0125 \text{ x}^2$, $r^2 = 0.6217$, maximum point = 0.51 m.

The HMS x soil layer interaction for TOC (Table 1) had variations in the 0.20-0.60 m layer. A larger amount of TOC was found in the CTL, which indicates positive effects from preserving the harvest residual phytomass in relation to carbon content, even over a period of only 90 days. Total organic carbon increased significantly in the 0.20-0.60 m layer (Table 1) for the CTL and in the 0.40-0.60 m layer for B, both with quadratic responses ($\hat{y}_{CTL}^{**} = 9.1848 + 0.3811 \text{ x} - 0.0046 \text{ x}^2$, $r^2 = 0.6134$, maximum

| Dremonta | | F values | | CV | | HMS | |
|--|--|---------------------|---------------------|-----------|---------------------|---------------------|-------------|
| Property | HMS | L | HMS × D | % | CTL | В | TL |
| | | m | | | | | |
| TOC (g kg ⁻¹) | 15.993** | 94.310** | 4.302 [*] | 22 | 13.0 A | 10.7 B | 9.2 C |
| CS (Mg ha ⁻¹) | 6.178^{**} | 33.868** | 1.773 ^{ns} | 35 | 31.7 A | 29.6 A | 22.3 B |
| L (m) | 0.00-0.10 | 0.10-0.20 | 0.20-0.40 | 0.40-0.60 | 0.60-0.80 | 0.80-1.00 | |
| TOC (g kg ⁻¹) | 12.2 | 9.7 | 13.5 | 16.0 | 8.7 | 8.5 | |
| | $\hat{y}_{TOC}^{**} = 10.0705 + 0.1486 \text{ x} - 0.0020 \text{ x}^2$, $r^2 = 0.5706$, maximum point = 0.37 m | | | | | | |
| CS (Mg ha ⁻¹) | 18.2 | 14.8 | 40.6 | 47.4 | 25.3 | 24.8 | |
| | \hat{y}_{cs}^{**} = 8.3235 + 1.2688 x - 0.0125 x ² , r ² = 0.6217, maximum point = 0.51 m | | | | | | |
| Dec. | | | TOC (g | y kg⁻¹) | | | |
| HMS | 0.00-0.10 | 0.10-0.20 | 0.20-0.40 | 0.40-0.60 | 0.60-0.80 | 0.80-1.00 | F values |
| CTL | 12.4 a | 10.5 a | 18.1 a | 20.0 a | 9.0 a | 7.8 a | 20.710 ** |
| | $\hat{y}_{\text{TOC(CTL)}}^{**}$ = 9.1848 + 0.3811 x - 0.0046 x ² , r ² = 0.6134, maximum point = 0.41 m | | | | | | |
| В | 12.2 a | 9.5 a | 12.9 b | 15.1 b | 8.9 a | 7.3 a | 19.286 ** |
| | $\hat{y}_{TOC(B)}^{**}$ = 9.8554 + 0.1417 x - 0.0019 x ² , r ² = 0.4863, maximum point = 0.37 m | | | | | | |
| TL | 12.0 a | 9.8 a | 9.5 b | 9.5 c | 7.5 a | 6.8 a | 2.791^{*} |
| $\hat{y}_{TOC(TL)}^{*}$ = 11.4488 - 0.0528 x, r ² = 0.8200, range of variation = -0.528 g C kg ⁻¹ 0.10 m ⁻¹ | | | | | | | |
| F values | 0.033 ^{ns} | 0.299 ^{ns} | 16.238 ** | 20.170 ** | 0.570 ^{ns} | 0.192 ^{ns} | |

Table 1. F values, coefficients of variation (CV) and determination (r^2), regression equations, decomposition (Dec.) of interaction, harvest management systems (HMS) × soil layer (L), for total organic carbon (TOC; g kg⁻¹) and carbon stock (CS; Mg ha⁻¹), after 90 days of Harvest Management System implantation

Mean values followed by the same uppercase letters in the row (HMS) and lowercase letters in the column (TOC) do not differ statistically from each other by the Tukey test for p<0.05.^{ns}: not significant; ^{**} and ^{*} significant at 1 and 5 %, respectively. ⁽¹⁾ CTL: cut-to-length (Maintaining 100 % of the residual phytomass in the area); B: bare (Removal of 100 % of the vegetation residual phytomass and setting up a net to collect all leaves, branches, etc); TL: tree-length (Removal of the bark and maintaining the other residual phytomass in the area).



point = 0.41 m; \hat{y}_{B}^{**} = 9.8554 + 0.1417 x - 0.0019 x², r² = 0.4863, maximum point = 0.37 m; \hat{y}_{TL}^{*} = 11.4488 - 0.0528 x, r² = 0.8200, range of variation = -0.528 g C kg⁻¹ 0.10 m⁻¹).

In the CTL, where the residual phytomass from the harvest was maintained, there are increases in TOC (Table 1) from this phytomass. However, in the B and TL, where the residual phytomass was removed (totally and partially), the TOC content already declines after a period of only 90 days in the HMS, in which TOC ranged from 12 to 16 g kg⁻¹. Macroaggregates (>250 μ m) predominate in the distribution of the soil aggregates by

| Table 2. F values and coefficient of variation | (CV) for distribution of the aggregate | s (in %) in relation to the harvest mar | nagement |
|--|--|---|----------|
| systems (HMS) and the soil layer (L) | | | - |

| Variable - | Diameter | | | | | | |
|--------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|--|
| variable | >2000 μm | 2000-1000 μm | 1000-500 μm | 500-250 μm | 250-102 μm | 102-53 μm | |
| | F values | | | | | | |
| HMS | 1.945 ^{ns} | 0.051 ^{ns} | 0.882 ^{ns} | 1.694 ^{ns} | 2.807 ^{ns} | 0.640 ^{ns} | |
| L (m) | 45.350** | 25.770** | 29.399** | 30.030** | 34.707** | 1.580 ^{ns} | |
| HMS × D | 1.425 ^{ns} | 0.514 ^{ns} | 0.812 ^{ns} | 0.641 ^{ns} | 2.810*# | 0.494 ^{ns} | |
| CV (%) | 10 | 75 | 71 | 64 | 53 | 101 | |
| HMS ⁽¹⁾ | Mean (%) | | | | | | |
| CTL | 85.39a | 1.89a | 2.17a | 3.71a | 2.77a | 0.45a | |
| В | 80.03a | 2.00a | 2.70a | 5.71a | 4.11a | 0.57a | |
| TL | 79.96a | 2.09a | 3.21a | 5.75a | 4.69a | 0.39a | |
| L (m) | | | | | | | |
| 0.00-0.05 | 89.79a | 0.80b | 1.30b | 2.50b | 2.07b | 0.41a | |
| 0.05-0.10 | 87.00a | 1.24b | 1.41b | 3.38b | 2.69b | 0.44a | |
| 0.10-0.20 | 67.47b | 3.95a | 5.38a | 10.07a | 7.11a | 0.68a | |

Mean values followed by the same lowercase letters in the column do not differ statistically from each other by the Tukey test for p<0.05. ^{ns}: not significant; ^{**} and ^{*} significant at 1 and 5 %, respectively. [#]: decomposition not presented. ⁽¹⁾ HMS: harvest management systems. CTL: cut-to-length (Maintaining 100 % of the residual phytomass in the area); B: bare (Removal of 100 % of the residual vegetation phytomass and setting up a net to collect all leaves, branches, etc.); TL: tree-length (Removal of the bark and maintaining the other residual phytomass in the area).

Table 3. F values, coefficient of variation (CV), and decomposition (Dec.) of interaction, harvest management systems (HMS) \times aggregates diameter (μ m), for the carbon of the aggregates (g kg⁻¹) in the 0.00-0.05 soil layer (L), after 90 days of Harvest Management System implantation

| Variable | F values | | | CV | | HMS | |
|-----------------------------------|--------------------|---------------------|---------------------|---------------------|---------------------|------|------|
| variable | HMS | Diameter | HMS × D | % | CTL ⁽¹⁾ | В | TL |
| | | μm | | | | | |
| C aggregates (g kg ¹) | 0.08 ^{ns} | 16.686** | 2.796^{*} | 26 | 15.0 | 14.5 | 14.9 |
| | | Diamete | r (µm) | | | | |
| | >2000 | 2000-1000 | 1000-250 | 250-53 | | | |
| C aggregates (g kg ¹) | 19.7A | 17.2A | 12.4B | 9.9B | | | |
| Dec. | | Diamete | r (µm) | | | | |
| | >2000 | 2000-1000 | 1000-250 | 250-53 | F values | | |
| HMS | | C aggregate | es (g kg¹) | | | | |
| CTL | 18.5abA | 18.1aA | 14.0aAB | 9.4aB | 5.004** | | |
| В | 15.8bA | 17.6aA | 12.2aA | 12.3aA | 1.966 ^{ns} | | |
| TL | 24.8aA | 15.9aB | 11.1aBC | 7.8aC | 15.307** | | |
| F values | 6.020** | 0.381 ^{ns} | 0.586 ^{ns} | 1.482 ^{ns} | | | |

Mean values followed by the same uppercase letters in the row and lowercase letters in the column do not differ statistically from each other by the Tukey test for p<0.05.^{ns}: not significant; ^{**} and ^{*}: significant at 1 and 5 %, respectively.⁽¹⁾ CTL: Cut-to-Length (Maintaining 100 % of the residual phytomass in the area); B: Bare (Removal of 100 % of the vegetation residual phytomass and setting up a net to collect all leaves, branches, etc.); TL: Tree-Length (Removal of the bark and maintaining the other residual phytomass in the area).



diameter (Table 2) (Tisdall and Oades, 1982), especially the >2000 μ m class, which is prevalent in all the HMS, constituting at least 80 % of the aggregates.

Aggregates >2000 μ m are important in the 0.00-0.10 m soil layer, coinciding with the TOC content higher in this layer than in the 0.10-0.20 m layer due to its nearness to the residual phytomass left on the surface. Of the aggregates analyzed for TOC in the 0.00-0.05 m soil layer (Table 3), those of greater diameter (>2000 μ m and >1000 μ m) lead to the greatest retention of TOC in the soil, without differences among the HMS at 90 days. Nevertheless, it is important to note that microaggregates (<250 μ m) also retain TOC within them, though in smaller amounts than in the larger aggregates.

The soil under the management systems studied has isolated particles of microaggregate dimension (Figure 3d). The silicon content identified in the isolated particles, combined with oxygen, indicates the presence of quartz, which is the predominant mineral in the isolated particles (Figures 4, 5, and 6).



Figure 3. Images obtained in a scanning electron microscope (SEM; Magnitude = 250 X; EHT = 20.00 kV) of the aggregates with diameter from 4000 to 2000 μ m (a), 2000 to 1000 μ m (b), 1000 to 250 μ m (c), and less than 250 μ m (d), in the 0.00-0.05 m soil layer and different harvest management systems (HMS). ⁽¹⁾ CTL: cut-to-length (maintaining 100 % of the residues in the area); B: bare (Removal of 100 % of the residues and setting up a net to collect all leaves, branches, etc.); TL: tree-length (Removal of the bark from the area and maintaining the other residues in the area).







Figure 4. Images obtained in a scanning electron microscope (SEM) and energy dispersive X-ray (EDX) of the aggregates with diameter from 0 to 53 μ m (a) in the 0.00-0.05 m soil layer in the cut to length harvest management systems.



Figure 5. Images obtained in a scanning electron microscope (SEM) and energy dispersive X-ray (EDX) of the aggregates with diameter from 0 to 53 μ m (a) in the 0.00-0.05 m soil layer in the bare harvest management systems.







Figure 6. Images obtained in a scanning electron microscope (SEM) and energy dispersive X-ray (EDX) of the aggregates with diameter from 0 to 53 μ m (a) in the 0.00-0.05 m soil layer in the tree length harvest management systems.

In the HMS × soil layer decomposition (Table 5), the Macro values remain near 3 %, now for the three HMS in the 0.10-0.20 m layer, and show that the CTL, which maintains all residual phytomass at the surface, has a higher mean value for Macro in the 0.00-0.05 m layer (11.84 %). The decomposition of the HMS × carbon of the aggregates interaction (Table 3) confirms the maintenance of TOC in the larger aggregates (Figure 3), particularly in the CTL and TL, both with maintenance (total and partial, respectively) of the residual phytomass from eucalyptus harvest. However, in the B, with the removal of 100 % of the residual phytomass, differences are not observed for TOC among aggregates of different diameters (Table 3).

The response observed in the B for carbon of the aggregates (Table 3) highlights the importance of the management system adopted, in which removal of the residual phytomass reduced the TOC content and, consequently, the stability of the aggregates of greater diameter (Table 2), without difference in carbon retention among the different diameters of aggregates that compose the soil in this HMS.

Although soil BD and Macro properties did not vary among the HMS, they varied among the layers evaluated. Macroporosity approaches 3 % in the 0.10-0.20 m layer, suggesting severe limitation of water infiltration in the soil (Table 4). The HMS did not significantly affect total porosity, exhibiting mean values for CTL, B and TL of 38.74, 37.11 and 38.14 %, respectively (Table 4). Mechanical penetration resistance Weight (PR) (Table 4 and Figure 7) had significant variations for the HMS and soil layers evaluated. The CTL led to the lowest PR, differing from the others, which are similar to each other.

In this study, the CTL has the lowest PR, lower at the surface, a direct and indirect effect of TOC, whereas under the surface, the benefits are not evident. As the depth of the profile increased, the PR values increased, with a quadratic response $(\hat{y}_{PR\#}^{**} = 0.668 + 2.008 \text{ x} - 0.3020 \text{ x}^2; \text{ r}^2 = 0.9983)$, and PR reached a maximum level in the 0.33 m depth.

Table 4. Mean values for macroporosity (Macro) and microporosity (Micro), total porosity (TP), soil bulk density (BD), mechanical penetration resistance (PR), F values, and coefficient of variation (CV) by harvest management system (HMS) and soil layer (L), after 90 days of Harvest Management System implantation

| Source of variation | BD | Macro | Micro | ТР | PR [#] | |
|---------------------|---------------------|---------------------|---------------------|---------------------|-----------------|------------------|
| | Mg m⁻³ | | % | | MPa | |
| F values | | | | | | |
| HMS | 1.639 ^{ns} | 1.800 ^{ns} | 0.460 ^{ns} | 2.359 ^{ns} | 5.42 | 7** |
| L (m) | 17.330** | 22.807** | 6.856** | 6.628** | 41.06 | 0** |
| $HMS \times L$ | 1.682 ^{ns} | 1.957 ^{ns} | 0.318 ^{ns} | 1.309 ^{ns} | 0.12 | 7 ^{ns} |
| CV (%) | 3 | 45 | 8 | 7 | 25 | |
| HMS ¹ | | | | | | |
| CTL | 1.49a | 7.22a | 31.52a | 38.74a | 1.72 | а |
| В | 1.52a | 5.54a | 31.57a | 37.11a | 2.16 | b |
| TL | 1.49a | 5.81a | 32.33a | 38.14a | 2.24 | b |
| Layer (m) | | | | | Layer (r | n) ^{##} |
| 0.00-0.05 | 1.46a | 8.38a | 30.52a | 38.91a | 0.00-0.10 | 1.03 |
| 0.05-0.10 | 1.51b | 6.59a | 31.28a | 37.87ab | 0.10-0.20 | 2.18 |
| 0.10-0.20 | 1.54b | 2.81b | 33.34b | 36.16b | 0.20-0.30 | 2.61 |
| | | | | | 0.30-0.40 | 2.55 |

Mean values followed by the same lowercase letters in the column do not differ statistically from each other by the Tukey test for p<0.05. ^{ns}: not significant; ^{**} and ^{*}: significant at 1 and 5 %, respectively. ⁽¹⁾ CTL: Cut-to-Length (Maintaining 100 % of the residual phytomass in the area); B: Bare (Removal of 100 % of the vegetation residual phytomass and setting up a net to collect all leaves, branches, etc.); TL: Tree-Length (Removal of the bark and maintaining the other residual phytomass in the area); [#]: soil humidity in field: 8-10 % (volume); water holding capacity: 27 % (volume) according to USDA (2014). ^{##}: regression equation for PR in-layer $\hat{y}_{PR##}^{**}$: 0.668 + 2.008 x - 0.3020 x², r² = 0.9983.

Table 5. Decomposition of the interaction between harvest management systems (HMS) \times soil layer (L) for macroporosity

| Call Joyan | | E velves | | |
|------------|----------|----------------|---------|---------------------|
| Soli layer | CTL | В | TL | - r values |
| m | | Macropores (%) | | |
| 0.00-0.05 | 11.84aA | 7.02aB | 9.00aAB | 5.046** |
| 0.05-0.10 | 7.32aA | 6.76aA | 5.34abA | 0.617 ^{ns} |
| 0.10-0.20 | 2.49bA | 2.84bA | 3.08bA | 0.051 ^{ns} |
| F values | 12.359** | 9.322** | 5.040** | |

Mean values followed by the same uppercase letters in the row and lowercase letters in the column do not differ statistically from each other by the Tukey test for p<0.05.^{ns}: not significant; ^{**} and ^{*} significant at 1 and 5 %, respectively.⁽¹⁾ CTL: Cut-to-Length (Maintaining 100 % of the residual phytomass in the area); B: Bare (Removal of 100 % of the vegetation residual phytomass and setting up a net to collect all leaves, branches, etc.); TL: Tree-Length (Removal of the bark and maintaining the other residual phytomass in the area).

DISCUSSION

The TOC and CS content found in the uppermost soil layer (Table 1) is related to the input of residual phytomass from eucalyptus harvest, a recurrent response frequently reported in the literature (Costa Jr et al., 2012; Guan et al., 2015; Jesus et al., 2015; Marques et al., 2016; Sena et al., 2017). Below the surface, the increases observed in TOC and CS in deeper layers (0.20-0.60 m – Table 1) for the CTL and B coincide with the layer of greatest abundance of the root system in eucalyptus. This increase may also be the result of the downward movement of TOC in sandy soil, as also reported by Marques et al. (2016) evaluating primary forest in the Amazon region under *Latossolos*



Soil Mechanical Resistance to Penetration (MPa)



Figure 7. Soil mechanical resistance to penetration in the harvest management systems (HMS).

(plateau), *Argissolo* (slope), and *Espodossolo* (lowland). It may reflect effects from previous crops, though this is a less likely hypothesis due to climate conditions (hot and humid climate) and microbial activity (Costa Jr et al., 2012; Cunha et al., 2012), it is a question under analysis.

Commercial eucalyptus plantations can incorporate carbon in the soil due to the biomass deposited annually in the form of an organic covering and of dead roots (Silva et al., 2012), a fact observed in the CTL in which the residual harvest phytomass was maintained. However, in the B and TL, where residual phytomass was removed (totally and partially) over a period of 90 days, the TOC and CS content (Table 1) already manifests a decline at the layer of 0.20-0.40 m, indicating that removal of the litter that would accumulate over a period of 6 years could rapidly reduce the TOC in cultivated areas on sandy soil (Rocha et al., 2018). These results suggest that the residual harvest phytomass should be maintained in the area for more effective conservationist management.

Total organic carbon does not differ statistically between CTL and B, which have a higher content of TOC than in TL. However, the initial values for TOC in the 0.00-0.20 m soil layer indicate that this decreased in CTL from 14 to 12.4 g kg⁻¹, in B from 16 to 11 g kg⁻¹, and in TL from 16 to 11 g kg⁻¹, with smaller differences in the CTL (100 % residual phytomass at the surface) than in the TL or B HMS, where the residual phytomass was partially or totally removed.

In a study on sandy soil, Soares et al. (2017) found changes of greater magnitude in carbon stocks in the organic matter fractions in the 0.00-0.10 m soil layer due to the large contribution of the eucalyptus residual phytomass. The organic fraction of the soils do not have the same stability as the mineral fraction under agricultural use; thus, intensive use of the soil can contribute to its degradation (Six and Paustian, 2014; Sena et al., 2017), compromising the stability of its aggregates. Consequently, losses of TOC occur, which increases soil bulk density and reduces total porosity.

Macroaggregates (>2000 μ m) are predominant in the 0.00-0.10 m soil layer and all HMS, as shown in (Table 2). This predominance occurs in areas with low soil movement, as these areas normally exhibit greater stability of aggregates and greater organic carbon content (Six and Paustian, 2014; Sena et al., 2017). This explains the occurrence of 80 % of aggregates >2000 μ m in the sandy soil in the three HMS examined, indicating that the effects of total or partial (B or TL) litter removal over 90 days before sampling



did not reduce aggregate stability. Greater quantity of macroaggregates stable in water (>2000 μ m) in *Latossolos* under Cerrado was reported by Salton et al. (2008) in a study on different soil management systems with low soil movement, which was observed in other studies (An et al., 2010) as well, including a study on sandy soil (Sena et al., 2017).

The higher TOC content observed in the soil layer 0.00-0.10 m (Table 1) is associated with the higher stability of the aggregates >2000 μ m (Table 2), shows the importance of this carbon on maintenance of the stability of the aggregates and, therefore, on the structural stability of the soil (Salton et al., 2008; Guan et al., 2015). The larger aggregates (>2000 μ m) and those of lower stability tend to subdivide into smaller aggregates (Six et al., 2000), a process directly related to the reduction in TOC content (Nichols and Toro, 2011; Bast et al., 2014).

The loamy sandy soil (843 to 878 g kg⁻¹ of sand in the soil) in the area evaluated may have produced macroaggregates of lower stability and stable microaggregates. An explanation for greater stability of the microaggregates is the presence of single quartz particles of sand size (Figures 3, 4, 5 and 6), which grants greater resistance and absence of response to the HMS. The quartz particles show indications of coating (Figure 3 – diameter <250 μ m), suggesting that they were part of larger aggregates, and their greater stability also involves the amount of sand (quartz) they contain, which is greater than 80 % (Figures 4, 5 and 6).

Soil BD is not restrictive (1.46 to 1.54 Mg m⁻³) to water infiltration in the soil or to root development of eucalyptus, and neither are Micro and TP (Table 4). Reports of BD in loamy sandy soil ranging from 1.40 to 1.80 Mg m⁻³ were considered restrictive when higher than 1.55 Mg m⁻³ (Reichert et al., 2003); 1.65 Mg m⁻³ for soils with less than 200 g kg⁻¹ of clay (Reinert et al., 2001), and 1.60 Mg m⁻³ with macroporosity smaller than 10 % (Michelon, 2005).

Nevertheless, macroporosity has restrictive values (Table 5), near 3 % for the three HMS in the 0.10-0.20 m soil layer, indicating that the CTL, which maintains all the residual phytomass at the surface, has a higher mean value for Macro in the 0.00-0.05 m layer (11.84 %). However, the occurrence of Macro = 3 % in the layer immediately below can limit liquid and gas exchanges, as cited by Carter (2002), who considers macroporosity of around 10 % as adequate.

The presence of macroaggregates is positively associated with soil organic matter content (Costa Jr et al., 2012; Six and Paustian, 2014), as observed at the surface (Tables 1 and 3). The BD increased with the depth of the profile, coinciding with the reduction in TOC content, which is related to the aggregation of the soil (Salton et al., 2008; Sena et al., 2017). The Macro observed (Table 4) suggests that the reduced Macro is restricted to a narrow, compacted depth, which ended up not being represented in the BD samples; variations from 1.46 to 1.54 Mg m⁻³ in these samples are not compatible with a Macro of 3 %. Micro and TP did not vary for HMS and depths (Table 4). In an area growing eucalyptus (2 years) in sandy soil, a similar response was observed (Sena et al., 2017), low macroporosity (5 %) under the surface layer (0.10-0.30 m), with density lower than the limit established by Reichert et al. (2003), emphasizing the existence of a narrow, compacted depth. Jesus et al. (2015) evaluated physical properties after harvest activities and concluded that the aforementioned 0.10-0.20 m layer is most affected, which is shown in the BD, TP, and PR.

Total porosity (Table 4) was not significantly affected by the HMS and soil layers, with mean values near 40 %. The ideal TP of soil is 50 %, or 0.5 $m^3 m^{-3}$ of its total volume (Kiehl, 1979); under these parameters, the HMS evaluated have limiting conditions for plant development.

Mechanical penetration resistance (PR) (Table 4), closely linked to properties like bulk density and total porosity, showed significant variations for the HMS and soil layers evaluated. Although BD, macro, microporosity and TP do not vary among HMS (Table 4),

it is possible to relate the behavior of PR to macroporosity (Table 5), which in CTL is 40 % higher than B and 32 % higher than TL, the greater the PR and the lower is the macroporosity (Sena et al., 2017), as can be observed in the B and TL.

Lower PR (1.72 MPa) for CTL, compared to B and TL, which are higher and similar to each other (2.16 and 2.24 MPa, respectively), could be related to the maintenance of residual crop phytomass on the soil surface. However, this residual phytomass was removed only 90 days before, and that is not a period of sufficient length to affect soil PR in deeper layers, suggesting that these effects are still related to the process that was previously established in setting up the crop (6 years before) and mechanized activities at harvest (7 months before). These effects are also recorded in the CTL; however, they are diluted due to greater organic input at the surface. Jesus et al. (2015) observed that harvest under the CTL allows the effects of this activity to be mitigated, with lower PR. Maintaining bark in the area is one of the reasons for this.Studies on soils, including soils in the eucalyptus crop, adopt PR values ranging from 2.0 to 3.0 MPa as limiting plant development (Betioli Júnior et al., 2012; Jesus et al., 2015; Reichert et al., 2018). Thus, values considered limiting to the root development of eucalyptus were not observed.

As observed by Sena et al. (2017) in a study on sandy soil growing eucalyptus in the east of Mato Grosso do Sul, a soil can be considered degraded when it combines the following properties: low total porosity, low macroporosity, high density, low TOC, low CS and high PR, just as observed in the HMS in which the cover material coming from the harvest was removed. In light of the above, sandy soils and their physical properties manifest the need to adopt systems that favor soil structuring, such as those that increase organic matter content (Jesus et al., 2015; Marques et al., 2016; Sena et al., 2017). Nevertheless, mid-tolong-term evaluations are necessary, given the dynamic changes in soil physical quality imposed by the harvest management systems, edaphic and climatic conditions, and management operations in the next crop (such as nitrogen fertilization).

CONCLUSION

Keeping the residual phytomass on the soil surface can positively impact total organic carbon, with lower reduction in total organic carbon under the cut-to-length harvest management system. However, carbon stock is greater at the layer of 0.20-0.60; as the soil has a sandy texture, carbon moves through the soil profile, which has lower soil mechanical penetration resistance at the surface (0.00-0.10 m), once more under the cut-to-length system.

In the cut-to-length harvest management system, maintaining residual crop phytomass on the soil surface provides better soil physical conditions, with greater macroporosity (0.00-0.05 m), aggregates with more carbon, and lower soil mechanical penetration resistance compared to systems that maintain only part of the harvest residual phytomass or no harvest residual phytomass on the surface.

ACKNOWLEDGEMENTS

We thank CAPES and FAPESP for providing the Doctorate scholarship and the financial support, and Eldorado Brasil, a pulp company, for allowing study in their experimental areas and logistical support. São Paulo Research Foundation (FAPESP, grant number 2017/14049-2).

SUPPLEMENTARY MATERIAL

Supplementary data to this article can be found online at https://www.rbcsjournal.org/ wp-content/uploads/articles_xml/1806-9657-rbcs-45-e0200190/1806-9657-rbcs-45e0200190-suppl01.pdf



AUTHOR CONTRIBUTIONS

Conceptualization: 向 Kátia Luciene Maltoni (equal).

Data curation: (D) Karla Nascimento Sena (equal).

Formal analysis: (b) Glaucia Amorim Faria (equal), (b) Karla Nascimento Sena (equal), (b) Kátia Luciene Maltoni (equal) and (b) Maria Júlia Betiolo Troleis (equal).

Funding acquisition: (b) Karla Nascimento Sena (equal) and (b) Kátia Luciene Maltoni (equal).

Methodology: (D) Glaucia Amorim Faria (equal), (D) Karla Nascimento Sena (equal) and (D) Kátia Luciene Maltoni (equal).

Project administration: (D) Karla Nascimento Sena (equal), (D) Kátia Luciene Maltoni (equal) and (D) Maria Júlia Betiolo Troleis (supporting).

Resources: (b) Karla Nascimento Sena (equal) and (b) Kátia Luciene Maltoni (equal).

Software: 💿 Glaucia Amorim Faria (equal) and 💿 Karla Nascimento Sena (equal).

Supervision: 💿 Kátia Luciene Maltoni (lead).

Validation: (D) Glaucia Amorim Faria (equal), (D) Karla Nascimento Sena (equal) and (D) Kátia Luciene Maltoni (equal).

Visualization: (D) Karla Nascimento Sena (equal).

Writing - original draft: (D) Karla Nascimento Sena (equal).

Writing - review & editing: 💿 Karla Nascimento Sena (equal).

REFERENCES

Alvares CA, Stape JL, Sentelhas PC, Gonçalves JLM, Sparovek G. Köppen's climate classification map for Brazil. Meteorol Z. 2013;22:711-28. https://doi.org/10.1127/0941-2948/2013/0507

An S, Mentler A, Mayer H, Blum WEH. Soil aggregation, aggregate stability, organic carbon and nitrogen in different soil aggregate fractions under forest and shrub vegetation on the Loess Plateau, China. Catena. 2010;81:226-33. https://doi.org/10.1016/j.catena.2010.04.002

Bast A, Wilcke W, Graf F, Lüscher P, Gärtner H. The use of mycorrhiza for eco-engineering measures in steep alpine environments: effects on soil aggregate formation and fine-root development. Catena. 2014;39:1753-63. https://doi.org/10.1002/esp.3557

Ben-Dor E, Banin A. Determination of organic matter content in arid-zone soils using a simple "Loss-on-Ignition" method. Comm. Soil Sci Plant Anal. 1989;20:1675-95. https://doi.org/10.1080/00103628909368175

Betioli Júnior E, Moreira WH, Tormena CA, Ferreira CJB, Silva AP, Giarola NFB. Intervalo hídrico ótimo e grau de compactação de um Latossolo Vermelho após 30 anos sob plantio direto. Rev Bras Cienc Solo. 2012;36:971-82. https://doi.org/10.1590/S0100-06832012000300027

Carter MR. Quality, critical limits and standardization. In: Lal R, editor. Encyclopedia of soil science. New York, Marcel Dekker; 2002. p. 1062-5.

Costa Jr C, Piccolo MC, Siqueira Neto M, Camargo PB, Cerri CC, Bernoux M. Carbono em agregados do solo sob vegetação nativa, pastagem e sistemas agrícolas no bioma Cerrado. Rev Bras Cienc Solo. 2012;36:1311-21. https://doi.org/10.1590/S0100-06832012000400025

Cunha EQ, Stone LF, Ferreira EPB, Didonet AD, Moreira JAA. Atributos físicos, químicos e biológicos de solo sob produção orgânica impactados por sistemas de cultivo. Rev Bras Eng Agric Ambinetal. 2012;16:56-63. https://doi.org/10.1590/S1415-43662012000100008



Du H, Zeng F, Peng W, Wang K, Zhang H, Liu L, Song T. Carbon storage in a *Eucalyptus* plantation chronosequence in southern China. Forests. 2015;6:1763-78. https://doi.org/10.3390/f6061763

Ferreira DF. SISVAR: A computer analysis system to fixed effects split plot type designs. Rev Bras Biometria. 2019;37:529-35. https://doi.org/10.28951/rbb.v37i4.450

Flint LE, Flint AL. Porosity. In: Dane JH, Topp GC, editors. Methods of soil analysis: Part 4 - Physical methods. Madison: Soil Science Society of America; 2002. p. 241-54.

Gee GW, Bauder JW. Particle-size analysis. In: Kluter A, editor. Methods of soil analysis: Part 1 - Physical and mineralogical methods. 2nd ed. Madison: Soil Science Society of America; 1986. p. 383-411.

Grossman RB, Reinsch TG. Bulk density and linear extensibility. In: Dane JH, Topp GC, editors. Methods of soil analysis: Part 4 - Physical methods. Madison: Soil Science Society of America; 2002. p. 201-28.

Guan F, Tang X, Fan S, Zhao J, Peng C. Changes in soil carbon and nitrogen stocks followed the conversion from secondary forest to Chinese fir and Moso bamboo plantations. Catena. 2015;133:455-60. https://doi.org/10.1016/j.catena.2015.03.002

Huang J, Hartemink AE. Soil and environmental issues in sandy soils. Earth-Sci Rev. 2020;208:103295. https://doi.org/10.1016/j.earscirev.2020.103295

Indústria Brasileira de Árvores - IBÁ. Relatório anual 2020. Available from: https://iba.org/datafiles/publicacoes/relatorios/relatorio-iba-2020.pdf.

IUSS Working Group WRB. World reference base for soil resources 2014, update 2015: International soil classification system for naming soils and creating legends for soil maps. Rome: Food and Agriculture Organization of the United Nations; 2015. (World Soil Resources Reports, 106).

Jesus GL, Silva IR, Almeida LFJ, Santos MA, Leite FP, Neves JCL. Eucalyptus productivity, soil physical properties and organic matter fractions influenced by traffic intensity and harvest residues. Rev Bras Cienc Solo. 2015;39:1190-203. https://doi.org/10.1590/01000683rbcs20140494

Kiehl EJ. Manual de edafologia: relação solo-planta. São Paulo: Agronômica Ceres; 1979.

Krishna MP, Mohan M. Litter decomposition in forest ecosystems: a review. Energ Ecol Environ. 2017;2:236-49. https://doi.org/10.1007/s40974-017-0064-9

Laclau J, Levillain J, Deleporte P, Dieu Nzila J, Bouillet J, Saint André L, Versini A, Mareschal L, Nouvellon Y, Thongo M'Bou A, Ranger J. Organic residue mass at planting is an excellent predictor of tree growth in Eucalyptus plantations established on a sandy tropical soil. For Ecol Manage. 2010;260:2148-59. https://doi.org/10.1016/j.foreco.2010.09.007

Machado CC, Silva EM, Pereira SR. O setor florestal brasileiro e a colheita florestal. In: Machado CC, editor. Colheita florestal. 2. ed. Viçosa, MG: Universidade Federal de Viçosa; 2008. p. 15-42.

Marques JDD, Luizão FJ, Teixeira WG, Vitel CM, Marques EMD. Soil organic carbon, carbon stock and their relationships to physical attributes under forest soils in central Amazonia. Rev Arvore. 2016;40:197-208. https://doi.org/10.1590/0100-67622016000200002

Michelon CJ. Qualidade física dos solos irrigados do Rio Grande do Sul e do Brasil Central [dissertação]. Santa Maria: Universidade Federal de Santa Maria; 2005.

Nichols KA, Toro M. A whole soil stability index (WSSI) for evaluating soil aggregation. Soil Till Res. 2011;111:99-104. https://doi.org/10.1016/j.still.2010.08.014

Nimmo JR, Perkins KS. Aggregate stability and size distribution. In: Dane JH, Topp GC, editors. Methods of soil analysis: Part 4 - Physical methods. Madison: Soil Science Society of America; 2002. p. 317-28.

Reichert JM, Cechin NF, Reinert DJ, Rodrigues MF, Suzuki LEAS. Ground-based harvesting operations of Pinus taeda affects structure and pore functioning of clay and sandy clay soils. Geoderma. 2018;331:38-49. https://doi.org/10.1016/j.geoderma.2018.06.012



Reichert JM, Reinert DJ, Braida JA. Qualidade dos solos e sustentabilidade de sistemas agrícolas. Cienc Amb. 2003;27:29-48.

Reinert DJ, Reichert JM, Silva VR. Propriedades físicas de solos em sistema de plantio direto irrigado. In: Carlesso R, Petry MT, Rosa GM, Ceretta CA, editores. Irrigação por aspersão no Rio Grande do Sul. Santa Maria: UFSM; 2001. p. 114-33.

Rocha JHT, Gonçalves JLM, Brandani CB, Ferraz AV, Franci AF, Marques ERG, Arthur Junior JC, Hubner A. Forest residue removal decreases soil quality and affects wood productivity even with high rates of fertilizer application. For Ecol Manage. 2018;430:188-95. https://doi.org/10.1016/j.foreco.2018.08.010

Salton JC, Mielniczuk J, Bayer C, Boeni M, Conceição PC, Fabrício AC, Macedo MCM, Broch DL. Agregação e estabilidade de agregados do solo em sistemas agropecuários em Mato Grosso do Sul. Rev Bras Cienc Solo. 2008;32:11-21. https://doi.org/10.1590/S0100-06832008000100002

Santos HG, Jacomine PKT, Anjos LHC, Oliveira VA, Lumbreras JF, Coelho MR, Almeida JA, Araújo Filho JC, Oliveira JB, Cunha TJF. Sistema brasileiro de classificação de solos. 5. ed. rev. ampl. Brasília, DF: Embrapa; 2018.

Schmidt MWI, Torn MS, Abiven S, Dittman T, Guggenberger G, Janssen IA, Kleber M, Kogel-Knabner I, Lehman J, Manning DAE, Nannipieri P, Rasse DP, Weiner S, Trumbore SE. Persistence of soil organic matter as an ecosystem property. Nature. 2011;478:49-56. https://doi.org/10.1038/nature10386

Sena KN, Maltoni KL, Faria GA, Cassiolato AMR. Organic carbon and physical properties in sandy soil after conversion from degraded pasture to *Eucalyptus* in the Brazilian *Cerrado*. Rev Bras Cienc Solo. 2017;41:e0150505. https://doi.org/10.1590/18069657rbcs20150505

Silva CF, Pereira MG, Miguel DL, Feitora JCF, Loss A, Menezes CEG, Silva EMR. Carbono orgânico total, biomassa microbiana e atividades enzimáticas do solo de áreas agrícolas, florestais e pastagem - Processos e propriedades do solo. Rev Bras Cienc Solo. 2012;36:1680-9. https://doi.org/10.1590/S0100-06832012000600002

Six J, Paustian K. Aggregate-associated soil organic matter as an ecosystem property and a measurement tool. Soil Biol Biochem. 2014;68:4-9. https://doi.org/10.1016/j.soilbio.2013.06.014

Six J, Paustrian K, Elliott ET, Combrink C. Soil structure and organic matter: distribution of aggregate-size classes and aggregate-associated carbon. Soil Sci Soc Am J. 2000;64:681-9. https://doi.org/10.2136/sssaj2000.642681x

Soares BEM, Silva IR, Barros NF, Teixeira RS, Fonseca S, Vasconcelos AA, Souza RN. Soil organic matter fractions under second-rotation eucalyptus plantations in eastern Rio Grande do Sul. Rev Arvore. 2017;41:e410107. https://doi.org/10.1590/1806-90882017000100007

Soil Survey Staff. Keys to soil taxonomy. 12th ed. Washington, DC: United States Department of Agriculture, Natural Resources Conservation Service; 2014.

Teixeira PC, Donagemma GK, Fontana A, Teixeira WG. Manual de métodos de análise de solo. 3. ed. rev e ampl. Brasília, DF: Embrapa; 2017.

Tisdall JM, Oades LM. Organic matter and water stable aggregates in soil. J Soil Sci. 1982;33:141-63. https://doi.org/10.1111/j.1365-2389.1982.tb01755.x

Vicente LC, Gama-Rodrigues EF, Gama-Rodrigues AC, Marciano CR. Organic carbon within soil aggregates under forestry systems and pasture in a southeast region of Brazil. Catena. 2019;182:104139. https://doi.org/10.1016/j.catena.2019.104139

White GN. Scanning electron microscopy. In: Ulery AI, Dress LR, editors. Methods of soil analysis: Part 5 - Mineralogical methods. Madison: Soil Science Society of America; 2008. p. 269-97.

Xie ZB, Zhu JG, Liu G, Cadisch G, Hasegawa T, Chen CM, Sun HF, Tang HY, Zeng Q. Soil organic carbon stocks in China and changes from 1980s to 2000s. Global Change Biol. 2007;13:1989-2007. https://doi.org/10.1111/j.1365-2486.2007.01409.x

17