Effects of water-soluble fraction of petroleum on growth and prey consumption of juvenile *Hoplias aff. malabaricus* (Osteichthyes: Erythrinidae)


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(With 2 figures)

**Abstract**

The influence of the water-soluble fraction of petroleum (WSF) on prey consumption and growth of juvenile trahira *Hoplias aff. malabaricus* was investigated. Juveniles were submitted to either WSF or Control treatment over 28 days, and jewel tetra *Hyphessobrycon eques* adults were offered daily as prey for each predator. Total prey consumption ranged from 16 to 86 individuals. Despite the initially lower prey consumption under WSF exposure, there were no significant differences in overall feeding rates between the two treatments. Water-soluble fraction of petroleum had a negative effect on the growth in length of *H. aff. malabaricus* juveniles. Although unaffected, prey consumption suggested a relative resistance in *H. aff. malabaricus* to WSF exposition and the lower growth of individuals exposed to WSF than the Control possibly reflects metabolic costs. The implications of the main findings for the individual and the food chain are discussed, including behavioral aspects and the role played by this predator in shallow aquatic systems.

**Keywords:** body size, ecotoxicology, growth performance, juvenile trahira, predation, survival.

1. **Introduction**

Crude oil and its products are the main source of energy in modern society and account for 41% of the global energy demand, recently estimated at 8,677 million tons ([IEA, 2012](#)). Crude oil, or petroleum, has an organic origin and is classified as light, medium or heavy, depending on the dominant type of hydrocarbon, *i.e.*, aromatic compounds containing one benzene ring, polynuclear aromatic compounds containing two or more benzene rings, or aliphatic-rich (long- or short-chain) compounds, respectively ([Marchand et al., 2002](#)). Petroleum may also contain other compounds, such as sulfur, nitrogen, and oxygen, as well as trace metals, such as nickel and...
vanadium (Johnston, 1984; Marchand et al., 2002). Most environmental disasters worldwide are related to petroleum exploration and related activities, such as transport (Wang and Fingas, 2003). The degree of an oil spill’s impact depends on the amount and type of oil involved (Marchand et al., 2002), as well as on which organisms are involved and how they are exposed to the petroleum (Silva et al., 2009; Smit et al., 2009). According to the literature, oil spills have mostly large-scale consequences, such as the loss of biodiversity and ecosystem services (Sánchez et al., 2006; Antonio et al., 2011).

In water, petroleum undergoes physical and chemical processes (e.g., evaporation, dissolution, emulsion, photolysis, biodegradation), which generate a water-soluble fraction (WSF) (Daling et al., 1990; Pérez-Cadahía et al., 2004). The WSF of petroleum and its derivative products contain a mixture of polycyclic aromatic hydrocarbons (PAH); monoaromatic hydrocarbons, often referred to as BTEX (benzene, toluene, ethylbenzene and xylenes); phenols; heterocyclic compounds containing nitrogen and sulfur; and heavy metals. In aquatic organisms, WSF of petroleum may impair physiological integrity, and its chemical compounds may be disseminated throughout the food chain (Pérez-Cadahía et al., 2004). Therefore, under petroleum contamination, ecological interactions between trophic levels may extend the effects on an individual (Omorgie and Ufodike, 2000; Akaishi et al., 2004) to higher levels of the biological organization, eventually affecting the structure and function of communities and ecosystems (Collier et al., 1996; Elliott et al., 2002; Peterson et al., 2003). Although ecotoxicological studies on aquatic organisms are increasing in number (Akaishi et al., 2004; Hansen et al., 2011; Weber et al., 2013), the evaluation of petroleum toxicity on top consumers is still a neglected topic.

Fish species are well represented in all levels of consumers in aquatic food chains; therefore, they are an essential component of nutrient cycling and energy flow in marine and continental waters (Vanni, 2002; Whittfield and Elliott, 2002). The limited capacity for dispersal of the sedentary fish species inhabiting hydrological disconnected environments makes them highly vulnerable to conditions imposed by environmental contamination. Therefore, understanding how and to what extent WSF affects consumers represents the first step in predicting the effects of contamination on aquatic food chains (Van der Oost et al., 2003).

The trahira Hoplias aff. malabaricus (Bloch) is distributed in most Central and South American watersheds (Oyakawa, 2003). By preferentially inhabiting vegetated areas that favor their ambush feeding strategy (Ferreira, 2007; Petry et al., 2010), H. aff. malabaricus may play an important role as a main predator in communities found in streams and lakes (Mazzeo et al., 2010; Corrêa et al., 2012). In these environments, H. aff. malabaricus are able to mediate prey species coexistence by piscivory (Petry et al., 2010). They might also do so in coastal brackish water bodies of southeastern Brazil, which are inhabited mostly by freshwater fish species, including trahiras (Di Dario et al., 2013).

The southeastern coast of Brazil is subjected to intensive petroleum-related activities involving drilling and transport over the Campos Basin, an area of approximately 100,000 km$^2$ between the 21st and 23rd southern parallels, which currently accounts for more than 80% of the Brazilian petroleum production (ANP, 2011). Although coastal lagoons in this region are mostly maintained by small streams, sporadic saltwater intrusion may occur via sandbar breaching events or during extreme high tides events (Caliman et al., 2010). In addition, diffuse pollutant inputs may originate from road runoff, domestic usage and licensed discharges of oil derivatives used as energy sources in urban areas (Marchand et al., 2002). Specifically, this study aimed to assess the effect of chronic exposure of Hoplias aff. malabaricus juveniles to WSF of petroleum by evaluating prey consumption rates and the conversion of acquired food into somatic growth.

2. Material and Methods

2.1. Sampling and acclimatization

Juveniles of H. aff. malabaricus were obtained in September 2012 via diurnal seining in coastal lagoons situated in northern Rio de Janeiro State. The ten juveniles ranged in size between 8.5 and 12.5 cm total length (Table 1). Acclimatization lasted 15 days and consisted of the individual maintenance of fish under a 12 h photoperiod at 25 °C in 12 L glass aquariums containing filtered and UV-sterilized freshwater. A PVC tube 50 mm in diameter was installed in the bottom of each aquarium to be used as refuge. Adults of jewel tetra Hyphessobrycon eques (Steindachner) with mean size of 2.98 ± 0.24 cm were commercially acquired and supplied to H. aff. malabaricus for consumption. Five preys were added to each aquarium, corresponding to a ratio of 2% of the predator mass per day (Monteiro et al., 2013). Feeding ceased three days before the beginning of the experiment, and non-consumed preys were removed. The number of preys consumed per fish was registered and replaced daily. Three times a day, the aquariums were checked, and the dead preys were removed and replaced. Every 72 h, the feces were removed by suction and the water level was re-established. Environmental parameters (temperature and dissolved oxygen) were measured daily with a YSI-85 meter and adjusted daily with air conditioner and pumps to 25 ± 2 °C and 6 ± 1 mg L$^{-1}$, respectively.

2.2. Water-soluble fraction of petroleum

The WSF was prepared with heavy crude oil (17° API; Trevisan et al., 2009) obtained from the Jubarte oil field situated 77 km from the southern coast of Espirito Santo State, at the Campos Basin, Brazil, following the method proposed by Anderson et al. (1974), with some modifications. One part crude oil (1.2 kg) was mixed with four parts of filtered and sterilized tap water (4.8 L) (1:4 w/v) and homogenized in the dark with a magnetic stirrer for 24 h at 500 rpm. Afterward, the mixture was
transferred to a separation funnel and left for 2 h, after which the water fraction was collected. Undiluted, this fraction was considered 100% WSF and was prepared and stored at the experimental room temperature.

A sample of the 100% WSF of petroleum was sent to the CQA laboratory (Centro de Qualidade Analítica, Campinas, Brazil) for total hydrocarbon and BTEX analysis, as well as for the evaluation of its iron, barium, strontium and sulfates contents. This laboratory employed methods for evaluating solid wastes established by the Environmental Protection Agency (EPA), American Public Health Association (APHA), American Water Works Association (AWWA) and Water Environment Federation (WEF). According to the results of the analysis, the quantity of hydrocarbons was below the level of detection (< 0.2 µg L⁻¹). The evaluation of metal contents indicated that the WSF contained 0.09 mg L⁻¹ of iron, 0.09 mg L⁻¹ of barium, 0.08 mg L⁻¹ of strontium and 16.60 mg L⁻¹ of sulfates.

### 2.3. Feeding experiment

The purpose of this assay was to evaluate the effect of WSF on prey consumption and growth in length and weight in the juveniles of *H. aff. malabaricus*. Therefore, the total length (*L*<sub>f</sub>) and weight (*W*) of each juvenile were recorded at the beginning and at the end of the experiment. The acclimatization conditions were maintained during the 28 days of the experiment, and the juveniles were distributed into two treatment groups with five replicates each: 1) the control set, 0% WSF (herein named Control); and 2) the exposed group, 50% WSF (herein named WSF). The consumption rates were obtained as the proportion of preys consumed per day and per week. The increments in length (*G*<sub>LENGTH</sub> = ln *L*<sub>f_final</sub> – ln *L*<sub>f_initial</sub>) and weight (*G*<sub>WEIGHT</sub> = ln *W*<sub>f_final</sub> – ln *W*<sub>f_initial</sub>) were considered the finite growth rates in length and weight, respectively.

Once a week, the water was exchanged for freshly prepared water containing the same proportions of WSF for each treatment. The initial length and weight of the juveniles did not differ significantly between the two treatment groups (Student *t*-test: *t*<sub>length</sub> = –0.30; d.f. = 8; *P* = 0.77; Student *t*-test: *t*<sub>weight</sub> = –0.50; d.f. = 8; *P* = 0.63). At the end of the experiment, all juveniles were anesthetized for the final measurements by submersion in filtered and sterilized freshwater containing 6.7 ml L⁻¹ of benzocaine stock solution (80 g dissolved in 1 L ethanol).

### 2.4. Statistical analysis

The consumption and finite growth rates were evaluated as the response variables in this experiment. Repeated measures analysis of variance (rm-ANOVA) was performed using the general linear modeling procedure of the statistical package STATISTICA (StatSoft, 2007) and was applied to verify the effect of the treatment and time of exposure on the juvenile consumption rates. The rm-ANOVA design considered the treatment as the main factor, whereas the weekly consumption rate was considered as the repeated measures. Student’s *t*-test was used to evaluate any possible effect of the experimental treatment on the number of prey consumed and finite growth rates. Assumptions of normality, homoscedasticity and sphericity were met (sphericity required in the rm-ANOVA) and *P* < 0.05 was adopted as the statistical significance level.

### 3. Results

No mortality was observed in the Control units, and only one juvenile died after 13 days of WSF treatment, which was excluded from all analyses. Prey consumption per juvenile showed variation along the 28 days of the experiment, showing in mean a slight increase with time (Figure 1). Although fish in the Control group consumed...
more prey overall than those in the WSF units (Table 1), differences in the total number of consumed prey between the two treatments were non-significant (Student t-test: $t = 0.42; d.f. = 7; P = 0.69$).

Juveniles in the Control consumed more preys during the first two weeks (1.6 ± 1.2 prey.day$^{-1}$) than those submitted to petroleum (1.2 ± 1.1 prey.day$^{-1}$). At the third week, the juveniles from the Control showed a decay in their prey consumption, reestablishing it at the fourth week, maintaining it again at higher levels than the WSF treated individuals (Figure 1b). According to the rm-ANOVA, there was a significant effect of time on the consumption rates of *H. aff. malabaricus* juveniles (Table 2). Again, no significant effect was observed for treatment neither for the interaction between treatment and time of exposure (Table 2).

The final length and weight of the individuals were higher than the initial values in both treatments (Table 1; Figures 2a and 2b). The mean increment in length in the Control (2.16 ± 0.53 cm) was more than twice the increase observed in the treated individuals (0.93 ± 0.38 cm) after 28 days of experiment (Table 1; Figure 2a). Significant differences in growth in length were found when the finite rates were evaluated (Figure 2c). Although the mean increment in weight was higher in the Control (11.34 ± 4.51 g) than those exposed to WSF (8.75 ± 6.35 g) (Figure 2b), differences in weight finite rates were not significant (Figure 2d).

### 4. Discussion

Several studies have assessed the effects of petroleum either as dispersed particles or as a soluble fraction (Collier et al., 1996; Heintz et al., 2000; Akaishi et al., 2004; Pérez-Cadahía et al., 2004; Hansen et al., 2011). Any intrusion of petroleum (e.g., from accidental spills on coastal waters) into enclosed water bodies will delay dispersion and increase the persistence of this pollutant. By inhabiting coastal lagoons, the trahira *Hoplias aff. malabaricus* may experience this source of contamination.

A large number of studies demonstrating that WSF reduces water quality and the well-being of aquatic organisms are available. The negative effects of petroleum and its compounds on feeding and conversion of food in somatic growth have been demonstrated experimentally in a wide range of fish species, including species occupying marine, estuarine and freshwater environments (Heintz et al., 2000; Omoregie and Ufodike, 2000; Saborido-Rey et al., 2007). In addition, any effect on the individual-level might affect their development, progeny and ecological interactions with other individuals. The reduction in growth in length of *H. aff. malabaricus* juveniles exposed experimentally to petroleum observed in the present study adds to the limited amount of available evidence regarding the effect of petroleum on top predators at the whole organism level.

Although the juveniles exposed to WSF of petroleum in the present study grew less in length than the controls, they did not show differences in either consumption rates or weight. In the natural environment, the trahira *H. aff. malabaricus* exhibits ambush feeding on fish species, thereby expending low levels of energy on predation (Ferreira, 2007; Ramsdorf et al., 2009; Petry et al., 2010). The small artificial environment in which this experiment was carried out favored prey detection by the juveniles, which did not spend much energy on feeding activity. This finding related to the feeding behavior might explain in part the absence of a conspicuous effect from the treatment on consumption rates. Omoregie and Ufodike (2000) found that juveniles of the widespread introduced Nile tilapia *Oreochromis niloticus* (Linnaeus) fed less aggressively and ingested only 80% of the supplied food under the highest concentrations of WSF. The susceptibility to the experimental routine (weekly water and WSF renewal) was observed in the juveniles of the trahira *H. aff. malabaricus* mainly at the end of the first two weeks, by the decrease in predation rates in both treatments (Figure 1a). The increase in predation rate observed in both treatments on the last

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**Figure 1.** Prey consumption rates of *Hoplias aff. malabaricus* juveniles (Mean ± S.E.) after 28 days in exposition to two levels of treatment (Control and WSF). (a) Number of preys consumed daily; (b) Number of preys consumed weekly. Letters above bars indicate the differences according to the rm-ANOVA.

**Table 2.** Effects of water-soluble fraction of petroleum on weekly consumption rates (number of preys consumed per week) in juveniles of trahira *Hoplias aff. malabaricus* exposed for four weeks (28 days) to two levels of treatment (Control and WSF), as determined by rm-ANOVA.

<table>
<thead>
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<th>Effect</th>
<th>SS</th>
<th>d.f.</th>
<th>MS</th>
<th>F</th>
<th>P</th>
</tr>
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</tr>
<tr>
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<td>1</td>
<td>0.02</td>
<td>0.18</td>
<td>0.69</td>
</tr>
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<td>Error</td>
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<tr>
<td>Repeated measure</td>
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<td>0.15</td>
<td>32.59</td>
<td>&lt;0.01</td>
</tr>
<tr>
<td>Time</td>
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<td>0.01</td>
<td>2.77</td>
<td>0.07</td>
</tr>
<tr>
<td>Treatment × Time</td>
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<td>21</td>
<td>0.01</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>
week of the experiment was expected, due to fish growth (Loureiro and Hahn, 1996; Corrêa et al., 2012).

The absence of significant differences in weight increment between the treatments in this study may be expected because there is more investment of energy in growth in length rather than growth in weight (fat reserves) during the juvenile ontogenetic stage (Teixeira de Mello et al., 2006). Therefore, differences would be expected between the two treatment groups with respect to length growth rate in juveniles of *H. aff. malabaricus*. In fact, trahiras exposed to WSF in this study grew significantly less in length than those of the Control group. This outcome strongly suggests a negative effect of WSF on metabolism. Individuals in both treatments maintained similar consumption rates; therefore, we assume that those exposed to petroleum were unable to efficiently allocate the energy obtained from food to somatic growth in length. Nwabueze and Agbogidi (2010) reported a lack of growth in length for the African catfish *Heterobranchus bidorsalis* Geoffroy St. Hilaire submitted to concentrations over 50% WSF. We suppose that juveniles of *H. aff. malabaricus* exposed to petroleum used their energy for physiological compensation, rather than investing in somatic growth, which might be related to the activation of their detoxification mechanisms, as observed in juveniles of other fish species (Omoregie and Ufodike, 2000; Saborido-Rey et al., 2007; Esenowo and Ugwumba, 2010).

The detoxification metabolisms are associated with the commitment to homeostasis in individuals exposed to pollutants (Esenowo and Ugwumba, 2010). The degree of accumulation and retention of organic contaminants such as hydrocarbons is primarily dependent on the overall content and their distribution in tissues (Elliott et al., 2002). In mussels, it has been found that even low levels of PAHs (5 ng L$^{-1}$) cause bioaccumulation and an increase in the levels of oxygenase enzymes involved in biotransformation (Pérez-Cadahía et al., 2004). After reviewing physiological and performance indicators of stress in fish, Schreck (1990) states that individuals may compensate for sublethal chronic conditions, recover in sublethal acute conditions or become exhausted and die.

Although this study detected a moderate effect of WSF in *H. aff. malabaricus*, the subtle changes in somatic growth may be important in the long term and may have consequences on the population level, such as reduced growth or population decline (Elliott et al., 2002). According to Heintz et al. (2000), recruitment and

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**Figure 2.** Growth of juveniles of *Hoplias aff. malabaricus* (Mean ± S.E.) after 28 days exposed to two levels of treatment (Control and WSF). (a) Increase in total length (cm); (b) Total weight (g); (c) Growth in length (finite rates in length = ln $L_{T_{\text{final}}}$ - ln $L_{T_{\text{initial}}}$); (d) Growth in weight (finite rates in weight = ln $W_{\text{final}}$ - ln $W_{\text{initial}}$).
population dynamics may be affected by sublethal effects, such as reduction in growth.

The sensitivity of fish to petroleum was evaluated early on in Alaskan freshwater and anadromous species, and it was experimentally demonstrated that the median tolerance limit at 96 h (i.e., the concentration at which half of the animals survived the given period of exposure) may vary from one species to the other (Moles et al., 1979). These authors also observed that salmonids (the most sensitive group) maintained in saltwater were more sensitive to petroleum than those tested in freshwater, demonstrating that salinity also has an important effect. Several studies (Rantin et al., 1993; Luz and Portella, 2005) have reported that erythrinids, especially H. aff. malabancus, have a high degree of resistance to environmental conditions involving hypoxia, long-term feeding deprivation and temperature extremes (Rantin et al., 1992; Rios et al., 2002; Rios et al., 2004; Petry et al., 2007), which negatively affect most of the sympatric fish species (Soares et al., 2006). The resistance characteristics of the trahira are most likely related to its sedentary behavior and low metabolism in contrast to the highly migrant and more active salmonids. The differences in energy expenditure between these aquatic top predators may explain, at least in part, their differences in susceptibility to petroleum.

It is thus concluded that 28 days of exposure to 50% WSF was non-lethal to juvenile H. aff. malabancus, neither affecting prey consumption nor the growth rate in weight. Nevertheless, it affected negatively the growth rate in length of the individuals. Considering the ecological importance of the trahira as predator and potential keystone, the negative effect of WSF of petroleum reported herein may in long term affect the stability of its communities.

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