Acute toxicity test of agricultural pesticides on silver catfish (*Rhamdia quelen*) fingerlings

Teste de toxicidade aguda de pesticidas agrícolas em alevinos de jundiá (*Rhamdia quelen*)

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- NOTE -

ABSTRACT

Toxicity risks of agricultural pesticides to fishes are pivotal. Currently, many questions remain unsolved regarding the toxicity of commonly used pesticides to silver catfish (*Rhamdia quelen*), a South American catfish. The present studies have been designed to investigate the acute toxicity and the lethal concentration (LC₅₀) of four herbicides, two fungicides and two insecticides to silver catfish fingerlings. All experiments were carried out in triplicates, in a static bioassay system, using commercially available pesticides. The data was analyzed through the Trimmed Spearman-Karber method available from the Environmental Protection Agency. The 96hLC₅₀ and 95% lower and upper confidence limits, respectively, for the following pesticides were determined: glyphosate (7.3mg L⁻¹; 6.5–8.3), atrazine (10.2mg L⁻¹; 9.1–11.5), atrazine+simazine (10.5mg L⁻¹; 8.9–12.4), mesotrione (532.0mg L⁻¹; 476.5–594), tebuconazole (5.3mg L⁻¹; 4.9–5.7), metilparation (4.8mg L⁻¹; 4.3–5.3), estrobilurina and triazol (9.9mg L⁻¹; 8.7–11.2). Diflubenzuron was also tested and caused no fish mortality up to 1g L⁻¹. The toxic concentration of these pesticides to silver catfish fingerlings fell above the concentration used for application in the field and, except following accidental application or misplacing of empty recipients, it should not cause fish mortality. Nonetheless, the data obtained will be useful to study the long-term effect of these products on the hematological, biochemical, hormonal and immunological parameters of silver catfish and related fish species in South Brazil.

Key words: bioassay, toxicity, pesticides, LC₅₀ *Rhamdia quelen*.

INTRODUCTION

The silver catfish (*Rhamdia quelen*), a *Heptapteridae* fish, is one of the most widespread inhabitants of South American Rivers. In artificial pond cultures, silver catfish presents high reproduction rate
and fast weight gain, mainly in the warmer months of the year (GOMES et al., 2000). This species might be used as a model to improve management of several fish species of this family. Unfortunately, most artificial ponds used for fish culture are located close to or inside agricultural areas, or are fed with water springs that run through cultivated soil. Because of modern pest management practices, large amounts of herbicides, fungicides and pesticides are used in these areas for crop protection. In addition, some pesticides might be added directly to water to control macrophytes and predatory insects (SZAREK et al., 2000). As a result, small amounts of these products might be found in waters used for fish culture (VAN DER OOST et al., 2003). Contamination of water with large amounts of pesticides leads to fish mortality but the effects of small amounts are mostly unknown. Traditionally, survival, growth and reproduction of individuals are chosen as endpoints of the classic laboratory tests for ecotoxicity (VAN DER OOST et al., 2003). Chemicals used in agriculture may affect fish communities by altering species composition of plankton communities. In addition, exposures may also result in a disturbance of the reproductive endocrine systems (KIME et al., 1995).

To date, few data are available regarding toxicity of herbicides, insecticides and fungicides to *R. quelen*. In irrigated rice culture commingled with silver catfish, herbicides are applied directly to the water and may affect aquatic life. The toxicity of herbicides used in irrigated rice culture has been recently evaluated (MIRON et al., 2005), and a few of these compounds were considered non-toxic for silver catfish at concentrations considered effective for weeds. In this research, the focus was to investigate the acute toxicity of pesticides commonly used in soy wheat, and corn cultures in which the product might reach water springs or ponds accidentally or through runoff of soil particles after rain (VAN DER OOST et al., 2003). Thus, the aim of this study was to determine the LC50 of commonly used agricultural products in *R. quelen* fingerlings. These compounds were chosen according to their importance for agriculture in Southern Brazil, which is based on soybean, corn, and wheat production.

**MATERIALS AND METHODS**

This study was conducted between July 2004 and August 2005, at the facilities of the Universidade de Passo Fundo, Rio Grande do Sul, Brazil (28°15’S / 52°24’W, 687 m above sea level). The fish used in the study were 60-day-old mixed-sex silver catfish fingerlings weighing between 2 and 4g. They were kept in 500-L fiberglass tanks up to distribution into experimental aquaria. Water exchange rates of 20% were used each day, at the same time as food wastes were suctioned from the tanks. During an acclimation period of 7 days, the fish were kept under natural photoperiod and fed two times a day (10:00 and 16:00h) at 5% of body weight with commercial extruded food (42% crude protein, 3,400Kcal kg⁻¹ DE).

All water parameters were checked daily before introduction of fingerlings and up to the time the product was applied to the water. Water temperature and dissolved oxygen concentrations were measured with an YSI model 550A oxygen meter (Yellow Spring Instruments, USA). The pH values (Bernauer pH meter), total ammonia-N (colorimetric test), total alkalinity and hardness were also measured.

For the LC50 determinations, 210 fingerlings were uniformly distributed in 21 40-L plastic aquaria, keeping fish density below or equal to 1g L⁻¹, according to the Brazilian Association for Technical Rules (ABNT). Each product was tested using 5 to 6 different concentrations, with 3 repetitions each. Three aquaria were kept as control (without herbicide). Fingerlings were observed at 12 h intervals, for 96h (acute toxicity) when the test was concluded. During the experimental period, fingerlings were not fed and water exchange was stopped.

All products used were purchased from local stores. The generic, commercial, and chemical names, and pesticide group of each product tested are shown in table 1, and the concentrations used are shown in table 2. The concentration used for each trial was calculated using the concentration (g L⁻¹) stated on the product’s label. Before addition, each product was mixed in a small volume of water from each aquarium and then added to the water using a glass pipette. Fingerlings were then observed for 96h and the mortality recorded; swimming behavior (normal, erratic swimming, lethargy) was checked, recorded and compared to the control group.

All dead fish were frozen and then shipped to the biological garbage collector. The fish that remained alive after each experiment were killed by thermal shock in ice-cold water and discarded as described above. After each toxicity trial, the water contaminated was kept for at least 30 days in fiberglass tanks and then percolated in septic ponds. After each experiment, aquaria were cleaned with running water followed by rinsing with ethanol. Before reusing, aquaria were filled with water and tested for remaining toxicity by adding silver catfish fingerlings that were observed for at least 5 days for mortality or behavioral changes.
The 96-h LC\(_{50}\) for each pesticide was calculated based on the mortality data, recorded at 12 hr intervals for each concentration of the product, using the Trimmed Spearman-Karber method (Version 1.5) available from the Environmental Protection Agency (USA). Comparisons of water pH or alkalinity among the different treatments were made by one-way analysis of variance and Tukey test. Analysis was performed using the software InsTat (Sigma), and the minimum significance level was set at P<0.05.

**RESULTS**

Throughout all the trials, the water temperature averaged 22±2\(^\circ\)C, pH ranged from 6.2 to 7.0, dissolved oxygen ranged from 5.6 to 7.5mg L\(^{-1}\) and total ammonia was lower than 0.5mg L\(^{-1}\). The total hardness and alkalinity were 60 and 65mg CaCO\(_3\), respectively. All values were within the acceptable limits for fish culture in pond water as reported previously (BOYD, 1982). None of the products, even at the highest concentration used, altered the water quality parameters. Lethargy, swimming at the water surface and erratic swimming (mainly vertical swimming) were the main behavioral changes observed throughout the experiment, in the presence of tebuconazole, strobilurin plus triazol, glyphosate and atrazine or atrazine plus simazine; hyper excitability was observed in fish exposed to methyl-parathion and increased abdominal volume was observed in fish exposed to atrazine or atrazine plus simazine (data not shown). Tebuconazole, at 16mg L\(^{-1}\), caused fish death almost immediately following addition to the tank. The behavioral changes were observed with different pesticides, usually at the higher concentrations tested, but were not used to assess the effects of the products. The nominal concentration of each pesticide, the concentrations tested for toxicity, the 96-h LC\(_{50}\) obtained for each product, and the concentration used in the field are depicted in table 1 and 2, respectively.

**DISCUSSION**

In South Brazil, most fish ponds are still built in wetlands inside agricultural areas. Water contamination with agricultural pesticides is a potential threat to productivity and a major cause of fish mortality. However, water contamination with pesticides at non-lethal concentrations might pass unnoticed except for loss in productivity, which, in most cases, might be difficult to assess. In addition, there are no data on the accumulation of such chemicals in fish and how this could affect human health. Thus, it becomes necessary to determine the concentration of agricultural pesticides capable of affecting fish biochemical and physiological parameters that contribute for productivity losses. With this in mind, the 96-h LC\(_{50}\) of several commonly used agricultural pesticides were determined in this study. Silver catfish fingerlings were used because this fish species is ubiquitous in rivers and ponds in South Brazil and has been intensively cultivated for commercial purposes.

The most toxic product tested was Folidol 600 (methyl-parathion, 600g L\(^{-1}\); table 2), which is used in fish culture ponds to kill the aquatic larval stages of predatory insects that threaten fish larvae. The Folidol 96 h LC\(_{50}\) was 4.8mg L\(^{-1}\), a value similar to that found previously (MURTY et al., 1984) for Mystus cavasius (5.9mg L\(^{-1}\)) and lower than that found for the mosquito fish Gambusia affinis (8.4mg L\(^{-1}\)) (BOONE & CHAMBERS, 1997). The acute effects of methyl-parathion were also determined for matrinxã (Brycon cephalus) in which, besides major behavioral changes, the 96 h LC\(_{50}\) was determined at 6.0mg L\(^{-1}\) (AGUIAR et al., 1997).

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**Table 1 - Pesticides tested for acute toxicity on Rhamdia quelen fingerlings.**

<table>
<thead>
<tr>
<th>Generic name</th>
<th>Commercial name*</th>
<th>Chemical name</th>
<th>Usage</th>
</tr>
</thead>
<tbody>
<tr>
<td>Glyphosate</td>
<td>Roundup</td>
<td>N-phosphonomethylglycine</td>
<td>Herbicide</td>
</tr>
<tr>
<td>Atrazine</td>
<td>Atrazina</td>
<td>2-chloro-4-ethylamine-6-isopropylamino-S-triazine</td>
<td>Herbicide</td>
</tr>
<tr>
<td>Atrazine + Simazine</td>
<td>Herbimix</td>
<td>6-chloro-N,N-diethyl-1,3,5-triazine-2,4-diamine + 2-chloro-4-ethylamine-6-isopropylamino-S-triazine</td>
<td>Herbicide</td>
</tr>
<tr>
<td>Mesotrione</td>
<td>Callisto</td>
<td>[2-[4-(methylsulfonyl)-2-2 nitrobenzoyl]-1,3 cyclohexanedione</td>
<td>Herbicide</td>
</tr>
<tr>
<td>Methyl-parathion</td>
<td>Folidol 600</td>
<td>O-O-dimethyl O-4-nitrophenyl thiophosphate</td>
<td>Insecticide</td>
</tr>
<tr>
<td>Diflubenzuron</td>
<td>Dimilin</td>
<td>1-(4-chlorophenyl)-3-(2,6 diflurobenzol) urea</td>
<td>Insecticide</td>
</tr>
<tr>
<td>Tebuconazole</td>
<td>Folicur</td>
<td>2-[2-(4-chlorophenyl)ethyl]-3,3-dimethyl-1-(1H-1,2,4-triazol-1-yl)buta n-2-ol</td>
<td>Fungicide</td>
</tr>
<tr>
<td>Strobilurin and triazol</td>
<td>Opera</td>
<td>Pyraclostrobin.methyl N-(2-[1-1(4-chlorophenyl)-1h-pyrazol-3 yl]oxymethyl[phenyl]N-methoxy carbamate</td>
<td>Fungicide</td>
</tr>
</tbody>
</table>

* Commercial names might be trademark protected by law. All products were purchased on local stores.
Atrazine is one of the most widely used herbicides and, because of its considerable persistence and mobility in soil and water, it is considered a common terrestrial and aquatic contaminant (OULMI et al., 1995). The 96-h LC50 of atrazine for trout embryos and larvae ranged from 0.87 to 1.11 mg L−1, and concentrations as low as 10 μg L−1 caused kidney damage in chronic exposed rainbow trout (OULMI et al., 1995). For *Tilapia mossambicus*, the atrazine 96-h LC50 was 8.8 mg L−1 (PRASSAD & REDDY, 1994); chronic exposure effects were tested on Tilapia using 1/8 of this dose and several disturbances in osmotic balance of exposed fish were found. In common carp (*Cyprinus carpio*) the 96-h LC50 was 18.8 mg L−1 (NESKIVICK et al., 1993). According to these data, the 96-h LC50 of atrazine for *R. quelen* (10.2 mg L−1) was similar to *T. mossambicus*; however, *R. quelen* was more sensitive to atrazine than *C. carpio* but more resistant than rainbow trout.

The 72-h LC50 of simazine for the larval stage of *Sparus aurata* was 4.19 mg L−1 (ARUFE et al., 2004); relevant behavioral changes were also noticed for *R. quelen* in the present study, mostly related to erratic swimming and hyper excitation. Taken together, the low 96-h LC50, combined with the strong behavioral changes and the fact that Folidol is frequently used directly in water, indicate that methyl-parathion is a potentially harmful compound for fish, including *R. quelen*.

Atrazine (250 mg L−1) + Simazine (250 mg L−1) 1, 2, 4, 8, 16, 32 10.5 (8.9 – 12.4) 1500 - 1750 b
Atrazine (250 mg L−1) + Simazine (250 mg L−1) 1, 2, 4, 8, 16 10.2 (9.0 – 11.5) 1000 - 1500
Atrazine (250 mg L−1) + Simazine (250 mg L−1) 1, 2, 4, 8, 16, 32 10.5 (8.9 – 12.4) 1500 - 1750 b

The concentrations of each product used in each experimental trial and 96-h LC50 values determined based on the active ingredient of each product. The concentration of each product tested was calculated based on the concentration of the active ingredient of each product as stated on the label. Concentrations tested (mg L−1) 96h LC50 (mg L−1) (95% lower and upper confidence interval) Concentration to target species (g ha−1)

Concentrations of each product used in each experimental trial and 96-h LC50 values determined based on the active ingredient of each product. The concentration of each product tested was calculated based on the concentration of the active ingredient of each product as stated on the label. Concentrations tested (mg L−1) 96h LC50 (mg L−1) (95% lower and upper confidence interval) Concentration to target species (g ha−1)

Table 2 - Environmental lethal concentrations of agricultural pesticides in *Rhamdia quelen*.

Acute toxicity test of agricultural pesticides on silver catfish (*Rhamdia quelen*) fingerlings. (MOORE & LOWER, 2001). A formulation containing only simazine was not available commercially in South Brazil at the time the experiments were carried out; thus, the 96-h LC50 of simazine alone on silver catfish fingerlings has not been determined.

The 96-h LC50 determined for the glyphosate-based herbicide Roundup®, in *R. quelen* (7.3 mg L−1) was much lower than that reported by other authors in other fish species as *Lepomis macrochirus* (120 mg L−1) and *Oncorhynchus mykiss* (86 mg L−1) (HUMBURG & COLBY, 1989). For commercial formulations, the LC50 found varied from 3 to 197 mg L−1 in rainbow trout and from 13 to 33 mg L−1 in coho salmon (HOLTBY & BAILLIE, 1989). For the *Gambusia yucatanana*, the 96-h LC50 of glyphosate-based formulae was 17.8 mg L−1 (RENDÓN-VAN-OSTEN et al., 2005), a concentration closer to the values reported here. One of the reasons for the higher toxicity of commercial glyphosate-based formulases might be attributed to the presence of the surfactant (POEA), which is more toxic than glyphosate itself. The 96-h LC50 for isolated POEA on rainbow trout (*O. mykiss*) ranged from 0.65 to 7.4 mg L−1 (giesy et al., 2000); for commercial Roundup®, the 96-h LC50 ranged from 8.7 to 27 mg L−1 and for pure glyphosate salt the 96-h LC50 ranged from 140 to 240 mg L−1 (giesy et al., 2000). These data show the higher toxicity of POEA, in agreement with data already reported (FOLMAR et al., 1979; TSUI & CHU, 2003). In addition, the isolated glyphosate salt, because of its rapid degradation in water but relatively high stability in the soil, was described as secure for fish (giesy et al., 2000). For *R. quelen*, Roundup® was found to have a low 96-h LC50 and more studies on active ingredients will be conducted in the near future.

Tebuconazole is a fungicide used also in plant cultures or as wood preservative against fungi
and insects (LEBOKOWSKA et al., 2003). The 48h LC$_{50}$ of tebuconazole-based formulae for Poecilia reticulata is 8.5 mg/L and the 96h LC$_{50}$ is 4.5 mg L$^{-1}$ (LEBOKOWSKA et al., 2003). This concentration is higher than that reported for R. quelen (4.76 mg L$^{-1}$), probably due to the fact that the used commercial formulation to P. reticulata was less concentrated. In addition, according to previous data (LEBOKOWSKA et al., 2003), the LC$_{50}$ of tebuconazole alone, for unspecified fish species, ranged from 1.6 to 8.7 mg L$^{-1}$; thus, the LC$_{50}$ of tebuconazole obtained for R. quelen is within this range.

Data on acute toxicity of mesotrione, diflubenzurin and strobilurin plus triazol for fish are scarce. The 96-h LC$_{50}$ of mesotrione for R. quelen was higher than 500 mg L$^{-1}$, a concentration higher than that achieved during field application. Mesotrione has been reported to be almost non-toxic to warm and cold-water fish (EPA, 2001) and the toxicity of strobilurin plus triazol has been studied only in rats and was considered slightly toxic by the oral route. Similarly, concentrations up to 1000 mg L$^{-1}$ of dimilin (diflubenzurin) caused no mortality in silver catfish fingerlings during the 96h exposure time.

It is worth mentioning that most differences in toxicity of the pesticides might be attributed either to the formulæ used, water quality parameters or to the biochemical pathway affected by the drug and the fish species used, that might be naturally less or more tolerant to water contamination.

Regarding behavioral changes, the most prominent effect of methyl-parathion on R. quelen was the induction of abnormal, erratic swimming and hyper excitability. A similar effect was also reported for M. cavasius with the same product (MURTY et al., 1984). This erratic swimming occurs most likely due to acetylcholinesterase (AChE) inhibition by organophosphorus compounds, as described for several fish species (MURTY et al., 1984; AGUIAR et al., 2004; RENDÓN-VAN-OSTEN et al., 2005) including R. quelen (MIRON et al., 2005). Atrazine was also reported as inducing behavioral abnormalities in Carassius auratus, most commonly erratic swimming, as also observed in the present work, after short-term exposure to sub lethal concentrations (SAGLIO & TRIJASSE, 1998; GREYMORE et al., 2001).

Thus, because the compounds tested are widely applied around the world in soybean, wheat and corn cultures or other agricultural activities and since the Heptapteridae family represents an important fish family cultivated all over the world, the data on acute toxicity here presented might be useful to study the effects on several biochemical, hematological, immunological and physiological parameters of fish during acute or long term exposure with sub lethal doses of these products. In addition, it will provide tools for further studies considering environmental risk assessment (ERA).

**CONCLUSIONS**

Agricultural pesticides were toxic to silver catfish fingerlings at concentrations higher than that used against their target species; nonetheless, water run off or accidental spilling might have deleterious effect on silver catfish raised in ponds within agricultural areas. Further research on the effect of sub-lethal concentration of selected pesticides is being held to investigate bioaccumulation and the effect at the molecular level.

**BIOETHICS AND BIOSecurity Committee Approval**

This study has been approved by the Ethic Committee Research of the Universidade de Passo Fundo (Nº 630/CONEP), RS and has been developed in accordance with national and institutional guidelines for the protection of human subjects and animal welfare.

**ACKNOWLEDGEMENTS**

This work was supported with grants from SCI-EMBRAPA 0149/2001/2 (Secretaria de Cooperação Internacional, Empresa Brasileira de Pesquisa Agropecuária) and Universidade de Passo Fundo (UPF). All undergraduate students have scientific fellowships (Deniz Anziliero, Pibic/CNPq; Monique Lorenson and Ariane Marteninghe, Pibic/UPF, Daiana Martins and Tális de Oliveira, Pivc/UPF). Leonardo Bologna da Silva is a MSc student and has a CNPq fellowship. CNPq granted Fellowship to LCK (300259/2003-4) and LJGB (305905/2006-6).

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