ABSTRACT – This study was conducted in a forest under restoration process, which belongs to the company Holcim Brasil S/A, in the municipality of Barroso, state of Minas Gerais (21°00' to 22°00'S and 43°00' to 44°00'W), where 40 plots (2 x 2 m) were set, spaced at 10 m, forming eight strata parallel to the watercourse present in the area. Floristic composition and natural regeneration stratum were characterized, and the formed strata allowed evaluating whether the riparian vegetation and watercourse influence on the local regeneration. It was found 162 individuals of 13 families, 18 genera and 22 species, and 10,125 individuals/ha were estimated. Successional classes from pioneer and early secondary and zoochorous dispersion syndrome prevailed among species and individuals. The watercourse and riparian vegetation did not exercise significant influence (p > 0.05) on the number of species and regenerating individuals among the different strata of the forest. The diversity index of Shannon-Wiener (H') and equability of Pielou (J') were 2.691 and 0.870, respectively. The species *Psidium guajava* and Myrtaceae families presented the highest VI (value of importance). Natural regeneration analysis showed the low floristic diversity in the area, suggesting that corrective management actions should be adopted.

Keywords: Ecological restoration; Forest restoration; Restoration indicators.
1. INTRODUCTION

Despite the economic importance, limestone mining can promote considerable negative environmental impacts such as the suppression of native vegetation for the opening of new mining areas (MARTINS NETO; RAMALHO, 2010). Therefore, on the legal scope, there has been a particular concern about mining activities, as for example, the Law 11.428/2006, which provides the adoption of a compensatory measure that, includes forest restoration of an area equivalent to the mining project area.

Once the process of restoring these environments starts, it is necessary to evaluate and monitor them to check if the proposed goals were met. For this purpose, the evaluation markers and monitoring tools were used. Those are tools used to detect whether the process of succession needs or does not need intervention (MARTINS, 2009), and to conclude whether the environment has already a position for self-supporting, as well.

The regenerating stratum is young individuals or seedlings that inhabit the forest understory and allows the self-perpetuation of plant community over time (RODRIGUES et al., 2009). The evaluation of the natural regeneration under the canopy of areas at restoration process is an important tool for analyzing the evolution of restoration communities (MELO; DURIGAN, 2007) according to Marangon et al. (2008), regeneration results from the interaction of restoring natural processes of the ecosystem, being part of the development cycle and the forest setting up.

On the other hand, the identification of ecological barriers that prevent or hinder natural regeneration and decrease the self-perpetuation of the ecosystem is an essential factor for the adaptation of restoration methods and for the adoption of correction management techniques (ENGEL and PARROTA, 2003; RODRIGUES et al., 2011).

The objective of this study was to use natural regeneration stratum as an indicator of restoration process performed as environmental compensation for limestone mining in the municipality of Barroso, state of Minas Gerais, through the floristic and structural analysis.

2. MATERIAL AND METHODS

2.1. Study area

The study was carried out in an area under seven years of restoration process, which belongs to the company Holcim Brasil S/A in the municipality of Barroso (21º00’ to 22º00’S and 43º00’ to 44º00’ W), state of Minas Gerais, in the middle region of Campos das Vertentes. The climate is classified as Cwb (Köppen), that is, mesothermal with well-defined seasons (OLIVEIRA-FILHO; MACHADO, 1993). The average annual temperature in the city is 18.8°C and average annual rainfall of 1300 mm. The soils in the area are predominantly Yellow-Red Latosol and haplic cambisol (EMBRAPA, 2006). The altitude in the municipality ranges from 900 m to 1,200 m. As for the vegetation, the occurrence of semi-deciduous forest, riparian forest and Cerrado-field (Brazilian savannah) stand out (MENINI NETO et al., 2004).

The place of study presents approximately 13 ha and its surroundings consist of pastures, mine dumps, a stream named “Córrego do Monjole” as well as areas of natural regeneration process, belonging to the company Holcim Brasil S/A. The vegetation in the study area results from a planting of native and exotic species (Table 1), spaced by 3 x 3 m, totaling approximately 555 individuals and 30 species. The implementation was carried out in 2006 by the company Holcim Brasil S/A as a compensatory measure for limestone mining carried out by the company. Before restoration, the site was dominated by Urochloa sp. pasture and some individual trees located on the banks of the watercourse.

2.2. Characterization of natural regeneration

At the study site, a 0.5 ha area was delimited, where 40 plots of 2x2m, spaced at 10 m were enclosed for the evaluation of regenerating individuals. Therefore, eight strata were formed parallel to the watercourse, composed of five plots, where the stratum 1 was the closest and stratum 8 the most distant from the stream (Figure 1) in order to evaluate the interference of the watercourse as well as the riparian vegetation on the natural regeneration of forest under restoration.

All shrub-tree individuals with a height equal to or higher than 30 cm and CBH (circumference at breast height of 1.30) of less than 15.0 cm, present in the plots were identified and had their respective diameter measured at the ground level and total height measured. The sampled species were further classified into successional categories: pioneers, early secondary, late secondary and unclassified (GANDOLFI et al., 1995; MARTINS; RODRIGUES, 2005) and, according to their seed dispersal syndromes: zoochoric, anemochory and autochory (van der PIJL, 1982).
Table 1 – List of species used in planting the forest under restoration process, company Holcim Brasil S/A, Barroso, MG.

By using Fitopac 2.1 software (SHEPHERD, 2010), the following parameters were calculated: density, frequency, dominance, importance value, described by Mueller-Dombois and Ellenberg (1974), the diversity index of Shannon-Wiener (H') (MAGURRAN, 1998) and equability (J') (PEILOU, 1975).

Sum and percentage of individuals and species were obtained in each successional category and dispersion syndrome. To verify the effect of distance of the watercourse and riparian vegetation on species richness and number of regenerating individuals, each stratum was considered a treatment and each plot, a replicate. The means calculated for the number of species and the number of individuals per stratum were compared using analysis of variance (ANOVA) using the F test. Later, the Tukey test was applied at a level of 5% of significance. Statistical analyzes were performed using STATISTICA 7.0 software (STATSOFT, 2004).

3. RESULTS

This study sampled 162 individuals of 13 families, 18 genera and 22 species (Table 2), where all species were native. An estimated density of 10,125 individuals/ha was estimated. Out of the 30 species used in the planting, 24 (80%) were not found in the stratum of natural regeneration, and only Lithrea molleoides, Schinus terebinthifolia, Croton urucurana, Inga vera, Trichilia catigua and Psidium guajava, common species for planting and the regeneration stratum, were found.

Regarding the successional classification, it is observed the almost totality of species distributed in pioneer species (11 species, 50%) and early secondary (9 species; 40.9%), where only 2 (9.1%) species belong to late secondary class. At the level of individuals, the largest proportion found in the study was the pioneer class (107 individuals, 66%), followed by early Secondary class (48 individuals; 29.6%) and late secondary
The richness of sampled species and density of individuals were higher than those observed in other natural regeneration surveys conducted in environments where seedlings were planted at similar restoration time (SIQUEIRA, 2002; BASTOS, 2010). Yet, these parameters were quite lower than those found in seasonal mature semideciduous forests (MARTINS et al., 2008; GARCIA et al., 2011). These observations prove that the restoration process is advancing, however, many species still need to be recruited from the seed bank or from external sources, so that the area under restoration resembles forests at a more advanced stage of succession.

The low floristic similarity among planted and regenerating individuals may be related to reproductive immaturity of those introduced in the planting, which are not producing propagules able to compose the regenerating stratum yet. The sampling method may also have affected because the plots of the natural regeneration may have not sampled species that present an essentially aggregated pattern. In addition, the dense layer of invasive alien grasses present in the study area, especially Melinis minutiflora, may be inhibiting the regeneration of shrub-tree species. This result proves that the seedlings coming from external sources are essential to the restoration process of this environment, highlighting the importance of different sources of propagules (local and immigrant) in the maintenance of floristic diversity in secondary forests (ALVEZ; METZGER, 2006).

The predominance of pioneer and early secondary species in the regenerating stratum is due to some features such as: abundant and continuous production of seed throughout the year, dispersion over great distances, formation of seed bank and high phenotypic plasticity (WHITMORE, 1990). The presence of these species is important for the restoration process because, due to rapid growth, they are able to shade the ground in a short time by naturally controlling the appearance of weeds and providing the staking of ombrophilous plants. In addition, because of their relatively short life cycle, they produce a considerable amount of biomass that, in turn, will turn into organic matter, which will be incorporated to the soil. The large amount of produced fruit will serve as food, mainly for birds, ensuring seed dispersal and the dynamics of the implanted forest (GOLÇALVES et al., 2005).
The results found for dispersion syndrome corroborate other natural regeneration surveys carried out in semideciduous seasonal forests (BASTOS, 2010; FERREIRA et al., 2010; MIRANDA NETO, 2011) since tropical forests are characterized by presenting high proportions of plant species whose dispersion is performed by animals (STEFANELLO et al., 2010). The importance of zoochory dispersion syndrome for the restoration of the environment lies in the fact that this process allows the displacement of the propagule away from the mother plant, which reduces competition for local resources and potential rate of predation by insects or rodents (JANZEN, 1970; HOWE et al., 1985).

Although not significant (p > 0.05), the strata closest to the watercourse had the highest means for number of species and individuals (Figure 2).

The Shannon-Wiener diversity index (H’) demonstrates that the study area has low diversity of species. However, the result was higher than that found by Bastos (2010), in areas at seven (H’ = 2.48) and nine years (H = 1.60) of restoration. It is important to aim for a high diversity of species in restored...

### Table 2 – Floristic composition of the species sampled in the regeneration stratum of a forest under restoration process, in the municipality of Barroso, state of Minas Gerais.

<table>
<thead>
<tr>
<th>Family</th>
<th>Species</th>
<th>NI</th>
<th>CS</th>
<th>SD</th>
<th>VI (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Anacardiaceae</td>
<td>Lithrea molleoides (Vell.) Engl.</td>
<td>21</td>
<td>P</td>
<td>Zoo</td>
<td>12,32</td>
</tr>
<tr>
<td></td>
<td>Schinus terebinthifolia Raddi</td>
<td>10</td>
<td>P</td>
<td>Zoo</td>
<td>7,88</td>
</tr>
<tr>
<td></td>
<td>Tapirira obtusa (Benth.) J.D. Mitch.</td>
<td>14</td>
<td>Si</td>
<td>Zoo</td>
<td>6,76</td>
</tr>
<tr>
<td>Asteraceae</td>
<td>Baccharis dracunculifolia DC.</td>
<td>13</td>
<td>P</td>
<td>Ane</td>
<td>7,45</td>
</tr>
<tr>
<td></td>
<td>Gochnatia polymorpha (Less.) Cabrera</td>
<td>3</td>
<td>P</td>
<td>Ane</td>
<td>2,87</td>
</tr>
<tr>
<td></td>
<td>Vernonia polyanthes (Spreng.) Less.</td>
<td>8</td>
<td>P</td>
<td>Ane</td>
<td>5,13</td>
</tr>
<tr>
<td>Erythroxylaceae</td>
<td>Erythroxylum decidualum A. St.-Hil.</td>
<td>2</td>
<td>St</td>
<td>Zoo</td>
<td>1,30</td>
</tr>
<tr>
<td>Euphorbiaceae</td>
<td>Croton urucurana Baill.</td>
<td>1</td>
<td>P</td>
<td>Auto</td>
<td>0,52</td>
</tr>
<tr>
<td></td>
<td>Alchornea glandulosa Poepp.</td>
<td>2</td>
<td>P</td>
<td>Zoo</td>
<td>0,88</td>
</tr>
<tr>
<td>Fabaceae</td>
<td>Inga vera Willd.</td>
<td>4</td>
<td>Si</td>
<td>Zoo</td>
<td>2,23</td>
</tr>
<tr>
<td>Lauraceae</td>
<td>Ocotea odorifera (Vell.) Rohwer.</td>
<td>5</td>
<td>St</td>
<td>Zoo</td>
<td>2,36</td>
</tr>
<tr>
<td>Meliaceae</td>
<td>Trichilia catigua A. Juss.</td>
<td>3</td>
<td>Si</td>
<td>Zoo</td>
<td>1,33</td>
</tr>
<tr>
<td>Moraceae</td>
<td>Maclura tinctoria (L.) D. Don ex Steud.</td>
<td>1</td>
<td>Si</td>
<td>Zoo</td>
<td>0,53</td>
</tr>
<tr>
<td>Myrtaceae</td>
<td>Psidium cattleianum Sabine</td>
<td>1</td>
<td>P</td>
<td>Zoo</td>
<td>0,53</td>
</tr>
<tr>
<td></td>
<td>Psidium firmum O. Berg</td>
<td>6</td>
<td>P</td>
<td>Zoo</td>
<td>3,91</td>
</tr>
<tr>
<td></td>
<td>Psidium guajava L.</td>
<td>28</td>
<td>P</td>
<td>Zoo</td>
<td>20,80</td>
</tr>
<tr>
<td></td>
<td>Psidium rufum DC.</td>
<td>3</td>
<td>Si</td>
<td>Zoo</td>
<td>1,65</td>
</tr>
<tr>
<td>Piperaceae</td>
<td>Piper aduncum L.</td>
<td>14</td>
<td>P</td>
<td>Zoo</td>
<td>7,85</td>
</tr>
<tr>
<td>Primulaceae</td>
<td>Raphanea ferruginea (Ruiz e Pav.) Mez</td>
<td>4</td>
<td>Si</td>
<td>Zoo</td>
<td>2,18</td>
</tr>
<tr>
<td></td>
<td>Raphanea umbellata (Mart.) Mez</td>
<td>3</td>
<td>Si</td>
<td>Zoo</td>
<td>2,03</td>
</tr>
<tr>
<td>Rubiaceae</td>
<td>Guettarda uruguensis Cham. e Schldtl.</td>
<td>14</td>
<td>Si</td>
<td>Zoo</td>
<td>8,53</td>
</tr>
<tr>
<td>Sapindaceae</td>
<td>Allophylus edulis A. St.-Hil., Camb. e A. Juss.</td>
<td>2</td>
<td>Si</td>
<td>Zoo</td>
<td>0,96</td>
</tr>
</tbody>
</table>

NI = number of sampled individuals; CS = successional class: P = pioneer, Si = early secondary, St = late secondary; SD = dispersion syndrome: Zoo = zoochory, Ane = anemochory, Auto = autochory; IV = importance value.

The Shannon-Wiener diversity index (H’) demonstrates that the study area has low diversity of species. However, the result was higher than that found by Bastos (2010), in areas at seven (H’ = 2.48) and nine years (H = 1.60) of restoration. It is important to aim for a high diversity of species in restored...

![Figure 2](https://example.com/figure2.png)

**Figure 2** – Mean number of individuals and number of regenerating species, per stratum in the forest restoration process, Barroso, MG. Values followed by the same letter within each group did not differ significantly by Tukey test (0.05 > p ≥ 0.01).
ecosystems, especially in the case of rain forests, which are inherently rich in species. However, many studies have shown that plantings with few species or even monospecific can accelerate the natural regeneration of vegetation, acting as “catalytic” of the forest restoration process. In other words, richness or initial species diversity do not matter if they do not form a set fitted to local conditions and are not able to survive and to quickly cover the soil, thereby, facilitating the continuity of the succession (DURIGAN et al., 2010).

The Pielou’s equability value found in this study highlights that despite the low value of diversity found, the forest under restoration is heterogeneous, with low ecological dominance. This result is similar to that reported by Bastos (2010), in areas under restoration process for 7 (J’ = 0.860) and 9 years (J’ = 0.890) and exceed the results obtained by Garcia et al. (2011) and Higuchi et al. (2006), with indexes ranging from 0.700 to 0.740 in semi-deciduous seasonal forest fragments.

The high importance value (IV) presented by the species *Psidium guajava*, *Lithraea molleoides*, *Guettarda uruguensis*, *Schinus terebinthifolia*, *Piper aduncum* and *Baccharis dracunculifolia* as well as the families Myrtaceae, Anacardiaceae and Asteraceae corroborate many natural regeneration surveys in semi-deciduous seasonal forests (SOARES, 2009; ARANTES et al., 2012; BILA, 2012), where such species and families stood out in the regenerating stratum, as well. This fact may be attributed to the following characteristics: pioneering, wide geographical distribution, adaptation to anthropized environments, with poor fertile soils and high light intensity, intrinsic of these families and their species.

5. CONCLUSION

Natural regeneration has proved to be a good indicator of the status of regeneration of an area as it showed the low diversity of regenerating shrub-tree species, mostly belonging to the early stages of regeneration. The results suggest the existence of barriers that disrupt the development of tree species, suggesting that corrective management actions are taken to accelerate the successional process.

6. ACKNOWLEDMENTS

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7. REFERENCES


Natural regeneration stratum as an indicator of...


