



Study of the Cu, Mn, Pb and Zn dynamics in soil, plants and bee pollen from the region of Teresina (PI), Brazil

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ABSTRACT

The purpose of this study is to characterize native bee plants regarding their capacity to extract and accumulate trace elements from the soil and its consequences to the sanity of the produced pollen. The trace elements Cu, Mn, Pb and Zn were analyzed in soil, plants and bee pollen from Teresina region (PI), Brazil, by flame atomic absorption spectrophotometer. Considering the studied plant species, Cu and Pb metals presented in the highest levels in the roots of *B. platypetala* with 47.35 and 32.71 $\mu\text{g}\cdot\text{mL}^{-1}$ and *H. suaveolens* with 39.69 and 17.06 $\mu\text{g}\cdot\text{mL}^{-1}$, respectively, while in the aerial parts Mn and Zn metals presented the highest levels in *S. verticillata* with 199.18 and 85.73 $\mu\text{g}\cdot\text{mL}^{-1}$. In the pollen, the levels of Cu, Mn, Pb and Zn vary from 5.44 to 11.75 $\mu\text{g}\cdot\text{mL}^{-1}$; 34.31 to 85.75 $\mu\text{g}\cdot\text{mL}^{-1}$; 13.98 to 18.19 $\mu\text{g}\cdot\text{mL}^{-1}$ and 50.19 to 90.35 $\mu\text{g}\cdot\text{mL}^{-1}$, respectively. These results indicate that in the apicultural pasture the translocation (from soil to pollen) of Mn and Zn was more effective than in case of Cu and Pb, therefore, the bee pollen can be used as food supplement without causing risks to human health.

Key words: bee pollen, bee plant, trace elements, translocation capacity.

INTRODUCTION

The bee pollen is a natural material of high complexity, whose composition significantly varies according to environmental variables such as: the region where it is produced, predominant flora, bees' preference, geochemical characteristics of the soil and the period of the year (Alcoforado Filho and Gonçalves 2000). From a macroscopic point of view, it is composed of botanical pollen grains that bees collect from the anthers of male or hermaphrodite flowers and moisten with

their salivary secretion, forming pellets that are transported up to the hive (Barth 1989). When it is collected at the hive's entrance and dehydrated at most 42°C, until reaching 4% of maximum moisture level, the product is considered "dehydrated bee pollen". This is the most convenient presentation form for commercial purposes (Brasil 2001).

Despite being considered a "clean" substance, if the environment where it is produced is polluted, the pollutants can be present in its composition, sometimes in significantly high concentrations (Fredes and Montenegro 2006, Kump et al. 1996, Rashed and Soltan 2004). Two of the main pollutant sources that

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may be absorbed by the pollen are: 1) the atmospheric air, through particulate material in suspension (Ward 2000); and 2) the soil, through plants absorption and translocation (Turgut et al. 2004). In both cases the pollutants can be of organic nature as well as mineral nature. In the second case, there is a prevalence of minerals, among them heavy metals deserve special attention (Chen et al. 2004, Meers et al. 2005). This is because, in addition to the intrinsic toxicity of these metals, some bee plants also exhibit characteristics of phytoaccumulation of heavy metals, contributing to their biomagnification (Ajasa et al. 2004, Cintra et al. 2005, Ghosh and Singh 2005).

Except in cases of natural catastrophes (volcanisms, flood, etc), high levels of heavy metals in the soil are of predominately anthropic origin, as they usually occur at very low concentrations in a terrestrial environment, typically at trace and ultra-trace levels (Moura et al. 2006, Ward 2000). This means that the emergence of high levels of heavy metals in the pollen should be mainly credited to other factors and not to their expressive abundance in this place. Two factors that may be determinant in situations like that are: 1) these metals are of intrinsically cumulative character in living tissues; 2) there are some polliniferous plants that are also good accumulators of these metals (Azeredo et al. 2003, Barth 1989).

Actually, some plant species are known and well characterized regarding their capacity to accumulate high levels of heavy metals in their biomass. They are classified as hyperaccumulators (Meers et al. 2005, Pollard and Baker 1997, Visoottiviset et al. 2002, Wei et al. 2006). According to specialized literature (Yoon et al. 2006) there are reports of about 400 plant species that present this phenomenon. From this estimated amount, 26 plant species are selective for cobalt, 24 are selected for copper, 8 are selected for manganese, 145 are selected for nickel, 5 are selected for lead and 4 plant species are selected for zinc.

Although the hyperaccumulation phenomenon is not a nutritional need for the plant, but only a

tolerance that is superior than the average of other species, this also represents a peculiar characteristic of each species or group of species (Oven et al. 2002). Thus, in regions where apiculture is practiced for commercial purposes, the identification of bee plants with this characteristic is an important item to be evaluated, in order to assure that the product fully meets technical and sanitary specifications imposed by the regulatory agencies and the demanding consumer market of modern times.

This work aims to characterize the plants and bee pollen produced in Teresina, regarding the contents of Cu, Mn, Pb and Zn, as well as their consequences to the quality of the product.

MATERIALS AND METHODS

REAGENTS AND SOLUTION

All reagents employed in this study were of analytical grade and they were used without previous purification. In order to prepare the solutions, it was used ultra pure water (18 M Ω cm) obtained through the Milli-Q purification system (Millipore, RO15). The analytical curves to determine the studied heavy metals (Cu, Mn, Pb and Zn) were constructed by adequate dilutions of stock solutions of the respective high purity standards, containing 1,000 $\mu\text{g}\cdot\text{mL}^{-1}$ of each metal.

APPARATUS

For the plants samples pulverization it was used a mill (Fritsch, Pulverisette 14) that was equipped with a sieve of 0.50 mm mesh embedded in a stainless steel hollow.

The mineralization of the samples was carried out using a digestion system (Digester, Quimis Model) equipped with an aluminum block of 42 tubes capacity (0.1 L) and a temperature controller.

In order to obtain the analytical determinations, it was used an atomic absorption spectrophotometer (Varian, SpectrAA 220 FS). A hollow cathode lamp of each metal was used as a radiation source and it

TABLE I
Wavelength, experimental conditions and parameters of the calibration curve to the analyzed metals.

Metals	Wavelength (nm)	Slit width (nm)	Current lamp (mA)	Gas flow (Air/Acetilene) (L.min ⁻¹)	Linear range (µg.mL ⁻¹)	Regression equations	Correlation coefficient
Cu	324.8	0.5	4.0	11.00/2.0	0.1 – 1.3	A = 0.0622 x C – 0.0006	0.9992
Mn	279.5	0.2	5.0	13.50/2.0	0.1 – 2.2	A = 0.0575 x C – 0.0048	0.9966
Pb	217.0	1.0	5.0	13.50/2.0	0.1 – 1.3	A = 0.0212 x C + 0.0010	0.9932
Zn	213.9	1.0	5.0	13.50/2.0	0.1 – 1.9	A = 0.3306 x C + 0.0076	0.9999

was measured under optimum operating conditions with an air-acetylene flame. The instrumental conditions and parameters were selected according to the recommendations mentioned in the equipment's instructions manual (Analytical Methods 1989), as presented in Table I.

PREPARATION OF SAMPLES AND ANALYTICAL DETERMINATION

The collecting of samples was carried out in the Apicultural Sector of the Center for Agricultural Sciences of the Universidade Federal do Piauí, settled in an area covered by native plant species (localization: 05° 05' 21" S, 42° 48' 07" W, altitude: 72 m).

The plant species utilized in this study were: *Bauhinia platyptala*, *Hyptis suaveolens*, *Mimosa caesalpiniaefolia* and *Spermacoce verticillata*. Samples of roots and aerial parts of each plant were collected during the blossom period (March and June/2006). They were washed in current and distilled water and left to dry in a stove for 48 hours, at a temperature of 75° to 80°C. In the next step, samples were pulverized and dried again at the same temperature, until they reached a constant weight. Portions of approximately 1.0g of samples were digested in an acidic mixture of nitric and perchloric acid (Ghosh and Singh 2005). The obtained solutions were diluted to the volume of about 50 mL.

The pollen was monthly collected, from *Apis mellifera* hives of the same area, between January and December, 2006. After the samples have been dehydrated and homogenized, they were stored in plastic bags at a temperature of approximately

–18°C. The samples of 1 mm grain size dried previously to the constant mass at 50°C for 12 hours were mineralized according to the already described procedure for plants and analyzed by flame atomic absorption spectrophotometry.

RESULTS AND DISCUSSION

VALIDATION

The limits of detection (LOD) and quantification (LOQ) presented in Table II were determined according to Ribani et al. (2004), by using Equations (1) and (2).

$$\text{LOD} = 3.33 \times (s/S) \quad (1)$$

$$\text{LOQ} = 10 \times (s/S) \quad (2)$$

where “s” is the standard deviation of the measures referring to the blank (n = 20) and “S” is the inclination of the analytical curve.

The recovery test of each metal was carried out based on the improvement of the samples solutions, by adding adequate aliquots of metals standard solutions, in order to obtain final concentrations of 0.4 and 0.8 µg.mL⁻¹ in the work solutions. The results were calculated according to Equation (3).

$$\%R = (c_i - c_0/C_i) \times 100 \quad (3)$$

where c_i is the final concentration of the improved work solutions, c_0 is the concentration of the work solution without improvement and C_i is the final concentration of the standard solution. The results are presented in Table II.

TABLE II
LOD, LOQ values and recovery test used in the analytical determination of Cu, Mn, Pb and Zn.

Metals	LOD ($\mu\text{g}\cdot\text{mg}^{-1}$)	LOQ ($\mu\text{g}\cdot\text{mg}^{-1}$)	Recovery		
			C_i ($\mu\text{g}\cdot\text{mL}^{-1}$)	c_i ($\mu\text{g}\cdot\text{mL}^{-1}$)	R (%)
Cu	0.071	0.24	-	0.162 (2.45%)*	-
			0.40	0.531 (1.14%)	92
			0.80	0.903 (3.10%)	93
Mn	0.054	0.17	-	1.93 (0.572%)	-
			0.40	2.34 (1.19%)	102
			0.80	2.84 (0.701%)	113
Pb	0.072	0.23	-	0.422 (1.65%)	-
			0.40	0.784 (3.72%)	91
			0.80	1.19 (3.01%)	96
Zn	0.031	0.095	-	1.64 (1.70%)	-
			0.40	1.99 (4.41%)	87
			0.80	2.40 (3.45%)	95

* Average of the concentrations (% coefficient of variation, $n = 3$).

The observed recovery results situated between 87% (Pb) and 113% (Mn), with coefficients of variation lower than 10%. These data indicate that, despite the material complexity, the interferences due to the effects of the matrix are insignificant (Ribani et al. 2004) and therefore, the employed methodology is adequate to the referred product.

TRANSLOCATION OF METALS

The collected levels of metals in the soil and in the two different bee plant species (roots and aerial parts) are shown in Table III.

The levels of metals encountered in the soil are relatively low in relation to the reference values usually accepted as risk thresholds for human health and the environment (Falco et al. 2005). Although there is no consensus about this question, some specialists, institutions and governments use the limits ($X \pm 2s$) and $100(X \pm 2s)$ as criterion to characterize the existence of risk in soil that was not polluted by human activity-except for agricultural purposes-where X is the average and s is the standard

deviation of the levels presented in the referred area (Gil et al. 2004). According to this criterion, only Mn (with $X = 18.60$ and $s = 0.751 \mu\text{g}\cdot\text{mg}^{-1}$) represents some approximating threat to this critical range or risk threshold for non-polluted soils. However, it has to be taken into account that these parameters should only be considered as local or regional guidelines, and they are also limited in time. The time limitation, even in protected areas, is essentially related to the dynamics of metals by aerial or aquatic ways that allow its "spontaneous" occurrence in these areas, even when the generating source is miles away (Micó et al. 2007, Reimann et al. 2001).

It is important to observe that low levels of metals in the soil do not necessarily mean low levels of metals in the plant. Different species can present not only different tolerances to high levels, but also different capacities of extracting and directing the accumulation of metals from the soil to the different parts of the vegetable (Rodella 2005). The metals capacity of translocation from the soil to the roots (CTR), from the soil to the aerial parts (CTA) and

TABLE III

Levels of metals ($\mu\text{g.mg}^{-1}$) and the coefficient of variation (%) found in the soil and in the studied plants (except Cu in *S. verticillata*).

Metals	Soil ($\mu\text{g.mL}^{-1}$)	<i>B. platypetala</i>		<i>H. suaveolens</i>		<i>M. caesalpiniaefolia</i>		<i>S. verticillata</i>	
		Roots ($\mu\text{g.mL}^{-1}$)	Aerial parts ($\mu\text{g.mL}^{-1}$)	Roots ($\mu\text{g.mL}^{-1}$)	Aerial parts ($\mu\text{g.mL}^{-1}$)	Roots ($\mu\text{g.mL}^{-1}$)	Aerial parts ($\mu\text{g.mL}^{-1}$)	Roots ($\mu\text{g.mL}^{-1}$)	Aerial parts ($\mu\text{g.mL}^{-1}$)
Cu	1.72 (2.63%)	47.35 (16.19%)	8.09 (2.95%)	39.69 (9.72%)	9.75 (12.91%)	8.88 (11.61%)	4.87 (17.10%)	< LOD	< LOD
Mn	18.60 (4.06%)	18.29 (13.21)	66.58 (1.50)	59.60 (5.55%)	89.15 (4.30%)	17.32 (7.01%)	48.18 (3.49%)	54.67 (2.69%)	199.18 (2.27%)
Pb	0.58 (5.87%)	32.71 (12.43)	11.88 (0.24%)	17.06 (3.43%)	13.47 (5.26%)	14.74 (1.84%)	11.68 (4.70%)	17.10 (12.19%)	16.29 (2.03%)
Zn	1.84 (4.65%)	30.97 (5.65%)	35.19 (3.03%)	26.32 (5.08%)	40.37 (0.13%)	11.99 (4.86%)	31.06 (9.82%)	19.62 (1.11%)	85.73 (3.43%)

from the roots to the aerial parts, so called intrinsic translocation (CTI), are defined in the equations (4-6), where MCR, MCA and MCS are the metals concentrations in the roots, in the aerial parts and in the soil, respectively. Sometimes these parameters determine the effects produced by the metal in the plant, e.g. the development of toxicity (Nobre et al. 2004, Robinson et al. 1999).

$$\text{CTR} = \text{MCR}/\text{MCS} \quad (4)$$

$$\text{CTA} = \text{MCA}/\text{MCS} \quad (5)$$

$$\text{CTI} = \text{MCA}/\text{MCR} \quad (6)$$

Besides, according to Wei et al. (2006), these parameters are also important to distinguish the non-accumulating or moderately tolerant species, from the hyper-accumulating or highly tolerant species (Yoon et al. 2006). Generally, the enrichment factor (EF), that may assume any of the above expressions, since it has been specified, provides good information about the plant's potential to hyper-accumulation and through which process this state is achieved. Particularly, high values of CTR combined with low values of the CTI in a certain plant means an advantageous situation in case of bee plants and oleraceas, in general (except to edible tubers and rhizomes), because their development in contaminated soil normally occur, and they can accumulate high levels of metal (in roots), without compromising the useful aerial parts (leaves, flowers,

fruits and seeds). In places suspected of being contaminated by heavy metals, it is recommended to cultivate plants with this characteristic, because it enables the production of food in a potentially useless or dangerous type of soil. At the same time, this type of plants can be used as phytoremediators of the cultivated area, by removing it and burning it, in order to obtain the metal's bio-ore (Reimann et al. 2001, Anderson et al. 1999).

In regions without an intense industrial activity (as in the focused case), it is prudent to admit that the absorption of metals in the air by leaves and flowers is negligible. Otherwise, it will be necessary to monitor the air of the region in order to identify the metal's intake route in the plant (soil or air) and then decide whether the cultivation of species with high CTR and low CTI is viable or not.

The found values of the metals capacity of translocation, from the soil to the roots (CTR), from the soil to the aerial parts (CTA) and from the roots to the aerial parts (CTI) in *B. platypetala*, *H. suaveolens*, *M. caesalpiniaefolia* and *S. verticillata* species are demonstrated in Table IV.

Based on Table IV data, it is possible to critically evaluate the real significance of high and low levels of metals in the soil, considering the tested plant species. It is observed that Pb and Zn, despite being presented at low levels in the soil (0.58 and 1.84 $\mu\text{g.mg}^{-1}$), exhibit the highest values

TABLE IV
Values of CTR, CTA and CTI for the analyzed metals.

Species	Translocation	Metals			
		Cu	Mn	Pb	Zn
<i>B. platypetala</i>	CTR	27.53	0.98	56.40	16.83
	CTA	4.70	3.58	20.48	19.12
	CTI	0.17	3.64	0.36	1.14
<i>H. suaveolens</i>	CTR	23.07	3.20	29.41	14.30
	CTA	5.67	4.80	23.22	21.94
	CTI	0.24	1.50	0.79	1.53
<i>M. caesalpiniaefolia</i>	CTR	5.16	0.93	25.41	6.52
	CTA	2.83	2.59	20.14	16.88
	CTI	0.55	2.78	0.79	2.59
<i>S. verticillata</i>	CTR	-	2.94	29.48	10.66
	CTA	-	10.71	28.09	46.59
	CTI	-	3.64	0.95	4.37

of CTR and CTA in both species, while Mn, that has the highest level in the soil ($18.60 \mu\text{g}\cdot\text{mg}^{-1}$), exhibits very low values of CTR and CTA in comparison to Pb and Zn. Besides, all the values of CTI are very low, indicating that, even in cases where there is a significant accumulation of metals in the plants, this occurs preferentially in the roots and, therefore, these two species have an insignificant importance as tributaries of those metals in the bee pollen produced in this region.

SEASONAL VARIATIONS OF Cu, Mn, Pb AND Zn LEVELS IN THE BEE POLLEN FROM TERESINA REGION

The average levels of Cu, Mn, Pb and Zn that were found in the analyzed samples of bee pollen and their respective coefficients of variation are shown in Table V.

The levels of Cu, Mn, Pb and Zn indicated in Table V are very informative regarding the seasonal dynamics of occurrence of these metals in the bee pollen from the studied region. Mn and Zn metals are presented in higher levels while Cu and Pb are presented in lower levels. The causes of this separation may be related to the accumulating mechanism of Mn and Zn that, according to Singh

and Sinha (2005), Singh and Agrawal (2007) and Ali et al. (2006), involves complexation of the metal with groups ($-\text{NH}_2$) and ($-\text{SH}$) of amino acid and non-peptide thiol presented in tissue plants.

Cu, Mn and Zn, respecting the tolerable limits, are essential micronutrients for animals and plants, and they take part in important vital processes for both types of living beings (Franco 2002). Because it does not have a known (beneficial) biological function, Pb is only considered a highly toxic trace-element that cause many disturbances to plants and animals health, including man and the environment (Chen et al. 2004).

Under well defined experimental conditions, Krajncic and Nemec (2003) have demonstrated that Cu, Mn and Zn together participate in catalytic processes of some superior plant species, whose results are not only a sensible increase in the production of flowers, attributed to Mn, but also an increase in the photosynthesis rate and sucrose synthesis, both due to the joint participation of Cu and Zn. This somewhat would justify the fact that Mn and Zn have very similar levels in most of the samples (approximately 75%). From that point, there is a gradient of concentration determined by the plant's nutritional needs.

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RESUMO

O objetivo deste estudo é caracterizar plantas apícolas nativas, quanto a sua capacidade de extrair e acumular elementos-traço do solo e suas conseqüências na sanidade do pólen produzido. Os elementos-traço Cu, Mn, Pb e Zn foram analisados em solo, planta e pólen apícolas em Teresina (PI), Brasil, por espectrofotometria de absorção atômica com atomização em chama. Considerando as espécies de plantas estudadas, os metais Cu e Pb apresentaram nas raízes maiores teores de *B. platypetala* com 47,35 e 32,71 $\mu\text{g}\cdot\text{mL}^{-1}$ e *H. suaveolens* com 39,69 e 17,06 $\mu\text{g}\cdot\text{mL}^{-1}$, respectivamente, enquanto na parte aérea os metais Mn e Zn apresentaram os maiores teores, em *S. verticillata* com 199,18 e 85,73 $\mu\text{g}\cdot\text{mL}^{-1}$. No pólen os teores de Cu, Mn, Pb e Zn varia de 5,44 a 11,75 $\mu\text{g}\cdot\text{mL}^{-1}$; 34,31 a 85,75 $\mu\text{g}\cdot\text{mL}^{-1}$; 13,98 a 18,19 $\mu\text{g}\cdot\text{mL}^{-1}$ e 50,19 a 90,35 $\mu\text{g}\cdot\text{mL}^{-1}$, respectivamente. Esses resultados indicam que, no pasto apícola, a translocação (do solo ao pólen apícola) de Mn e Zn foi mais eficiente do que Cu e Pb, portanto, o pólen apícola poderá ser utilizado como complemento alimentar sem causar riscos à saúde humana.

Palavras-chave: pólen apícola, plantas apícolas, elementos-traço, capacidade de translocação.

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