

Density estimates and conservation of *Leopardus pardalis* southernmost population of the Atlantic Forest

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ABSTRACT. Using camera traps and capture/recapture analyses we recorded the presence and abundance of cat species at Turvo State Park, in southern Brazil. Ocelot [*Leopardus pardalis* (Linnaeus, 1758)] population density was estimated for two areas of the park, with differing management profiles. Density estimates varied from 0.14 to 0.26 indiv. km². Another five cat species were recorded at very low frequencies, precluding more accurate analyses. We estimate 24 to 45 ocelots occur in the reserve, which is probably too small for long-term maintenance of the population, if isolated. However, if habitat integrity and connectivity between the Park and the Green Corridor of Misiones is maintained, an estimated ocelot population of 1,680 individuals should have long-term viability.

KEYWORDS. *Panthera*, *Puma*, sympatric cats, camera trap, capture/recapture.

RESUMO. Estimativas de densidade e aspectos de conservação da população *Leopardus pardalis* mais austral da Mata Atlântica. Utilizando armadilhas fotográficas e análises de captura/marcação/recaptura, registramos a presença e abundância de felinos no Parque Estadual do Turvo, sul do Brasil. A densidade de jaguatiricas [*Leopardus pardalis* (Linnaeus, 1758)] foi estimada para duas áreas dentro do Parque, que apresentam diferentes formas de uso. A densidade estimada variou de 0,14 a 0,26 indivíduos por km². Outras cinco espécies de felinos foram registradas em frequências muito baixas, impedindo análises mais acuradas. Estimamos que 24 a 45 jaguatiricas devam utilizar a Unidade de Conservação, o que provavelmente representa uma população muito pequena para se manter ao longo prazo, caso isolada. Contudo, se a integridade e conectividade de habitat entre o Parque e o Corredor Verde de Misiones forem mantidas, uma população de jaguatiricas estimada em 1.680 indivíduos deve ser viável a longo prazo.

PALAVRAS-CHAVE. *Panthera*, *Puma*, felinos simpátricos, armadilha fotográfica, captura/recaptura.

The ocelot [*Leopardus pardalis* (Linnaeus, 1758)] is a medium sized cat (7 - 16 kg) with a broad geographic distribution in the Americas, ranging from the southwestern United States to northern Argentina and southern Brazil (EISENBERG & REDFORD, 1999). Although it is common over large areas, such as within the Amazon basin, this felid is regionally threatened and included in some local red lists. This is notably the case in the highly threatened biodiversity hot spot of the Atlantic Forest, where the ocelot is classified as Vulnerable on the red list of threatened species of Rio Grande do Sul State (FZB, 2014), among several other species. Ocelot use many habitat types, including seasonally flooded forests, cerrado and Atlantic Forest (WILSON & MITTERMEIER, 2009). It is the best known of the smaller Neotropical cats, with studies of its home range (LUDLOW & SUNQUIST, 1987; EMMONS, 1988; CRAWSHAW & QUIGLEY, 1989), diet (e.g., EMMONS, 1987; KONECNY, 1989) activity patterns, (e.g., LUDLOW & SUNQUIST, 1987; DI BITETTI *et al.*, 2006) and habitat use (e.g., LUDLOW & SUNQUIST, 1987; DI BITETTI *et al.*, 2006).

Camera traps are becoming increasingly important for wildlife studies (WEMMER *et al.*, 1996). They are especially useful for rare or difficult to observe species (TOMAS & MIRANDA, 2003). Additionally, for many animals, photographs allow individual identification to be made via unique pelage markings. Accordingly, capture-recapture statistical methods can be applied to such animals without

their actual physical capture (KARANTH *et al.*, 2003). Using such methods, camera traps have been used to estimate population densities of large cats, such as tiger *Panthera tigris* (Linnaeus, 1758) (KARANTH, 1995; KARANTH & NICHOLS, 1998) and jaguar *Panthera onca* (Linnaeus, 1758) (WALLACE *et al.*, 2003; SILVER *et al.*, 2004; MAFFEI *et al.*, 2004; SOISALO & CAVALCANTI, 2006; PAVIOLO *et al.*, 2008). However, in the Neotropics, camera trap studies have been applied to only a few species of the smaller cats such *L. pardalis* (TROLLE & KÉRY, 2003, 2005; MAFFEI *et al.*, 2005; DI BITETTI *et al.*, 2006; DILLON & KELLY, 2007; DI BITETTI *et al.*, 2008; MAFFEI & NOSS, 2008; GOULART *et al.*, 2009; FUSCO-COSTA *et al.*, 2010), *Leopardus geoffroyi* (d'Orbigny & Gervais, 1844) (CUELLAR *et al.*, 2006) and *Leopardus guttulus* (Schreber, 1775) (TORTATO & OLIVEIRA, 2005).

Population parameter estimates are fundamentally important for wildlife conservation and management, especially for those carnivores that are sensitive to disturbance. Here we use camera traps to evaluate the relative abundance and when possible, estimate density of felids at Turvo State Park, southern Brazil.

MATERIAL AND METHODS

Turvo State Park is located in southernmost Brazil on the border of Argentina (27°00'S – 53°40'W to 27°20'S – 54°10'W). The park comprises 17,500 ha of rainforest

of the Upper Uruguay River basin, and is part of the Atlantic Forest domain. The park lies within the Misiones Green Corridor, a well conserved area that extends from Iguazu National Park (in the State of Paraná, in Brazil) to the Misiones region of Argentina. This area actually represents the southern range limit for several species, such as *Panthera onca* and *Tapirus terrestris* (OLIVEIRA, 1994; EISENBERG & REDFORD, 1999), and a variety of other vertebrates (FONTANA *et al.*, 2003). Turvo State Park is surrounded by an agricultural landscape dominated by soybean, wheat and dairy cattle. The original forest now consists of small fragments of secondary or altered primary forest, often associated with *Eucalyptus* spp. cultivation. The only connection of the park with forested areas is through the adjacent region in the Misiones province of Argentina, crossing the Uruguay River (Fig. 1). There are two internal roads in the park (1) the “Estrada do Salto” (Salto Road) is 15 km long and is open to the public from 8 AM to 5 PM and (2) “Estrada do Porto” (Porto Road) is 8 km long and its use is restricted to researchers. Camera trapping was carried out along these two roads.

In 2005, eight camera stations were set on Salto road operating for four consecutive nights each month, from February to September. This sampling effort resulted in 312 trap nights that were considered eight (monthly) capture events. Four camera stations were set on Porto road. On Porto road, cameras were used until the end of photographic film, which varied from four to 14 days, monthly, from January to August. This sampling effort resulted in 412

trap nights, and was considered eight (monthly) capture events. The camera stations were separated from each other by 2 km (mean distance) in each road, and were composed by pairs of cameras placed facing each other to obtain the record of both sides all animals photographed.

In 2006 we set eight pairs of camera traps on each road, with four along the main road and four along trails perpendicular to the roads. Cameras were placed at intervals of 300 meters and were in operation for 60 – 70 consecutive days. Salto road was monitored from June to August resulting in a sample effort of 456 trap nights, consisting in 11 capture events of six days. Porto road was monitored from March to June resulting in a sample effort of 442 trap nights of 12 capture events of six days.

Animals were identified by their unique pelage patterns of spots and rosettes. We analyzed capture events with the program *Capture* (REXSTAD & BURNHAM, 1991). This program compares different models to determine the most parsimonious, and examines differences in capturability among individuals (sex, age, etc.) over time, and combinations thereof (REXSTAD & BURNHAM, 1991).

The effective sampling area was estimated by the sum of buffers added around each trap station resulting in a polygon formed by the union of the buffer zones. The size of the buffer was defined as the mean of the maximum distance moved (MMDM) by the animals of each species. This value was calculated considering the mean of the maximum distance moved by animals that were recorded by different cameras. For comparative purposes, we also show estimates using $\frac{1}{2}$ MMDM buffer, previously used by several authors (TROLLE & KERY, 2005; DILLON & KELLY, 2007; GOULART *et al.*, 2009).

RESULTS

We recorded six species of felids in the Park: southern tiger cat *Leopardus guttulus*, margay *Leopardus wiedii*, ocelot *L. pardalis*, jaguarundi *Puma yagouaroundi*, puma *Puma concolor* and jaguar *P. onca*. Only the ocelot was common, comprising 87.7% (n= 211) of all cat records. Jaguarundi showed intermediate abundance, with 10.9% of the felid records from cameras. For puma, we obtained only five records of three different animals, while for jaguar only one individual was recorded four times. Margay and southern tiger cat were each recorded only once (Tab. I).

Tab. I. Photographic records of felids in the forests of Turvo State Park, Rio Grande do Sul state, Brazil.

Species	Records	
	N	%
Ocelot - <i>Leopardus pardalis</i>	211	84.7
Margay - <i>Leopardus wiedii</i>	1	0.4
Southern tiger cat - <i>Leopardus guttulus</i>	1	0.4
Puma - <i>Puma concolor</i>	5	2.0
Jaguarundi - <i>Puma yagouaroundi</i>	27	10.9
Jaguar - <i>Panthera onca</i>	4	1.6

There were no meaningful differences in density estimates between 2005 and 2006 for either Porto and Salto

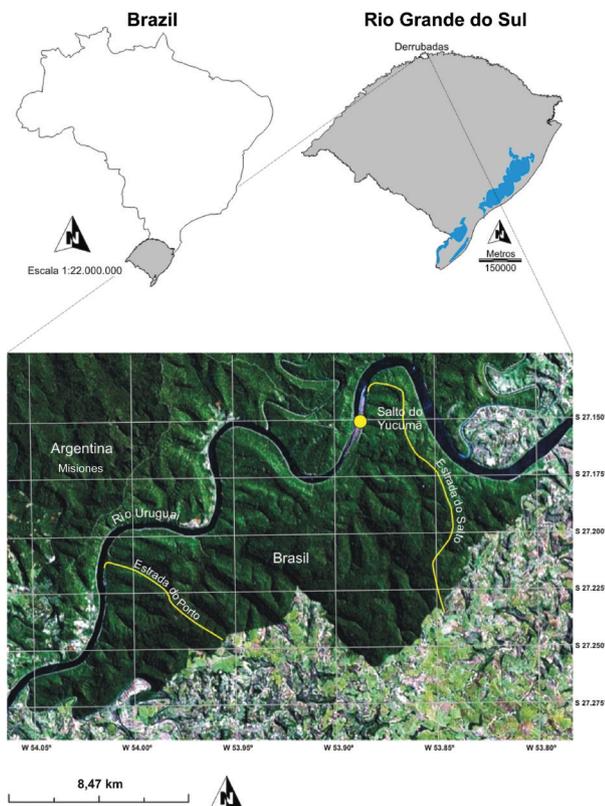


Fig. 1. Location of Turvo State Park at southern Brazil, and its connection with the region of Misiones, Argentina.

areas (Tab. II). In 2005, we obtained 98 ocelot records, 30 on the Porto road and 68 on the Salto road. We identified nine individuals, five males and four females, at Porto and 12 individuals, six males and six females, at Salto. In 2006, we obtained 113 records of ocelots, 31 at Porto and 82 at Salto. We identified five individuals at each road site: three males and two females at Porto, and three males, one female and one unknown sex at Salto. During the two years of study a total of 26 individuals were recorded using the sampled area.

Via photographs from different camera stations we recorded distance moved for 15 animals. The mean

maximum distance moved (MMDM) was 2.7 km. There were no statistical difference in either movement of males and females ($t = 0.18$, $df = 7$, $p = 0.8605$), or in the distance moved by individuals from the different road samples ($t = 1.41$, $df = 9$, $p = 0.1852$).

Based on MMDM, the effectively sampled area in 2005 was 58.12 km² on Porto road, and 92.09 km² on Salto road, and 29.99 km² on Porto road and 35.25 km² on Salto road in 2006. Density estimates varied between 0.14–0.26 individuals per km², with lower estimated densities for the Salto road site (Tab. II). Density estimates using ½ MMDM were three times greater than using the full MMDM.

Tab. II. Sampling areas, population size estimates (selected model) with 95% confidence interval (CI), effectively sampled area and density estimates (individuals / km²) at Turvo State Park, Rio Grande do Sul state, Brazil.

Sample	Estimated Population (Model)	95% CI	Sampled Area	Density
Porto 2005	15 (Mth)	11 – 41	½ MMDM = 22.54	0.66
			MMDM = 58.12	0.26
Salto 2005	13 (Mth)	13 – 22	½ MMDM = 39.78	0.33
			MMDM = 92.09	0.14
Porto 2006	6 (Mh)	5 – 17	½ MMDM = 9.17	0.65
			MMDM = 29.99	0.20
Salto 2006	5 (Mh)	5 – 14	½ MMDM = 11.92	0.42
			MMDM = 35.25	0.14

DISCUSSION

Although six species of cats were recorded at Turvo State Park, only ocelot was relatively abundant. This felid was recorded six times more than jaguarundi, the second most frequent cat in the park. The other smaller cats as well the larger species, jaguar and puma, seems occur in low abundance, recorded only sporadically.

The lack of differences between the years for the two areas is suggestive that camera arrangements did not affect results. The slightly higher density estimates for the Porto over the Salto site might be related to the latter area's public use of the roads, which could negatively impact ocelot. Although we do not know if prey availability is different between the two areas, all other environmental conditions appeared equal. In any event, this is highly indicative that abundance can change for varying reasons among sites within the same area/region.

Using 2005 data, we estimated ocelot densities of 0.14 to 0.26 individuals per km². These population density estimates (using the full MMDM) are the largest reported for the ocelot in the southern Atlantic Forest, and similar to that observed by FUSCO-COSTA *et al.* (2010) in southeastern Brazil. In Misiones, density estimates (also using full MMDM) were much smaller, with 0.03 individuals/km² (in altered areas) to 0.12 indiv./km² in preserved areas (DI BITETTI *et al.*, 2008). Another study in Atlantic Forest also found low densities, with 0.04 indiv./km² (GOULART *et al.*, 2009). In the Pantanal, estimates using the same method found 0.07 indiv./km² (TROLLE & KÉRY, 2005). However, our estimate seem perfectly plausible considering that we found unless 21 different animals using the Park during 2005. Another strong indication that Turvo State

Park has high ocelot densities is the notable capture rate of our study. We obtained 13.1 records/100-trap-nights⁻¹ in this study, a value considerably higher than the 2.5 – 8.1 records/100-trap-nights obtained by DI BITETTI *et al.* (2006), or 0.99 records/100-trap-nights (GOULART *et al.*, 2009) and 3.2 records/100-trap-nights (TROLLE & KÉRY, 2005). There is some discussion if capture rates reflect population density (CARBONE *et al.*, 2001; JENNELLE *et al.*, 2002), but we consider that comparisons of relative abundance using similar methods with the same species are more than justified for making inferences between areas. Another aspect that can be associated to density is movement pattern of individuals. Average maximum movement distances were smaller in this study (2.7 km) than the 3.96 km from Misiones (DI BITETTI *et al.*, 2006), which may suggest that ocelot have smaller home ranges at Turvo State Park and, hence, higher density.

Although has been used by several authors, effective sampling area estimates based on ½ MMDM model is probably very inaccurate. This model may substantially underestimate the area involved (TROLLE & KÉRY, 2005; MAFFEI & NOSS, 2008), and consequently, overestimate population density (SOISALO & CAVALCANTI, 2006). To avoid, or reduce, the problem of underestimating the sampled area, we adopted the full MMDM model, which gives more conservative density estimates. Also, full MMDM generated similar estimations of density observed for the species in the same habitat (DI BITETTI *et al.*, 2008). In our study, the use of ½ MMDM generates estimations three times bigger than estimated by MMDM, which seems unrealistic. (Tab. III).

The suggestion that ocelot density is positively correlated with rainfall and negatively correlated with

Tab. III. Habitat, method used to estimate effective sampled areas and density estimates (individuals / km²), using camera traps, in chronological order.

Habitat	Method for estimation of effective sampled area	Density	Source
Pantanal – flooded forest	½ MMDM	0.56	TROLLE & KERY, 2003
Pantanal – flooded forest	½ MMDM	0.11	TROLLE & KERY, 2005
Dry Forest	½ MMDM	0.24 – 0.66	MAFFEI <i>et al.</i> , 2005
Atlantic Forest	MMDM	0.08 – 0.13	DI BITETTI <i>et al.</i> , 2006
Scrub Forest	½ MMDM	0.30	HAINES <i>et al.</i> , 2006
Tropical rainforest	½ MMDM	0.26	DILLON & KELLY, 2007
Atlantic Forest	MMDM	0.05 – 0.12	DI BITETTI <i>et al.</i> , 2008
Atlantic Forest	½ MMDM	0.04	GOULART <i>et al.</i> , 2009
Atlantic Forest	MMDM	0.21	FUSCO-COSTA <i>et al.</i> , 2010
Amazonia	½ MMDM	0.75 – 0.95	KOLOWSKI & ALONSO, 2010
	MMDM	0.43 – 0.59	
Atlantic Forest	MMDM	0.14 – 0.26	Current study

latitude (DI BITETTI *et al.*, 2008), is not supported. The current study was carried out in the same region as that of DI BITETTI *et al.* (2008), but showed considerably higher density estimates. There are other variables, such as prey density or type of vegetation cover that could have a stronger influence (OLIVEIRA *et al.*, 2010). Consequently, we believe that more studies on ocelot density are needed across the species range before further generalization can be made regarding the patterns and their causes and effects.

Based on the area of Turvo State Park and our density estimates, we suggest a population of 24 to 45 individuals occurs within the Park. At least 26 different animals were recorded during two years of study, which is in accordance with our minimum estimate. We lack data on population turnover from migration or mortality, but seems clear that densities at Turvo Park are higher than previously observed for the Misiones region by DI BITETTI *et al.* (2008).

Turvo State Park, the Misiones Green Corridor and ocelot conservation: The population size (24-45) estimated for ocelots in Turvo is small below the minimum viable population of SHAFFER (1981), and the minimum effective population size proposed by FRANKLIN (1980) (both of 50 individuals) required to prevent inbreeding and its deleterious effects. However, to date the park is not isolated. In fact Turvo State Park is part of the Misiones Region of Argentina, a forested area known as the Green Corridor. This region is part of a relatively well-preserved area of about 12,000 km² that extends up to Iguacu National Park to the north. This area also represents the southernmost ranges of jaguar, tapir *T. terrestris* and white-lipped peccary *Tayassu pecari*. Movements of animals between the Misiones Region and Turvo State Park has been recorded for a variety of species, including jaguar (PAVILOLO *et al.*, 2006) and ungulates. We believe that the Uruguay River do not represent a barrier to this felid. Consequently, the Park's and the Argentinean population are one. Additionally, ocelot is also found in fragmented areas (OLIVEIRA, 1994), which also suggests that movement between the park and nearby areas is possible.

Although ocelots are listed as least concern in Brazil and worldwide (CASO *et al.*, 2015), this felid is threatened regionally. It was considered Vulnerable in Argentina (DIAZ

& OJEDA, 2000), and regionally in the state of Rio Grande do Sul (FZB, 2014), which comprise the Green Corridor of the southernmost Atlantic Forest. Given the known densities for ocelots in the Green Corridor, we would expect a total population between 600 and 3,120 individuals (based in the lower values from DI BITETTI *et al.*, (2008) and our higher estimates). Using a reasonable average density of 0.14 ocelots/km², we can estimate a population size around 1,680 individuals. A population of this size is nevertheless within the 1,250 effective minimum viable population size proposed for the long-term conservation of felids (OLIVEIRA, 1994). This is suggestive that as long as the gene flow is maintained and the area is not deforested, this, probably the southernmost population of ocelots, should endure.

However, the best estimate of 45 animals for the entire Park in a scenario of isolation would suggest a critical situation for the local ocelots. Hydroelectric plants are being planned for the region, which would probably cause the increase of river width, which depending on its size, could diminish or even halt gene flow between Turvo and Misiones. Thus, conservation of the ocelot and of all the large cats at Turvo State Park will only be guaranteed over the long term if the Misiones Green Corridor is preserved and remains connected, without barriers.

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