Mineralogy, zinc kinetic adsorption and sequential extraction of contaminated soil in Manaus, Amazon

Mineralogia, cinética de adsorção e extração sequencial de zinco de um solo contaminado em Manaus, Amazônia, Brazil

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ABSTRACT

Zinc is a heavy metal that enters the environment as consequence of both natural and anthropogenic activities. The association of the soil geochemical fractions and adsorption kinetics were studied to evaluate the behavior of the Zn in a contaminated Utisol soil. The contaminated soil samples were collected at depth 0-20, 20-40, 40-60, and 60-80cm into the Manaus Industrial District (MID) region. The study was performed in soil fractions at 50µm being the samples submitted to sequential extraction, batch experiments, X-ray diffraction (XRD) and Fourier transformed infrared spectroscopy (FT-IR). The adsorption kinetics was fitted according to Lagergren equation. The physical and chemical proprieties of soil showed that organic matter has low cation adsorption capacity and the existence of kaolinite poorly ordered, monoclinic and mixed with others mineral. The kinetics experiments performed indicated that the Zn is adsorbed by sorption process most probably with the formation of two bidentate surfaces. The Zn contents are above established by CETESB (reference value 60mg kg⁻¹) being observed remarkable differences between exchangeable and residual fraction.

Key words: Manaus Industrial District, utisol soil, kaolinite, XRD, FT-IR

RESUMO

O zinco é um metal pesado que é introduzido no ambiente como consequências de atividades naturais e antrópicas. A associação das frações geoquímicas do solo e a cinética de adsorção foram estudadas para avaliar o comportamento do Zn em um Argissolo Vermelho-Amarelo. As amostras de solo contaminado foram coletadas nas profundidades de 0-20, 20-40, 40-60 e 60-80cm na região do Distrito Industrial de Manaus (MID). O estudo foi realizado nas frações do solo 50µm, sendo as amostras submetidas à extração sequencial, experimentos de adsorção, difração de raios X (DRX) e espectroscopia no infravermelho com transformada de Fourier (IV-TF). A cinética

de adsorção foi ajustada de acordo com a equação de Lagergren. As propriedades físicas e químicas do solo mostram que a matéria orgânica tem baixa capacidade de adsorção e a existência de caulinita pobremente ordenada, monoclina e misturada com outros minerais. Os experimentos de cinética indicaram que o Zn é adsorvido por um processo de sorção, muito provavelmente, parecido com a formação de superficies bidentadas. A quantidade de Zn está acima do estabelecido pela CETESB (valor de referência de 60mg kg²), sendo observada notável diferença entre as frações trocável e residual.

Palavras-chave: Distrito Industrial de Manaus, utisolo, caulinita, DRX, IV-TF.

INTRODUCTION

In elevated concentration, the Zn can pose long-term risks for soil fertility and environment, affecting the microbiologic activities and increasing the phytotoxicity of the ecosystem. Zinc bioaccumulation is a health problem in which the world health organization (WHO) and Brazil's Ministry of Health (Law 2914/2011) regulated the concentration of 5.0 mg L^{-1} as maximum acceptable in drinking water. In fact Zn which has widely released it on the soil, is known as a heavy metal of the Priority Pollutants list (ATSDR, 2013).

In the soil, Zn reacts naturally with clay minerals, organic matter and oxides of metals such as iron and manganese (ABAT et al., 2012). As is known, soils serve as sinks being the clay minerals and organic contents important components for the

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control of the sorption/retention of contaminants such as heavy metals (ARIAS & SEN, 2009). Furthermore, colloidal constituents such as kaolinite are the components responsible for the chemical behavior of Zn in soil (WANG & Li, 2011). Specifically the kaolinite surface is the base of several phenomena, especially adsorption, which occur at the solid-solution interface, and are influenced by soil solution properties such as the ionic strength; pH values (ZHU, et al., 2012).

The study of the kinetics and the thermodynamics of the adsorption process provide important information on heavy metal phytotoxicity. Previous researchers have provided numerous kinetic models used to describe sorption processes. These include first-order, and second-order reversible ones, and first-order and second-order, irreversible ones, pseudo-first-order and pseudo-second-order ones based on the solution concentration (ADEBOWALE et al., 2008). The association in soil fractions as well as the kinetic adsorption capacity was studied in aim to predict the behavior of Zn in a polluted Utisol soil by galvanic industry from MID region that present high levels of heavy metal in its ecosystem.

MATERIAL AND METHODS

Sampling sites

Contaminated Utisol soil samples were collected in three sites sampling from the Manaus Industrial District-MID: IG03 (S 3°06'33,16"; W 59°57"33,75"), IG04 (S 3°06'64"; W59°57'52,51") and IG05 (S 3°06'53,99";W59°58'06,78"), in along of the profile: 0-20, 20-40, 40-60, and 60-80cm. Each soil sampling was subdivided in 4 subsamples. Soil samples were air-dried under room temperature for 24 hours. Approximately 1.0kg of the soil samples were then passed through a 2mm stainless steel sieve

in order to obtain the fine soil fraction (<2mm). The sand, silt, clay fractions, cation (Zn²⁺), organic matter contents, CEC and C/N values were determined by recommendation EMBRAPA (1997) (Table 1).

The fine soil was dispersed in water, sieved to separate the fraction $50\mu m$ and then dried at room temperature. This fraction was used in XRD, FT-IR and batch experiments.

X-Ray diffraction

The X-ray diffraction was obtained in a Geigerflex diffractometer D/max Series B (Rigaku) equipped with monochromator graphite and copper tube. Soil Samples < 2mm were triturated and pressed into an aluminum sample holder, and measurements were made from 5 to 70° (20) at steps of 0.02° with a counting of 0.6s step-1.

FT-IR

The infrared spectra were recorded range 4000 to 400cm⁻¹ on a Perkin Elmer spectrophotometer model Spectrum 2000 and 20 scans were performed with spectral resolution of 4cm⁻¹. The pellets containing about 1.0mg of fraction 50µm and 100mg of KBr were dried at 100°C for 24 hours to eliminate any existing moisture.

Kinetics experiments

The experiments were performed by adding portions of 1.000g of the 0.053mm fraction samples and 20mL of Zn 50mg L $^{-1}$ in 12 flasks in all soil samples (fractions<2mm). The flasks were stirred during 480min. and after intervals of 30min one flask was removed and filtered in *Millipore* (0.45 μ m and diameter 47mm). The Zn concentration was determined by FAAS (Spectrometer GBC AAS932 plus) using an air-acetylene flame and standard reagent certificates (SpecSol) with concentration

	Depth (cm)					
	00-20	20-40	40-60	60-80		
Sand (g kg ⁻¹)	809±2	780±9	801±0.3	750±5		
Silt (g kg ⁻¹)	24.5±0.1	40.0±0.1	23.0±0.6	19.0±0.3		
Clay (g kg ⁻¹)	157.0±0.4	179.5±0.6	182.0±0.4	222.0 ± 0.7		
pH_{KCl}	5.19±0.11	5.08 ± 0.72	5.37±0.34	5.18±0.92		
pH_{H20}	5.19±0.58	5.34 ± 0.07	5.90±0.95	5.63±0.38		
CEC (cmol _c kg ⁻¹)	3.05±0.66	4.24±0.46	4.55±0.69	1.48±0.24		
Organic matter (g kg ⁻¹)	11.7±0.8	1.89±0.03	17.6±0.9	12.8±0.2		
$Zn^{2+}(mg kg^{-1})$	10.34±0.36	8.28 ± 0.07	12.85±0.42	10.77±0.25		
C/N	17.30±0.23	21.10±0.34	19.95±0.14	17.70±0.90		

at 1.000±0.003mg g⁻¹. The flame and lengths waves were set for each element according to the instrumental manual. For all elements, the readings were made only with calibration curves presenting r²≥0.99. The Zn solution was adjusted to pH value 5.00±0.5 with HCl/NaOH that is the variation of the value found in the studied soil samples (Table 1). The data of each experiment were obtained in triplicate. The pseudo-second-order model of the Lagergren equation was chosen based on their applicability and recommendation found in several studies on soils (ADEBOWALE et al., 2008). The pseudo-second-order equation given below was used:

$$q_t = kq_e^2 t$$

The differential equation is as follows:

$$\frac{dq_t}{dt} = \frac{k(q_e + q_t)}{1 + k_t}$$

Integrating for the boundary conditions

t=0 to t=t

$$\frac{1}{q_e - q_t} = \frac{1}{q_e} + k_t$$

When this is linearized, it gives

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}$$

Where q_e is the amount of Zn^{2+} adsorbed at equilibrium (mg L⁻¹), q_t amount of Zn^{2+} adsorbed at time t (mg g⁻¹) and k_2 is the rate constant of the pseudo-second-order sorption (g mg⁻¹ min).

The constant k_2 is used to calculate initial sorption rate h, at $t\rightarrow 0$, as follows

$$h = k_2 q_2^2$$

Sequential extraction

The sequential extraction used by fine soil (50µm) was the same as what was recommended by TESSIER et al., (1979). The extract proceed was carried out at room temperature, in block digestion (Marconi, model MA4025-SCR), with agitation in shaker (Marconi, model MA570), and extraction of the analytical-reagents grade.

RESULTS AND DISCUSSION

It has been known for a long time that the clay and organic matter content of soils control the CEC. One of the most widely used indices for maturity of organic waste that is the C:N ratio in all soil samples is less than 20 indicating an advanced degree of stabilization as well as an equilibrium between mineralization and immobilization process during

organic matter decomposition (Table 1). The organic matter into Utisol soil has low adsorption capacity and consequently its contribution for the adsorption of the cations is relatively low. On the other hand, the ΔpH (pH_{KCl} - pH_{H2O}) values in all samples which indicates the predominance of negative charges.

Despite of the table 1 presents high iron contents X-ray diffraction patterns showed the presence of only kaolinite and quartz in all soil samples (Figure 1). In a particular reference to the reflections they indicate that the kaolinite (clay mineral) has low crystallinity and/or it is non-ordered. Furthermore, it possible to observe the relationship between the intensity of the reflections d_{131} and $d_1\bar{3}_1$, an important feature observed in XRD, which suggest the existence of a monoclinic kaolinite.

The FT-IR spectra in all soil samples show only bands relatives to kaolinite (Figure 1). In 3669, 3653, 3631cm^{-1} exist intense bands of absorption corresponding to the vibrations v_1 , v_2 and v_3 O-H, respectively. The lack of resolution of the vibration v_2 point to the existence of kaolinite with a very poor level of planning evidenced by the increased intensity v_4 in relation to v_1 . PROST et al. (1989) argue that the FT-IR spectrum as described above could be an indicative of the existence of dickite and nacrite instead of a disordered kaolinite. However XRD did not reveal specifically the existence of these minerals but the possibility of having a mixture of minerals.

Bands occur at 1115cm^{-1} Si-O, 1036 and 1009cm^{-1} on vibration Si-O-Si and 937 and 915cm^{-1} O-H. The asymmetry of at least 1.15 between the bands 1115 and 915cm^{-1} , as well as among v_1 - v_4 , points to an intercalate kaolinite. Therefore, FT-IR spectra suggest that kaolinites into Utisol soil are intercalated by cations. However their reactivity is probably low because of the intercalation process that is better for well-ordered kaolinite than the poorly ordered kaolinite. The kaolinite crystalline structure observed in this work is supported by MELO et al. (2001) that claim in Brazil the existence of kaolinites with relatively poor crystal order, with elongated, subspherical large aggregated and rounded forms as well as high stability.

The first-order reaction model was used to check all the results but the correlation coefficient is not high. However the rate law for a pseudo-second order could be fixed with very high correlation coefficient. According to GUERRA et al. (2010) this result depends on the control of the rate of reaction of an adsorption reaction. For example adsorption cation on Amazon kaolinite surface is better adjusted to the experimental kinetic data in constant

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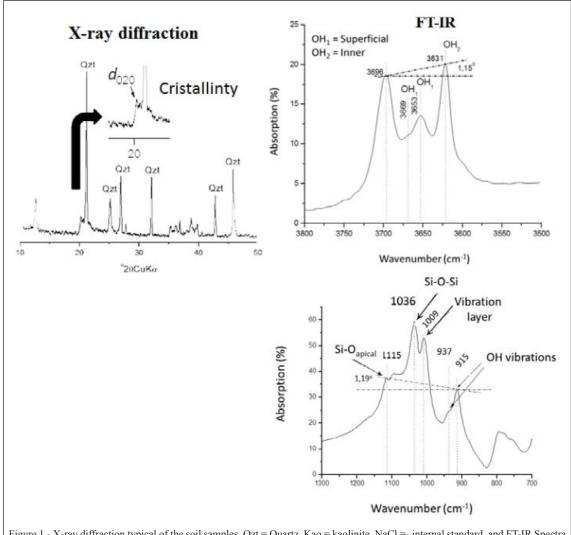


Figure 1 - X-ray diffraction typical of the soil samples. Qzt = Quartz. Kao = kaolinite. NaCl =- internal standard. and FT-IR Spectra of the soil samples.

temperature and monitoring the amount adsorbed with time. The sorption of Zn^{2+} ion onto Utisol soil could be as pseudo-second order process rather than first order. The fitting date for t/q_t versus time t for Zn^{2+} adsorption at different depth in the sampling sites IG03, IG04 and IG05, all parameters including the linear regression coefficients, R^2 , the initial adsorption rate, h, the pseudo-second-rate constant k_2 and equilibrium sorption capacity, q_e , are given in the table 2. The linear regression coefficients suggest two form of metal uptake processes: i) chemisorption and ii) more than one-step for the Zn^{2+} sorption.

As the experiment kinetics was carried out in pH values at 5.0 \pm 0.5 (values of the soil samples) the Zn²+ should be present in the forms of Zn²+, Zn(OH)+, and Zn(OH)_{2(s)}. At pH value ~5.0, the solubility of Zn(OH)+ and Zn(OH)_{2(s)} is high and

therefore, the Zn²⁺ is the specie that take part sorption process (SRIVASTAVA et al., 2005).

The main reason for this is due to the vertical stacking of kaolinite lattices which exposes a large vertical or edge surface area compared to the horizontal or planar surface area. Furthermore, the adsorption which takes place by this electrostatic mechanism is thermodynamically more stable and easier to adsorb than the free metal cation (USMAN, 2007). It occurs as a result of the kaolinite crystalline structure being poorly ordered and its negative charge depending on disruption of the structure at the edges particles.

Table 3 shows that the results found by applying sequential extraction methods to the soil samples. The values included are means of the three replicates with relative standard deviation ranged from 1 to 10% suggesting good repeatability. The

Table 2 – Pseudo-second-order kinetic parameters. q_e =Zn²⁺ adsorbed at equilibrium, q_E Zn²⁺ adsorbed at time t and k_2 = pseudo-second-order rate constant.

Sampling	Depth (cm)	$q_e({ m mg~g}^{ ext{-}1})$	k_2	h		2
			(g mg ⁻¹ min ⁻¹)		Y = AX + B	\mathbb{R}^2
IG03	00-20	0.8496	-0.1970	-0.1422	t/qt = 1.1769 t - 7.0286	0.9934
	20-40	0.5154	-0.1726	-0.0458	t/qt = 1.9401 t - 21.7984	0.9888
	40-60	0.4007	3.8773	0.622 7	t/qt = 2.4953 t + 1.6059	0.9921
	60-80	0.5517	0.1264	0.0384	t/qt = 1.8123 t + 25.9774	0.9844
IG04	00-20	0.4355	-0.1148	-0.0217	t/qt = 2.2957 t - 45.9042	0.9875
	20-40	0.4876	-0.1833	-0.0435	t/qt = 2.0505 t - 22.9372	0.9894
	40-60	0.4465	0.4664	0.0930	t/qt = 2.2392 t + 10.7487	0.9847
	60-80	0.4355	-0.1148	-0.0217	t/qt = 2.2957 t - 45.9042	0.9875
IG05	00-20	0.3340	-0.1795	-0.0200	t/qt = 2.9938 t - 49.9210	0.9719
	20-40	0.3310	0.5094	0.0558	t/qt = 3.0207 t + 17.9131	0.9755
	40-60	0.2361	0.2230	0.0124	t/qt = 4.2341 t + 80.37	0.9821
	60-80	0.2733	-1.7264	-0.1290	t/qt = 3.6577 t - 7.7490	0.9912

greatest values in the relative standard deviations were usually obtained in the phases and Zn present in the soil with lowest concentrations, especially if they are near the limits of detection of the analytical methods used. The Zn total contents (the sum of the steps of the sequential extraction) in all soil samples are above those established by CETESB (60mg kg⁻¹) for soils from State of São Paulo.

The presence of Zn in all operational fractions indicates that this metal is distributed between the bioavailable/mobile pool (exchangeable, carbonates, oxidisable and reducible fractions) and residual fraction, representing 10-60 and 90-40%

respectively. Furthermore the Zn content into each selective treatment fraction varied according to sampling site mainly in the exchangeable and residual fractions. It was observed a remarkable difference in the amount of Zn in the exchangeable fraction depending on depth and sampling site.

The Zn content into each sequential extraction fraction varied according to sampling site mainly in those fractions corresponding to the operationally defined as exchangeable and residual. The following order IG04>IG05>IG03 is observed for exchangeable fraction. Depending on changes of the redox condition mainly Eh the Zn can be easily

Table 3 - Zn contents (mg kg⁻¹) extracted of IG03, IG04 and IG05 by sequential extraction proposed by Tessier et al. (1979) from Manaus Industrial District –MID.

Sampling	Depth (cm)	Exchangeable Fraction	Bound to Carbonate	Bound to Iron and Manganese	Bound to Organic Matter	Residual fraction
IG03	00-20	4.0±0.5	4.6±0.3	1.10±0.03	7.4±0.5	143.2±3.4
	20-40	2.9±0.1	3.2 ± 0.1	1.10 ± 0.02	9.1±0.2	147.1±1.6
	40-60	4.6 ± 0.2	3.9 ± 0.7	4.00 ± 0.08	4.1±0.6	150.3±8.1
	60-80	5.3±0.3	3.7±0.4	1.10±0.07	12.9±0.9	161.4±7.0
IG04	00-20	75.3±0.7	15.6±0.9	13.30±0.06	29.4±0.8	104.1±5.8
	20-40	51.8±0.1	13.1 ± 0.4	9.70 ± 0.08	31.5±0.9	100.3 ± 8.4
	40-60	105.0 ± 0.7	19.7 ± 0.8	11.70 ± 0.16	18.9 ± 0.1	105.6±1.6
	60-80	77.0±0.3	12.6±0.2	8.90 ± 0.08	13.7±0.1	102.3±8.4
IG05	00-20	25.9±0.9	7.4±0.8	4.20±0.07	10.9±0.3	99.1±7.2
	20-40	28.3±0.7	11.4 ± 1.0	5.30 ± 0.71	17.0 ± 0.4	90.8±7.1
	40-60	28.3±0.2	7.5 ± 0.6	5.00 ± 0.71	16.0 ± 0.9	107.6±7.1
	60-80	24.6±0.9	7.0 ± 0.5	4.50±0.24	17.5±0.8	113.1±2.4

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released into the environment (TESSIER et al., 1979). Furthermore, the kinetic findings show the Utisol soil has low capacity of the adsorption and consequently the Zn mobilization is high.

Despite of the Utisol soil from Manaus have low carbonate content, the acid-extractable fraction, operationally defined as bound to carbonates, was high and varied depending on sampling site. Zinc may be present in this fraction as co-precipitated with other isostructural carbonates, e.g. siderite (FeCO₃) commonly found in the ore gangue (ETTLER et al., 2007). It imply that an appreciable contents of Zn would have been remobilized becoming readily available following a slight lowering of pH.

CONCLUSION

The C:N ratio indicated the presence of organic matter with low cation adsorption capacity. The XRD and FT-IR indicate the existence of kaolinite poorly ordered, monoclinic, mixed with others mineral, such as halloysite, dickite, nacrite, quartz, illite, feldspars. The kinetics data were better adjusted by pseudo-second order equation for Lagergren model. The adsorption kinetics showed the Zn is adsorbed by sorption process most probably with the formation of two bidentate surface complexes. The Zn contents are above established by CETESB being observed remarkable differences between exchangeable and residual fraction.

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