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Puls method for events simulation in a lot scale bioretention device

Método de Puls aplicado a um sistema de biorretenção em escala de lote para simulação de eventos

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ABSTRACT

Bioretention is a technique applied to treat and reduce runoff generated in urban areas. Although these facilities present complex processes, a simple model with few parameters can be useful for its project. This paper reports an application of Puls method as a bioretention simulation quantitative tool for events. A bioretention device in real scale was installed and monitored using four level sensors and a tipping-bucket rain gauge, both with data loggers. The method was applied in two ways: the facility as a single reservoir; and the facility as two serial reservoirs. Outflows depended on a single infiltration rate in the first case (K) and two infiltration rates in the second case (K1 and K2). The study used 15 events. Results showed suitable values for Nash-Sutcliffe (NS) and determination coefficients (R²) in the calibration phase, especially for bioretention simulated as a single reservoir, with NS values ranging from 0.64 to 0.95. NS ranged from 0.80 to 0.95 in the calibration phase for the device simulated as two serial reservoirs. The validation phase achieved lower NS and R² than calibration phase. Despite this, the present study suggests the Puls model as a good option for research on bioretention devices modelling.

Keywords: Bioretention; Compensatory techniques; LID; Puls model.

RESUMO

Biorretenção é uma técnica aplicada para tratar e reduzir o escoamento superficial gerado em áreas urbanas. Embora essas instalações apresentem processos complexos, um modelo simples com poucos parâmetros pode ser útil para seu projeto. Este artigo relata uma aplicação do método Puls como uma ferramenta quantitativa de simulação de biorretenção para eventos. Um dispositivo de biorretenção em escala real foi instalado e monitorado usando quatro sensores de nível e um pluviômetro de báscula, ambos com registradores de dados. O método foi aplicado de duas maneiras: a instalação como um único reservatório; e a instalação como dois reservatórios em série. As vazões dependiam de uma única taxa de infiltração no primeiro caso (K) e duas taxas de infiltração no segundo caso (K1 e K2). Foram utilizados 15 eventos e os resultados mostraram coeficientes de Nash-Sutcliffe (NS) e de determinação (R²) adequados na fase de calibração, especialmente quando a biorretenção foi simulada como um único reservatório, com valores NS variando de 0,64 a 0,95. Quando o dispositivo foi simulado como dois reservatórios em série, os valores de NS variaram de 0,80 a 0,95 na fase de calibração. A fase de validação resultou em NS e R² menores que a fase de calibração. Apesar disso, o estudo apresentou o modelo de Puls como uma boa opção para a pesquisa em modelagem de dispositivos de biorretenção.

Palavras-chave: Biorretenção; Técnicas compensatórias; LID; Modelo de Puls.

INTRODUCTION

Bioretention is a stormwater management technique, part of Low Impact Development (LID) approach to dampen peak flow rates and decrease runoff volumes. It also aims to increase local infiltration rates allowing storage in inner layers pores of the soil (HUNT et al., 2008). Aesthetically, this device occurs as gardens allocated in parking places, roadsides and residences (HUNT et al., 2015). Despite all benefits, hydraulics complexities, types of sediments and complex biological reactions difficult detailed simulations of bioretention by computational models (BARKDOLL et al., 2016).

Some used models

Common models used to simulate bioretention systems are DRAINMOD, RECARGA and SWMM (LIU et al., 2014; WINSTON; DORSEY; HUNT, 2016), GIFMod (MASSOUDIEH; AFLAKI; PANGULURI, 2016) and MUSIC (IMTEAZ et al., 2013).

DRAINMOD is a continuous simulation model. Brown, Skaggs and Hunt III (2013) used it to simulate bioretention cells in Nashville and Rocky Mount, North Carolina. Seepage parameters were easier calibrated when an internal water storage zone was present. Water levels could be used solely to calibrate the seepage parameters instead of trying to calibrate simultaneous processes. Initially, DRAINMOD was not designed to model runoff of urban areas, but it has been restructured to allow high temporal resolution inputs and outputs in these areas (LISENBEE; HATHAWAY; WINSTON, 2018).

RECARGA model simulates water movement over the entire bioretention, receiving data of soil moisture and water volume at each step of the process at the hydrological cycle (ATCHISON; POTTER; SEVERSON, 2006). Boancă et al. (2018) used this model and concluded that the variable soil texture is the most important element of control in the design and performance of bioretention cell, and hydraulic conductivity has a large effect on the duration of flooding in bioretention cell.

GIFMod (Green Infrastructure Flexible Model) can be used to simulate several facilities in the LID approach (MASSOUDIEH et al., 2017). MUSIC (Model for Urban Stormwater Improvement Conceptualization) simulates bioretention and other facilities that evolve dimensions and local soil properties (IMTEAZ et al., 2013). Montaseri, Hesami Afshar and Bozorg-Haddad (2015) used a combination of MUSIC and genetic algorithms to optimize a treatment train setting for the post-development scenario near the district of Central Canberra, in Australia.

SWMM (Storm Water Management Model) models bioretention as an artificial depression that contains planted vegetation in prepared soil layer to remove sediments and pollutants at runoff (ROSSMAN, 2015). Wang et al. (2016) e Li, Li and Li (2016) used it to simulate a bioretention facility.

Although a great variety of computational models are available to represent and evaluate bioretention systems, a model including important processes involved in biomass change, plant uptake and defoliation is required. This type of model can provide the identification of elements transferred to the groundwater (LIU et al., 2014).

Use of Puls model

Simple models also must be able to represent water balances. In Puls model, output flow rates are functions of input flow rates and reservoir features (ZOPPOU, 1999). Puls model has been gaining adepts to simulate source control facilities. For example, Barbassa, Angelini Sobrinha and Moruzzi (2014) applied it to evaluate the performance and assist the design of an infiltration well, and Tecedor et al. (2015), to modelling an infiltration plane.

Rosa (2016) employed the Puls method as one of the phases of BIRENICE model, created by the author to dimension and simulate a bioretention on a size bigger than a lot scale. Main parameters used were: return time, soil properties, catchment area and voids percentage. Infiltration, drain pipe, and infiltration plus drain pipe are the three ways to release water from a bioretention. According to Sievers and Pinheiro (2013), it is also possible to insert a coefficient to represent clogging. This method neglect evapotranspiration effect, but this omission may represent less than 1% in the reduction of output (JENNINGS, 2016).

The present research proposed the use of Puls Model (also known as Level Pool Routing) to simulate a quantitative process in a bioretention system at small-scale working as a reservoir where water output happened only by means of infiltration.

Therefore, the objective was to evaluate the application of a model based on Puls method treating the output as an infiltration rate and using monitoring quantitative data from the bioretention device. The motivation was the need to develop user-friendly models with few parameters so that managers can assess the benefits of these practices on hydrology (LIU et al., 2014). In Brazil, Rosa (2016) applied the software BIRENICE to simulate a bioretention in an urban catchment area of 2.3 ha. This study used the Puls model to simulate quantitative aspects only at a small-scale control.

MATERIAL AND METHODS

Studied area

A bioretention was installed, in real scale, at the Centre of Technology (CTEC), Federal University of Alagoas (UFAL), to retain and treat the roof runoff of a building called CEENG (Knowing and Experiencing Engineering – Conhecer e Experimentar a Engenharia, *in Portuguese*) and the grass runoff from a nearby area. Geographic coordinates are 9°33'06.8''S 35°46'31.7''W (Figure 1).

The monitored bioretention has a surface area of 6.64 m² (Figure 1) and three layers separated by a geotextile blanket, which allows the movement of water inside the device, without communication between the different materials. The first layer, at the bottom, is above local soil and filled with 10 cm of sand. 72 cm of gravel called *brita 1* fills the second layer, with granulometry ranging from 9.5 mm to 19 mm (BRASIL, 2009) and the third layer with 19 cm of black soil. These thicknesses represent mean values because they changed slowly along the length. Rosendo (2013) and Menezes (2015) detailed the sizing.

The surface of the device is around 30 cm below ground level. 9 *Philodendron* plants (popularly known as Imbé) were placed on the surface. According to Guihong and Ruijun (2017), this is

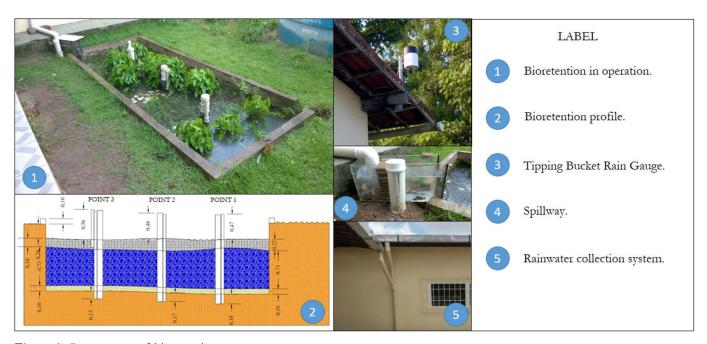


Figure 1. Components of bioretention.

one type of plant that can be applied in bioretention because its roots are great to improve layer structure and ornamentation, water retention and help to remove pollutants.

A tipping-bucket rain gauge was installed on a roof near the device (15 m) because is a safe place and far from trees. Three monitoring points were chosen for water balance and water quality (Figure 1). The levels were monitored using four level sensors. Three of them were at the bioretention and one was used at the spillway.

Figure 1 shows one level sensor at point 1 (far from spillway) used to verify the level inside the device, and two level sensors at point 3 (near the spillway) used to monitor the level inside the device (called S3) and above upper layer (called SS). Dynamics infiltration at the upper layer was observed, independently of the inner layer.

Collected data from sensors were formatted and separated into independent events. In this study, the word event means that precipitation raises water level inside the bioretention beyond sand layer, followed by a reduction to level zero recorded by sensors S1 and S3. No water elevation occurred during the next 12 hours, or during the 12 hours before events. Each event received a code following the Julian calendar with two units, according to the year, month and day that started the rain (YY/MM/DD). For example, the event occurred on June 24, 2015, received code 150624.

Puls model applied to bioretention

Puls model used a discretized expression for lumped continuity equation (Equation 1) and a relationship between storage and output flow rate from a reservoir. Unknowns were flow rate Q and storage S at time t + 1. They were obtained using a level versus output flow rate curve (TUCCI, 2005). Discretized Equation 1 became Equation 2:

$$\frac{S_{t+1} - S_t}{\Delta t} = \frac{I_t + I_{t+1}}{2} - \frac{Q_t + Q_{t+1}}{2} \tag{1}$$

$$Q_{t+1} + \frac{2S_{t+1}}{\Delta t} = I_t + I_{t+1} - Q_t + \frac{2S_t}{\Delta t}$$
 (2)

where: I_t and I_{t+1} are input flow rate in the reservoir at time intervals t and t+1; Q_t e Q_{t+1} are output discharge from the reservoir at time intervals t and t+1; S_t e S_{t+1} are storages in the reservoir at time intervals t and t+1; and Δt is the simulation time interval.

Output flow rate resulted from infiltration on the sides and bottom of the device. Angelini Sobrinha (2012) also considered this condition when modelled an infiltration well, and Lucas, Barbassa and Moruzzi (2013) when represented an infiltration trench. The discretized lumped continuity equation takes the form expressed in Equation 3.

$$Q_{\inf t + 1} + \frac{2S_{t+1}}{\Delta t} = I_t + I_{t+1} - Q_{\inf t} + \frac{2S_t}{\Delta t}$$
 (3)

The level versus flow rate curve was obtained using Equation 4, which presents a product of infiltration rate (K) by the total infiltration area:

$$Q_{\inf} = K \cdot \left[LC + 2H(L+C) \right]$$
(4)

where: Q_{inf} is the output flow rate by infiltration (m³/s); K is the infiltration rate (m/s); L is the width of bioretention (m); C is the length of bioretention (m); and H is the water level inside bioretention.

Traditional steps for simulation with Puls model are (LUCAS; BARBASSA; MORUZZI, 2013):

- 1. Plot Q_{inf} x H curve;
- 2. At the t+1 interval, the value of $[Q_{inft+1} + 2S_{t+1}/\Delta t]$ is determined because terms on the right of Equation 3 are known;
- 3. Output flow rate is obtained from [$Q_{inft+1} + 2S_{t+1}/\Delta t$] through Qinf x H curve;
- 4. The water level is calculated according to Δt .

Simulating one and two reservoirs

The method was applied in two pathways: the facility as a single reservoir, like an infiltration trench, in the first; and the facility as two serial reservoirs in the second. The output from the upper reservoir was the input to the lower reservoir. Therefore, in the first case, the parameter for calibration was a unique infiltration rate. In the second case, the parameters were the two infiltration rates.

A total of 37 rainfall events were recorded during the monitoring period between June 2015 and September 2016. This period was divided into two phases: in the first (23 events), water level monitoring performed from 2 level sensors and in the second phase (14 events), from 3 level sensors.

At the initial phase, selected events had maximum water levels greater than 20 cm. This value was defined by a sand layer thickness of approximately 10 cm. In the second phase, 3 events were selected because they have generated water level higher than surface level, above the upper layer of the bioretention (Table 1).

In the final phase of monitoring (May to September 2016), only 565 mm of precipitation were registered. This value is almost half of that value obtained in the same period of the previous year (971.4 mm). The second phase was much less rainy.

Table 1 presents the event code, total rainfall, rainfall duration and maximum water level inside bioretention. They were obtained at events monitoring from 150624 to 160529, collected from a tipping-bucket rain gauge, a calibrated spillway to provide input flow rates (hydrographs) and water level from level sensors inside bioretention at points S1 and S3.

Events 160602, 160619 and 160628 had the same information that previous events, and additional information from another level sensor used to measure water level in the upper layer. Thereby, it was possible to calibrate two infiltration rates, one at upper layer (level sensor SS) and other in the gravel layer (level sensors S1 and S2).

Data from events 150624 and 160529 were used to simulate bioretention like an infiltration trench, i.e., as a single water reservoir. In this case, it was necessary to determine fixed

Table 1. Events chosen to model the bioretention.

Quantity	Event	Precipitation	Duration	Maximum
Quantity	Event	(mm)	(h)	level (m)
1	150624	63.2	18.2	1.06
2	150722	42.4	36.8	1.09
3	150815	8.8	7.3	0.25
4	150818	10.6	10.0	0.27
5	150820	63.6	28.8	0.93
6	160121	15.4	2.7	0.74
7	160129	28.2	5.2	0.44
8	160413	45.4	31.8	1.05
9	160424	16.6	6.8	0.27
10	160522	37.8	9.2	0.70
11	160524	14.6	7.3	0.38
12	160529	146.8	33.1	1.19
13	160602	22.4	13.4	0.34
14	160619	29.6	24.2	0.47
15	160628	16.0	9.3	0.28

values for porosity and thickness for each layer, beyond calibrate infiltration rate per event.

Simulations of events 160602, 160619 and 160628 were based on a structure with two serial reservoirs. The output hydrograph from the upper reservoir was the input hydrograph to the lower reservoir. Water level data from an area above the upper reservoir (level sensor SS) were used to calibrate the infiltration rate.

Calibration and validation

Infiltration rate was the only parameter calibrated. Calibration was performed by trial and error methods, and visualization of graphs. Nash-Sutcliffe (NS) and determination R² coefficients were used in the evaluation of results. NS (Equation 5) ranges from 1 to -∞, with 1 as ideal value (NASH; SUTCLIFFE, 1970) and R² (Equation 6) ranges from 0 to 1, with 1 as also ideal value (MONTGOMERY; RUNGER, 2010). Beyond them, relative error (E) was used to verify discrepancies between observed and simulated maximum water levels.

$$NS = 1 - \frac{\sum_{i=1}^{n} (Xobs - Xsim)^{2}}{\sum_{i=1}^{n} (Xobs - \overline{X}sim)^{2}}$$
(5)

$$R^{2} = \frac{\left[\sum_{i=1}^{n} \left(Xobs - \overline{Xobs}\right) \cdot \left(Xsim - \overline{Xsim}\right)\right]^{2}}{\sum_{i=1}^{n} \left(Xobs - \overline{Xobs}\right)^{2} \cdot \sum_{i=1}^{n} \left(Xsim - \overline{Xsim}\right)^{2}}$$
(6)

where: n is the amount of data; Xobs is the observed data; $\overline{\text{Xobs}}$ is the mean of observed data; Xsim is the simulated data; and $\overline{\text{Xsim}}$ is the mean of observed data.

Roy-Poirier, Filion and Champagne (2015) used NS to compare simulated outflows. Figure 2 shows the outline used in the present study.

After inserting the Puls model in a worksheet, input flow rates (m³/s) and corresponding times were also inserted. A value

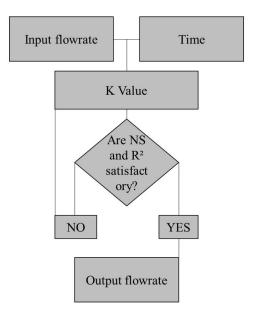


Figure 2. Flowchart of K calibration route (Bioretention as only one reservoir).

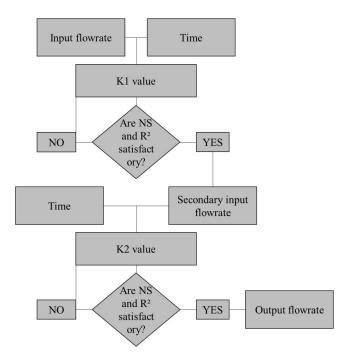


Figure 3. Flowchart of K calibration route (Bioretention as two serial reservoirs).

for K (mm/h) was adopted and NS and R² were calculated. In case of uncertainty, K values were changed to approach 1 for NS and R². This procedure was performed for the first 9 events of Table 1 because 3 subsequent events were used at validation. Figure 3 shows the procedures to represent the bioretention with 2 reservoirs, similarly to Figure 2.

Validation was performed in two ways. The first used the infiltration rate mean value from events 160522, 160524 e 160529. The second used the infiltration rate mean values from events where the bioretention was composed of two serial reservoirs, applying these values to simulate events without maximum levels above the upper layer.

Therefore, it was found K1 and K2 at the calibration process for events 160602, 160619 and 16062. Then, these values were used to simulate events 240615, 220715, 150820 and 160529, chosen randomly to validate the values of parameters inside bioretention.

RESULTS AND DISCUSSION

Simulating one reservoir

Table 2 exhibits result from calibrations: events code, K values, maximum observed water levels, $N^{\rm S}$ and R^2 coefficients and relative error. NS varied from 0.64 to 0.95. R^2 varied from 0.80 to 0.95. $K_{\rm med}$ was 26 mm/h, different from the result achieved by Rosendo (2013) of 50 mm/h.

According to Ohnuma Júnior, Silva and Mendiondo (2015), the infiltration rate is influenced by rainfall duration and intensity. Sample linear correlation coefficient between K and mean rainfall intensity was 0.66. Substituting rainfall intensity by rainfall duration, sample linear correlation coefficient decreased to -0.13.

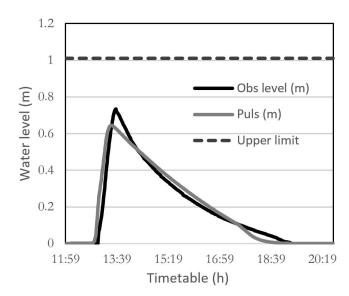


Figure 4. Event 160121.

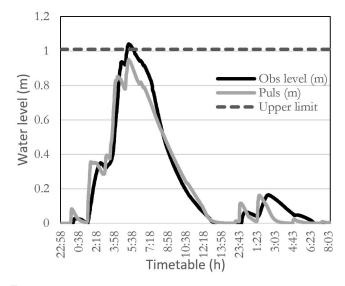


Figure 5. Event 160413.

Figures 4 and 5 show observed variations of water level (Obs. level), water level calculated (Puls) and upper limit of bioretention for the best coefficients in events 160121 and 160413.

While the maximum water level was inferior to the limit of the upper layer in Figure 4, the opposite happened in Figure 5. According to these figures, Puls model simulated the internal water level in a similar way to that observed with only one K calibrated as a parameter. Mean R² was 0.88 and mean NS was 0.84 at the calibration.

Tecedor et al. (2015) used the Puls model to simulate an infiltration plane and found an average R² of 0.72. They used values for infiltration rates from a double ring infiltrometer test. Ferreira, Barbassa and Moruzzi (2018) modelled two infiltration wells and they found mean R² of 0.92 and mean NS of 0.56. They performed a calibration similar to this study. Both studies implemented experiments with total control of water input through precipitation event simulation.

Table 2. Calibrated events for only 1 K value.

Event (1	K	Max. level	Max. level. Puls	NS	\mathbb{R}^2	${f E}$
	(mm/h)	(m)	(m)			(%)*
150624	17	1.06	0.77	0.79	0.84	27
150722	25	1.09	0.97	0.93	0.94	11
150815	20	0.25	0.26	0.84	0.84	4
150818	14	0.27	0.23	0.89	0.89	14
150820	21	0.93	0.56	0.71	0.80	40
160121	38	0.74	0.64	0.95	0.95	13
160129	42	0.44	0.56	0.64	0.85	26
160413	35	1.05	0.95	0.94	0.95	9
160424	22	0.27	0.29	0.86	0.90	5
Mean	26			0.84	0.88	

Table 3. Calibrated events for two K values.

Event		K (mm/h)	Max. level (m)	Max. level. Puls (m)	NS	\mathbb{R}^2	E (%)*
160602	K1	87	0.21	0.14	0.88	0.93	30
	K2	31	0.34	0.29	0.91	0.91	14
160619	K1	147	0.14	0.11	0.90	0.91	18
	K2	42	0.47	0.45	0.97	0.98	5
160628	K1	73	0.06	0.05	0.94	0.94	26
	K2	26	0.28	0.27	0.98	0.98	5
Mean	K 1	102			0.90	0.93	
	K 2	33			0.95	0.96	

^{*}E(%) = relative error.

Simulating two reservoirs

Table 3 presents the results of simulations considering the bioretention as two serial reservoirs. Results employed two values of K (K1 and K2, for the first and the second reservoir, respectively). The first average K (K1 $_{\rm med}$ = 102 mm/h) was the infiltration rate at the upper layer of bioretention. The second average K (K2 $_{\rm med}$ = 33 mm/h) was the infiltration rate inside the device. K1 $_{\rm med}$ was between 70 mm/h and 120 mm/h, found in the performed double ring infiltrometer test.

NS varied from 0.88 to 0.98 and R² varied from 0.91 to 0.98. The relative error between observed and simulated maximum levels was higher for the first reservoir, at event 160602. Calibration with two reservoirs was better than calibration with only one reservoir, evidencing the importance of monitoring two independent layers since they are able to store more water.

Figure 6 and Figure 7 display the performance of two bioretention reservoirs for the events 160619 and 160628. Level zero refers to the bottom of bioretention and level 1 refers to the beginning of the upper reservoir. The rising water level inside bioretention begins at level zero, while water level reaches the surface of the device at level 1. Graphs show the variation of water level recorded in the upper layer level sensor (SS), water levels calculated in the upper layer (SS Puls), observed level variations (Obs level) and calculated level (Puls level) inside the device.

According to Figures 6 and 7, excellent agreement occurred between observed and simulated values for both first and second reservoirs. Table 3 discloses K1 and K2 had great variations. For events 160619 and 160628, K1 achieved 147 mm/h and 73 mm/h, respectively. Probably, this happened because the Puls model was calibrated only using flow rate data provided by the spillway and by water level variation, ignoring other input volumes.

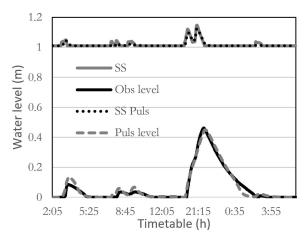


Figure 6. Event 160619.

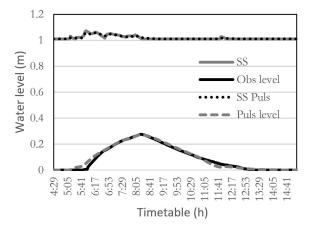


Figure 7. Event 160628.

Table 4. Validation with one value for K.

Event	K (mm/h)	Max. Level (m)	Max. Level (Puls) (m)	NS	\mathbb{R}^2	E (%)*
160522	26	0.70	0.74	0.76	0.93	6
160524	26	0.38	0.39	0.85	0.90	3
160529	26	1.19	1.77	0.48	0.27	49
Mean	26			0.70	0.70	

^{*}E(%) = relative error.

Table 5. Validation with two values of K.

Event	K (mm/h)	Max. Level (m)	Max. Level (Puls) (m)	NS	\mathbb{R}^2	E (%)*
150624	102	-	0.16	-	-	-
	33	1.06	0.52	0.43	0.89	51
150722	102	-	0.17	-	-	-
	33	1.09	0.62	0.89	0.88	43
150820	102	-	0.07	-	-	-
	33	0.93	0.34	0.52	0.92	63
160529	102	-	0.33	-	-	-
	33	1.19	1.05	0.60	0.67	11
Mean	102			-	-	
	33			0.61	0.84	

^{*}E(%) = relative error.

Table 6. Mean coefficients values: calibration and validation.

Layers	s Calibration		Valid	lation
quantity	NS	\mathbb{R}^2	NS	\mathbb{R}^2
1	0.84	0.88	0.70	0.70
2	0.90	0.93	-	-
	0.95	0.96	0.61	0.84

Validation

Initially, the average parameter K was introduced in Puls model to simulate events 160522, 160524 and 160529 (Table 1). Table 4 summarizes the results of the validation, which provided suitable NS and R^2 for events 220516 and 240516, while K of 26 mm/h was not validated for event 290516.

Lucas, Barbassa and Moruzzi (2013) found a greater variation in NS coefficient and relative error (-0.98 to 0.98 and 3.1% to 68%, respectively). They attributed this variation to simplifications in Puls model when considering constant infiltration and ignoring variation between initial and final infiltration capacity provided by the soil moistening. Tecedor et al. (2015) found mean R^2 of 0.71, similar to this study (0.70) when considering the low value of the last event. Thus, in an attempt to calibrate a K to provide the best representation of event 160529, a value of 42 mm/h was found.

Events 150624, 150722, 140820 and 160529 were used to validate K for two reservoirs. The lack of data recording by the level sensor at the SS point restricted the validation. Therefore, events from the first phase monitoring were used.

According to Table 5, NS and R^2 varied from 0.43 to 0.89 and from 0.67 to 0.92, respectively. In this case, NS were far from the unit, with the exception of event 150722. The relative error also showed high values. Adjustment of parameter K in the validation phase was insufficient in comparison with the calibration phase.

When comparing maximum levels calculated and observed, a high relative error was found in the first three simulations. This may have occurred because K calibration sample contained only 3 events, which may have been insufficient to perform calibration, also affecting validation. It would be more interesting to have a number of samples equivalent to that used in the calibration of bioretention as a single reservoir.

Another important point is that K is a constant value, which limits the representation of real infiltration process in the soil (ANGELINI SOBRINHA, 2012), although simulated 160619 and 160628 events were quite similar to the actual events. In addition, the quality of input data is important for better model calibration. Flow rate recorded at 1-minute intervals may have been insufficient to represent the hydrographs. Shorter intervals may be better. Table 6 summarizes the average coefficients generated in both calibration and validation.

CONCLUSION

This study presented the Puls model as a good option for research on modelling the bioretention facility. 15 events were used, and results showed suitable Nash-Sutcliffe (NS) and determination coefficients (R²) in calibration phase, especially for bioretention simulated as a single reservoir, with NS values ranging from 0.64 to 0.95. When the device was simulated as two serial reservoirs, NS values ranged from 0.80 to 0.95 in the calibration phase. Despite acceptable NS and R² in the calibration phase, determination of a single infiltration rate to represent bioretention was unfeasible.

The validation phase, using only three events, resulted in lower values of NS and R² than in the calibration phase, especially for event 160529. When bioretention was simulated as a single reservoir, values of NS ranged from 0.48 to 0.76. When the

device was simulated as two serial reservoirs, NS values ranged from 0.43 to 0.89.

Further studies can be carried out to simulate groups of events with similar characteristics, such as total precipitation, climatic conditions and rainfall return time. It requires further steps, including soil moisture monitoring. It is also interesting to evaluate infiltration rates as a function of depth, a simple alternative that keeps the essence of Puls model. In this case, new calibration parameters can arise, making the model less sophisticated.

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