

Acta Botanica Brasilica 32(4): 578-587. Oct-Dec 2018. doi: 10.1590/0102-33062018abb0013

Together yet separate: variation in soil chemistry determines differences in the arboreal-shrub structure of two contiguous rupestrian environments

Hernani Alves Almeida^{1*}, Thaíse de Oliveira Bahia¹, Irene Gélvez-Zúñiga¹ and Geraldo Wilson Fernandes¹

Received: January 15, 2018 Accepted: May 9, 2018

ABSTRACT

Rupestrian landscapes are characterized by vegetation mosaics comprised of different plant communities and strongly linked to environmental filters. These environments are nutrient-impoverished, and possess water retention deficits and high solar exposure. This study aimed to determine whether chemical properties of the soil shape the arboreal-shrub vegetation structure in two neighboring habitats; a rupestrian cerrado and a rupestrian grassland. We hypothesized that the habitat with higher soil chemical properties would have higher parameters for vegetation structure, and different species composition. We expected higher chemical properties of the soil to favor a greater variety of plant life-forms. A total of 1349 individuals of 85 species belonging to 24 families were recorded; 1141 individuals (65 species) in the rupestrian cerrado, and 208 individuals (20 species) in the rupestrian grassland. Overall, Asteraceae, Fabaceae and Melastomataceae had greater species richness in both areas. Both habitats floristic composition was strongly influenced by differences in soil chemistry. In the cerrado, species richness, composition, soil nutrient availability and acidity were greater than in the grassland. We demonstrated that variation in soil chemistry influences plant species richness and composition, and structural complexity of vegetation, in both contiguous rupestrian environments, and that soil chemistry heterogeneity yields different plant life-form strategies.

Keywords: arboreal-shrub, Asteraceae, *campo rupestre*, Cerrado, edaphic characteristics, environmental filters, Myrtaceae, phytosociology

Introduction

In Brazil, rupestrian grassland ecosystems, known as *campos rupestres*, are well known for their high plant biodiversity and endemism, comprising a mosaic of habitats that are shaped by, and strongly linked to, environmental filters (Fernandes 2016). These environments possess characteristics of old climatically-buffered infertile landscapes (OCBILs), including infertile soils, climatic stability and species with low capacity for dispersion

and adaptations for resource acquisition (Hopper 2009; Silveira *et al.* 2016). Distinct habitats have been recognized within campos rupestres, including: rocky outcrops, peat bogs, gallery forests, stony grasslands, sandy grasslands, natural islands or pockets of cerrado vegetation, islands of Atlantic Forest, and *Candeal* (the latter characterized by the dominance of *Eremanthus erythropappus*), among others (*e.g.* Carvalho *et al.* 2012; 2014; Negreiros *et al.* 2014; Coutinho *et al.* 2015; Coelho *et al.* 2016; Kuster *et al.* 2016).

st Corresponding author: hernanifh@gmail.com



¹ Ecologia Evolutiva e Biodiversidade, Instituto de Ciências Biológicas, Universidade Federal de Minas Gerais, 30161-970, Belo Horizonte, MG, Brazil

Together yet separate: variation in soil chemistry determines differences in the arboreal-shrub structure of two contiguous rupestrian environments

The major filters in rupestrian environments are climate, soil, fire, and water availability (e.g. Negreiros et al. 2014; Barbosa & Fernandes 2016; Figueira et al. 2016), among which the physico-chemical properties of the soil are considered one of the most important (Coutinho et al. 2015; Schaefer et al. 2016). In rupestrian environments, the distribution of various soils in the landscape has resulted in numerous habitats that are inhabited by different plant communities (e.g., Negreiros et al. 2014; Stradic et al. 2015; Mota et al. 2018). The origin of rupestrian habitats is largely due to the influence of topography, slope, soil type, texture, nutrient quantity, lithology and geomorphology, which together contribute to species composition and establishment (Messias et al. 2013; Schaefer et al. 2016).

Rupestrian grassland environments are nutrient-impoverished, and experience seasonal water deficit, high solar exposure, strong winds and characteristics influenced by rocks of chemically poor origin (Oliveira et al. 2015; see also Schaefer et al. 2016; Silveira et al. 2016). Such soils are located among quartzite or sandstone outcrops, and are derived from parent rocks of low nutritional value and with low macronutrient content (Schaefer et al. 2016). In general, rupestrian soils are acidic, shallow and possess high aluminum saturation, plus low levels of potassium (K), calcium (Ca), magnesium (Mg) and available phosphorus (P) (Negreiros et al. 2008). Features like soil texture, acidity, and Al, Ca and Mg ion concentrations are related to the arrangement of the soil mosaic, which is reinforced by levels of C, P, K and cation-exchange capacity (Schaefer et al. 2016).

In spite of the variety of habitats in the rupestrian landscape, sclerophylly is a common feature of its flora (Fernandes & Price 1991), particularly among the scattered shrubs within its herbaceous community (Messias et al. 2013; Barbosa et al. 2015; Stradic et al. 2015; Mota et al. 2016; 2018). Nutrient availability, along with soil depth, drainage, marked seasonality and heavy metal content have resulted in the establishment of species that are highly adapted to the various types of habitat (Benites et al. 2007; Vincent & Meguro 2008; Messias et al. 2011; Mews et al. 2016; Arruda et al. 2017). Different plant life forms are generally a response to changes in the landscape (Pausas & Austin 2001; Neves et al. 2016; Mota et al. 2018), while variation in plant phenotypes may also alter soil communities, thereby affecting the expression of plant traits (Nuland et al. 2016). Woody and/or herbaceous species are favored in deeper soils, while shallow and more impoverished soils harbor species with annual life cycles that have greater underground biomass to support environmental limitations (Mota et al. 2018). Likewise, Neves *et al.* (2016) reported that the most typical life forms of species of the Cerrado possess strategies that increase survivorship during hostile seasonal conditions. Plant species in P-impoverished soils of rupestrian grasslands experience less microorganism colonization and high root specialization, and possess several strategies for acquisition of mineral nutrition (Oliveira *et al.* 2015). The primarily source of organic matter in soils associated with rocky outcrops is from decomposition and microbial activity (Benites *et al.* 2007), hence favoring the establishment of vegetation due better nutrient levels and water retention.

In rupestrian areas, micro-environmental variation can result in functional and morphological differences among plants (Negreiros et al. 2014). Since the rupestrian landscapes of Serra do Cipó are mosaics with microenvironmental variation, we aimed to determine if variation in soil chemistry is shaping the structure and composition of arboreal-shrub vegetation of two contiguous habitats: a rupestrian cerrado and a quartzitic rupestrian grassland. We tested whether the habitat with higher soil chemical properties also had higher parameters for vegetation structure, as well as different species composition. Furthermore, since sizeable variation in soil properties can yield several strategies for acquiring nutritional resources (Araújo et al. 2011; Negreiros et al. 2014; Mota et al. 2018), we also tested whether higher soil chemical properties favor plants with a greater variety of life forms.

Materials and methods

Study area and habitat characterization

The study was conducted in Reserva Vellozia (19°16'45.7"S 43°35'27.8"W) in Serra do Cipó, Minas Gerais, Brazil. The region has a humid subtropical climate, "Cwb" according to Köppen classification, with a hot rainy season from November to March; a transition post-rainy period between April and May; a dry cold season from June to August; and a post-dry period between September and October (Rocha et al. 2016). The mean annual temperature at the study site is 18.5 °C, and the mean annual precipitation is 900 mm (Fernandes et al. 2016). The months with the lowest mean monthly precipitation and temperature are June, July and August. December has the highest mean monthly precipitation with 1000 mm, while January and February have the highest mean temperatures with 21.5 °C (Fernandes et al. 2016). The study area possesses patchy vegetation in the form of a mosaic of substrates, plant communities and micro-environments. Two contiguous habitats were chosen for test our hypothesis: a rupestrian cerrado island (Giulietti et al. 1987) (Fig. 1A) and a rupestrian grassland (Fig. 1B). Both areas are at 1250 m.a.s.l, and are 250 m from each other. The rupestrian cerrado island encompasses only 8000 m², and thus no more than 10 plots were set to assure sampling independence since both sampling areas were contiguous. Despite the fact that fire is a strong environmental filter in rupestrian environments (Figueira et al. 2016), some studies have reported no floristic differences after fires (Lenza et al. 2017), and so we disregard the effect of fire on the mosaic formation since no fires had been recorded for at least ten years in the study area.



Hernani Alves Almeida, Thaíse de Oliveira Bahia, Irene Gélvez-Zúñiga and Geraldo Wilson Fernandes



Figure 1. Habitats of rupestrian cerrado (A) and rupestrian Grassland (B) in Serra do Cipó, Minas Gerais, Brazil.

The rupestrian grassland (hereafter grassland) is characterized by a quartzitic peat bog substrate (Carvalho et al. 2014), commonly with acidic soils, high aluminum saturation, low nutrient concentration, low water-retention capacity and moderate levels of organic matter (Silva 2005). The rupestrian cerrado island (hereafter cerrado) occurs in patches with a dominance of shrub-tree species on a rocky outcrop substrate (Carvalho et al. 2014). Its soils are generally poorly drained oxisols with reduced effective depth and fertility (Amorim & Batalha 2006). The cerrado soils in the region are also nutrient impoverished with high aluminum saturation and a low development profile, and are abruptly becoming rupestrian grasslands (Carvalho et al. 2014).

Vegetation sampling

A circular area with a radius of 50 m was delimited in each habitat for choosing 10 randomized 10x10 m plots (N=20). Each plot was marked with PVC pickets and the soil coverage percentage was estimated using a 1 m2frame, in which all plants with a diameter at soil height (DSH) greater than ≥ 1 cm being measured with a caliper (mm). Plant height was measured with a tape (cm). Classic vegetation structure parameters were calculated for both habitats using species DSH and abundance following Mueller-Dombois & Ellenberg (1974). The parameters estimated were: relative frequency (RF), relative density (RD), relative dominance (RDo), and Importance Value Index (IVI). The IVI indicates the importance of a given species to the community structure by adding its relative frequency, density and dominance. Mean plant height (AH) and stem diameter (AD) were also determined for each species. Collected plant material was processed and identified in consultation with specialists and specialized literature, and through comparisons with the botanical material deposited in the Universidade Federal de Minas Gerais Herbarium (BHCB). Plant species were grouped into families following the Angiosperm Phylogeny Group (APG IV 2016). All the sampled species were identified and their life form characterized according to the virtual herbarium REFLORA (Flora do Brasil 2020).

Soil analysis

A sample of superficial soil (0-20 cm in depth) was collected to analyze the chemical properties of the soil in each plot per habitat (N= 10, total =20). Each sample was a composite of five sub-samples (plot center and vertices). The properties measured were: pH in water (pH), potassium (K), phosphorus (P), remaining phosphorus (P-rem), calcium (Ca²+), magnesium (Mg²+), aluminum (Al³+), hydrogen + aluminum (H+Al), exchangeable bases cmol_c/dm³ (EB), base-saturation % (V), effective cation-exchange capacity cmol_c/dm³ (t), aluminum saturation index % (m), pH 7.0 cation-exchange capacity cmol_c/dm³ (T), organic matter (OM), copper (Cu), manganese (Mn), iron (Fe), and zinc (Zn).

Statistical analysis

To test if differences in plant community structure between the two rupestrian habitats are explained by micro-environmental variation in soil chemistry, a general-linear model was performed (Crawley 2007). The structural parameters set as response variables were: species abundance and richness, relative density, mean plant height, mean stem diameter, and soil coverage; while habitats represented the explanatory variable. For each model, the appropriate error distribution for each response variable was used, according to Crawley (2007).

To determine if species composition is influenced by micro-environmental soil chemical properties NMDS (nonmetric multidimensional scaling) was performed using the Bray-Curtis similarity index, according to Hammer et al. (2001). After a second NMDS, the variables P-rem, Ca^2+ , Mg^2+ , H+Al, SB, V, T and Zn were removed. To test the differences in composition indicated by the NMDS, an ANOSIM (Similarity ANOVA – 5 % of significance) was performed (Clarke & Green 1988). To test whether the occurrence of life forms are explained by microenvironmental soil chemical variation between the two rupestrian habitats we performed another NMDS using the Bray-Curtis similarity index. After the differences in composition indicated by the NMDS were tested, an ANOSIM

Together yet separate: variation in soil chemistry determines differences in the arboreal-shrub structure of two contiguous rupestrian environments

was performed (at 5% of significance). Composition analyses were performed in PAST software (Hammer *et al.* 2001). To assess chemical soil variation between the two habitats, a Generalized Linear Model (GLM) was performed (Sólymos 2009). We used pH, K, P, P-rem, Ca²+, Mg²+, Al³+, H+Al, SB, V, t, m, T, OM, Cu, Mn, Fe and Zn as response variables and habitats as the explanatory variable. For each model, the appropriate error distribution for each response variable was used, according to Crawley (2007). The data were analyzed using R software (R Core Team 2014).

To analyze whether variation in soil chemistry explains species abundance, the "model selection" approach was used. The strength of candidate models was analyzed using Akaike's Information Criterion (AICc), adjusted for small sample sizes, by determining relative difference of AICc between models (ΔAICc) and associated model weights (AICc weighted) (Burnham & Anderson 2002). The effect of micro-environmental variation of soil chemistry on species abundance (response variable) was performed by comparing each model with a subset of models. After determining the best models, the soil properties with greater interaction with species abundance (equal to or greater than 0.5) were identified and used to rank which soil parameters best correlated with the patterns of composition in each habitat. The best models were considered those that were equal to or less than two AICc from the initial model. To select soil chemical variables, a Pearson correlation test was performed, and the variables with correlations greater than 0.8 were excluded from the models. The model selection analysis was performed according to Burnham & Anderson (2002) and were carried out in the software R version 3.2.2 (R Core Team 2014), using the MuMIn package (Bartón 2012).

Results

Differences in floristic composition between the rupestrian habitats were strongly influenced by microenvironmental soil chemical variation (R2= 0.89 and p<0.001), as hypothesized. In the contiguous habitats, cerrado and grassland, variation in soil chemistry was strongly aggregated, with some homogeneity within habitat (Fig. 2). Axis 1 explained 90 % of the segregation with a stress value of 0.11. As expected, soil chemistry properties varied considerably in each habitat. Furthermore, each habitat also exhibited differences in plant life forms (R2= 0.82, P= <0.001). Axis 1 explained 83 % of the variation with a stress value of 0.07 (Fig. 3). Additionally, a predominance of arboreal, shrub and palm shrub species were found for the cerrado, while shrubby species dominated the grassland.

The studied habitats varied in soil pH, K, P, Ca²+, Al³+, SB, V, t, m, T, MO, Cu, Mn, Fe, and Zn concentrations (Tab. 1). The cerrado habitat had higher mean values than the grassland habitat for K, Ca²+, Al³+, SB, t, T, V, Cu, Mn, and Zn. (Tab. 1, Fig. 4). The levels of Al, Cu, K, Mn and effective cation-exchange capacity (t) were significantly higher in

the cerrado habitat (Fig. 4 A, C, E, G, I); while soils of the grassland habitat had greater Fe, R, organic matter, pH levels and aluminum saturation (m) (Fig. 4 B, D, F, H, J).

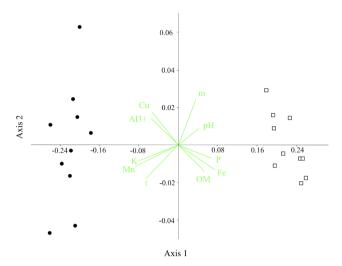


Figure 2. Soil chemical properties ordination in rupestrian cerrado (circle) and Grassland (square) habitat plots (N=20), in Serra do Cipó, Brazil. OM: organic matter, m: Aluminium saturation index, t: effective cation-exchange capacity.

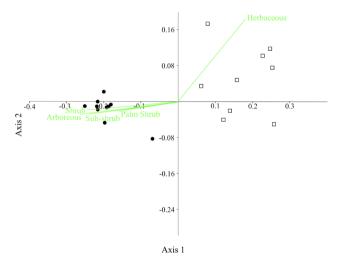


Figure 3. Life-form strategies ordination in rupestrian cerrado (circles) and Grassland (square) habitat plots (N=20), in Serra do Cipó, Brazil.

The most abundant plant species were influenced differently by micro-environmental variation in soil chemistry (Tab. 2). The most abundant and frequent species in the cerrado habitat were: *H. ochraceus, Mimosa radula* Benth., *M. guianensis, Erythroxylum campestre, Eugenia* sp., Ruellia villosa, M. maguirei, C. papillata, Vernonanthura polyanthes, Diospyros hispida, Roupala Montana, Byrsonima variabilis, Eugenia punicifolia, Kielmeyera petiolaris, and Psidium firmum. These species were more likely to occur with higher K, T and Al levels; and were negatively related with P, OM, m, and Zn levels. In the grassland habitat, C. hilariana, R. polymorpha, A. bongardii, and P. argenteus were



Hernani Alves Almeida, Thaíse de Oliveira Bahia, Irene Gélvez-Zúñiga and Geraldo Wilson Fernandes

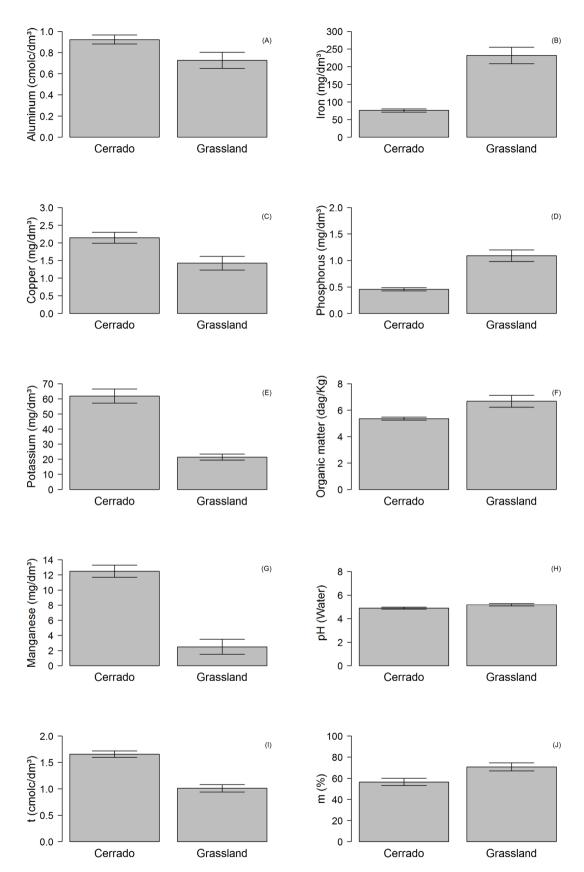


Figure 4. Soil physical properties in rupestrian cerrado and Grassland habitats in Serra do Cipó, Brazil. Aluminum (**A**), Iron (**B**), Copper (**C**), Phosphorus (**D**), Potassium (**E**), Organic Matter (**F**), Manganese (**G**), pH (**H**), t: effective cation-exchange capacity (**I**) and m: Aluminium saturation index (**J**).

Together yet separate: variation in soil chemistry determines differences in the arboreal-shrub structure of two contiquous rupestrian environments

the most abundant and frequent species. The species *P. argenteus*, *C. hilariana* and *A. bongardii* were less represented in the cerrado habitat, and were positively influenced by higher levels of organic matter, P, and Zn; and negatively related to Al and K concentrations.

Table 1. Soil chemical properties variations (Mean \pm SD) between two rupestrian habitats (rupestrian cerrado and Grassland) in Serra do Cipó, Brazil. H+Al: hydrogen + aluminum cmol_c/dm³, EB: exchangeable bases cmol_c/dm³, t: effective cation-exchange capacity cmol_c/dm³, T: cation-exchange capacity at pH 7.0 cmol_c/dm³, V: base-saturation %, m: aluminum saturation index %, OM: organic matter dag/Kg, P-Rem: remaining phosphorus ml/L. * indicates significant values (P<0.05).

8						
Attribute	"Ceri	ado"	Grassland			
	μ		μ			
pH*	4.85	± 0.09	5.25	± 0.08		
P*	0.46	± 0.02	1.15	± 0.10		
K*	59.4	± 4.80	19.8	± 1.28		
Ca ^{2*}	0.36	± 0.07	0.17	± 0.03		
Mg^2	1.76	± 1.58	0.05	± 0.01		
Al ^{3*}	0.94	± 0.04	0.69	± 0.07		
H+Al	8.04	± 0.24	7.79	± 0.13		
EB*	0.69	± 0.08	0.28	± 0.04		
t*	1.63	± 0.06	0.97	± 0.06		
T*	8.74	± 0.27	8.07	± 0.09		
V*	7.90	± 0.88	3.47	± 0.54		
m*	58.35	± 3.57	70.40	± 4.12		
MO*	5.23	± 0.16	6.93	± 0.40		
P-Rem	11.74	± 0.69	12.21	± 0.72		
Cu*	2.22	± 0.16	1.27	± 0.13		
Mn*	12.46	± 0.72	1.53	± 0.26		
Fe*	76.09	± 3.85	247.28	± 19.07		
Zn*	2.65	± 0.34	1.49	± 0.21		

The cerrado habitat had higher species richness and abundance, and individuals were taller and possessed greater diameters when compared to individuals of the grassland habitat (Tab. 3, Tab. S1 in supplementary material). Mean plant abundance in the cerrado habitat (106 \pm 9) was greater than that of the grassland (26 \pm 6) (Tab. 3). Mean plant height was higher in the cerrado (118.29 \pm 6.03cm) than in the grassland habitat (59.97 \pm 9.83cm) (Tab. 3). Relative density and soil coverage were the only vegetation structural parameters that did not differ between habitats (Tab. 3). In the cerrado habitat,78% of all the species were exclusive (51 out of 65), while in the grassland only30% of the species were exclusive (6 out of 20).

A total of 1349 individuals of 85 species, distributed among 24 families, were recorded; 1141 individuals of 65 species in the cerrado habitat, and 208 individuals of 20 species in the grassland habitat. Three species could not be identified. The families with greatest species richness in the cerrado habitat were Myrtaceae (11 species), Asteraceae (nine), Fabaceae and Malpighiaceae (seven), and Melastomataceae (six), which together represented 61% (40 out of 65 species) of all the species in this habitat. In the grassland habitat, the families with greatest species richness were Asteraceae (four species), Fabaceae (four), Melastomataceae (four), and Eriocaulaceae (two), accounting for 70% (14 out of 20) of all the species recorded (detailed in Tab. S1 in supplementary material).

In the cerrado habitat, *Chamaecrista papillata* (80.36), *Myrsine* (29.30), *Mimosa maguirei* (14.37), *Handroanthus ochraceus* (14.09) and *Eugenia* sp. (13.55) had the highest IVIs. The most abundant species was *M. guianensis* (197 individuals), followed by *Chamaecrista papillata* (102) and

Table 2. Soil properties effect in the indicator species abundance between two rupestrian habitats (rupestrian cerrado and Grassland) in Serra do Cipó. OM: organic matter, T: pH 7.0 cation-exchange capacity $cmol_c/dm^3$, m: Aluminum saturation index %.

	Species	Soil variables								
Habitat		OM	рН	k	T	Al	P	m	Zn	Cu
"Cerrado"	Handroanthus ochraceus	-	-	+	+	+				
	Mimosa radula						-			
	Myrsine guianensis				+	+		-		
	Erythroxylum campestre		-			+	-			
	Eugenia sp.	-	-	+	+				-	
	Ruellia villosa	-	-	+		+	-			-
	Mimosa maguirei		-	+			-			+
	Eugenia punicifolia	-								
	Chamaecrista papillata	-		+						
	Vernonanthura polyanthes		-		-					
	Diospyros hispida			+						
	Roupala montana		-			+		-	-	
	Byrsonima variabilis			+						
	Kielmeyera petiolaris		+	+			-	-		
	Psidium firmun					+				
Grassland	Cambessedesia hilariana						+		+	
	Richterago polymorpha		-			-	+		-	
	Actinocephalus bongardii	+				-				
	Paepalanthus argenteus			-	+	-		+	-	

M. maguirei (94). In the grassland habitat, Paepalanthus argenteus (61.34), Richterago polymorpha (50.32), Cambessedesia hilariana (29.13), Actinocephalus bongardii (23.51), and M. guianensis (21.15) had the highest IVIs. The most abundant species in this habitat were R. polymorpha (53), P. argenteus (44), and Microlicia crenulata (24).

Table 3. Vegetation structural variations (Mean \pm SD) between two rupestrian habitats (rupestrian cerrado and Grassland) in Serra do Cipó, Brazil. * indicates significant values (p<0.05).

Attribute	"Cerr	ado"	Grassland		
Attribute	μ		μ		
Abundance*	106.88	± 9.10	26.18	± 6.29	
Richness*	22.66	± 2.18	7.36	± 2.12	
Relative density	9.36	± 0.79	10.52	± 1.65	
Plant height (cm)*	118.29	± 6.03	59.97	± 9.83	
Stem diameter (cm)*	0.130	± 0.01	0.05	± 0.01	
Soil coverage	83.22	± 4.84	83.0	± 3.81	

Discussion

Our results demonstrate that the structure of the arboreal-shrub vegetation and the species composition of two contiguous habitats in the mosaic of two rupestrian environments in Serra do Cipó are shaped by microenvironmental variation in soil chemistry. We found different species composition, greater richness and abundance, and plants with greater diameters and height sin the cerrado, where the soil is chemically more complex. This result reinforces that micro-environmental variation of the soil acts as important drivers of vegetation in rupestrian ecosystems (Giulietti et al. 1997; Rapini et al. 2008; Fernandes 2016; Coelho et al. 2016; Schaefer et al. 2016). Furthermore, the results also indicate that even in contiguous areas, a rupestrian ecosystem can possess high environmental heterogeneity, which favors the formation of microhabitat mosaics and thus high flora richness and several plant life-form strategies (e.g., Negreiros et al. 2014; Stradic et al. 2015; Neves et al. 2016; Mota et al. 2018).

We demonstrated that micro-environmental variation in soil chemistry represents strong filters that favor certain plant life forms. Our results show that the predominant life forms of rupestrian cerrado habitat were woody shrubs, sub-shrubs, and small palm trees. This was also where the soil had higher concentrations of chemical compounds and nutrient levels, when compared to the soil of the grassland habitat. However, small species, predominantly herbs, were dominant in the grassland habitat, where the soil had less nutrition. Plant life forms reflect a set of characteristics that determine the establishment and survival of plants in a given environment. As a result, the dominant life forms in a given community determine its vegetation structure (Araújo et al. 2011; Messias et al. 2011).

In rupestrian grassland communities, slight environmental differences can result in large differences in functional traits of plants, ranking the soil, characterized by its low fertility, high Al concentration and low water retention, as the most important environmental filter for the presence of highly stress-tolerant species (Negreiros et al. 2014). Some predominant life form adaptations, such as woody species with the capacity for annual regrowth and/ or with subterranean structures for avoiding desiccation due to fire, are commonly found in rupestrian grasslands on rocky outcrops (Conceição et al. 2007). Therefore, plant life forms can be indicative of the environmental adaptation of a species in a given community, and indirectly provide information about the environment (Araújo et al. 2011; Schaefer et al. 2012). This pattern can be found in other environments where soil nutrients are scarce and where there is strong climatic pressures on plants, such as semi-arid and savanna areas of Brazil (Araújo et al. 2011; Neves et al. 2016; Mota et al. 2018). In such areas, there is pronounced environmental heterogeneity, which may influence species diversity, but mainly increases plant functional diversity (Araújo et al. 2011).

Nutrient availability and acidity, and species richness, were greater in the cerrado than in the grassland habitat. Acidity enhances soil solubility and carbonate, sulfate and phosphate concentrations (Mengel & Kirkby 1982; Bakker et al. 2005; Robins et al. 2013), which may explain the greater plant species richness of the cerrado habitat when compared to the grassland. Plants in the cerrado habitat were taller and had larger diameters, which can be favored by characteristics of the soil. Epigeal biomass formation by species in ferruginous rupestrian grasslands is closely related to aspects of the soil, such as water retention capacity, soil friability and granulometry (Teixeira & Lemos-Filho 2002). Additionally, basal area and vegetation structure of cerrado species are intimately related to soil characteristics, including the presence of clay and particle size, which enable greater physical support for roots (Assis et al. 2011).

Differences in the occurrence of species between the two studied habitats was largely associated with the higher acidity, Fe, Cu, Al, t and Mn of the cerrado, while in the grassland higher concentrations of Fe, OM, P and m had greater impact on species occurrence. In cerrado habitats, low basal saturation levels and restricted nutrient availability represent major challenges to plant nutrition (Arruda et al. 2017). Furthermore, cerrado soils differ in texture, acidity, Al, Ca and Mg concentrations (Mota et al. 2018). According to Oliveira et al. (2015), the low level of phosphorus in rupestrian grassland soils is an important environmental filter that favors root specialization strategies and greater leaf N:P proportions. Other factors, such as low resilience and low water retention capacity (Klink & Machado 2005; Negreiros et al. 2011), can also explain the heterogeneity of neighboring areas in rupestrian environments, which favor completely different vegetation types (Fernandes 2016).

In the grassland habitat, the establishment and abundance of some species, such as *C. hilariana*, were favored by the presence of P and Zn. Despite records of antagonist

effects of P and Zn (Olsen 1972; Reis-Junior & Martinez 2002), their interaction can also result in increased Zn absorption facilitated by the presence of P (Wallace *et al.* 1978). In fact, higher soil iron (Fe) concentration, as our results indicate, is associated with low pH levels, which could facilitate the release of P in this habitat (Hinsinger 2001).

The species abundance of the cerrado habitat was five times greater than that of the grassland habitat, which can be explained by the nutrient concentration and shallow soils of the latter, which are impoverished due to the quartzite and sandstone content of the parental rock (Benites *et al.* 2003). Furthermore, such conditions, combined with regimes of frequent fire, association with the mycorrhizal community and topography, can modify the shrub-tree vegetation structure, and thus reduce biomass, increase mortality rates and reduce species recruitment rates (Hoffmann & Moreira 2002; Miranda *et al.* 2004; Carvalho *et al.* 2014). Additionally, differences in nutrient availability in the soil, such as Ca², may result indifferent floristic compositions of cerrado vegetation (Haridasan 2008).

Several studies in the cerrado have reported Fabaceae to be the most representative family (Giulietti *et al.* 2005; Pinto *et al.* 2009; Cheung *et al.* 2016; Torres *et al.* 2017). Likewise, Myrtaceae has been recorded as the richest family in cerrado habitats in Goiás, in central Brazil (Moura *et al.* 2010). For the grassland habitat, Asteraceae was the richest family, which is in concordance with other studies in the region (e.g. Mota *et al.* 2016; see also Conceição *et al.* 2016; Silveira *et al.* 2016; Mota *et al.* 2018).

In general, the results shown here corroborate the hypothesis that micro-environmental variation in soil chemistry is directly related to vegetation structure and influence species composition of plants in rupestrian environments. Furthermore, such soil heterogeneity yields different plant life form strategies for maximizing the acquisition of nutrients. Plants in rupestrian ecosystems are known for being highly adapted in order to survive, reproduce and maintain populations in such heterogeneous mosaic landscapes, and to deal with the constraints that this harsh environment presents.

Acknowledgements

We thank A. Picorelli and E. Mota for fieldwork assistance, A. B. Viana-Junior for statistical support and D. Negreiros for the species identification. This study was supported by CNPq/Long-Term Ecological Research (PELD-CRSC-17), and also received support by ComCerrado/PPBio, CAPES, and Vale/FAPEMIG. We are grateful to Cedro Têxtil, Reserva Vellozia, Parque Nacional da Serra do Cipó, Pousada Serra Morena, and Pousada Pouso do Elefante for logistical support. TOB received a scholarship from CAPES/UFMG and IGZ from CNPq/DTI-C.

References

- Amorim PK, Batalha MA. 2006. Características edáficas de um cerrado hiperestacional em comparação com um cerrado estacional e um campo úmido: implicações para a estrutura da comunidade vegetal. Brazilian Journal of Biology 66: 661-670.
- APG IV. 2016. An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG IV. Botanical Journal of the Linnean Society 181: 1-20.
- Araújo FS, Costa RC, Lima JR, et al. 2011. Floristics and life-forms along a topographic gradient, central-western Ceará, Brazil. Rodriguésia 62: 341-366.
- Arruda DM, Fernandes-Filho EI, Solar RR, Schaefer CE. 2017. Combining climatic and soil properties better predicts covers of Brazilian biomes. The Science of Nature 104: 32.
- Assis ACC, Coelho RM, Pinheiro ES, Durigan G. 2011. Water availability determines physiognomic gradient in an area of low-fertility soils under Cerrado vegetation. Plant Ecology 212: 1135-1147.
- Bakker C, Rodenburg J, Bodegom PM. 2005. Effects of Ca-and Fe-rich seepage on P availability and plant performance in calcareous dune soils. Plant and Soil 275: 111-122.
- Barbosa NPU, Fernandes GW. 2016. Rupestrian grassland: past, present and future distribution. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 531-544.
- Barbosa NPU, Fernandes GW, Sanchez-Azofeifa A. 2015. A relict species restricted to a quartzitic mountain in tropical America: an example of micro refugium?. Acta Botanica Brasilica 29: 299-309.
- Bartón K. 2012. MuMIn: Multi-model inference. http://CRAN.R-project.org/. 30 Aug. 2017.
- Benites VM, Caiafa AN, Mendonça ES, Schaefer CE, Ker JC. 2003. Solos e vegetação nos complexos rupestres de altitude da Mantiqueira e do Espinhaço. Floresta e Ambiente 10: 76-87.
- Benites VM, Schaefer CEG, Simas FN, Santos HG. 2007. Soils associated with rock outcrops in the Brazilian mountain ranges Mantiqueira and Espinhaço. Brazilian Journal of Botany 30: 569-577.
- Burnham KP, Anderson DR. 2002. Model selection and multimodel inference: a practical information-theoretic approach. 2nd. edn. New York, Springer Science & Business Media.
- Carvalho F, Godoy EL, Lisboa FJ, et al. 2014. Relationship between physical and chemical soil attributes and plant species diversity in tropical mountain ecosystems from Brazil. Journal of Mountain Science 11: 875-883
- Carvalho F, Souza FA, Carrenho R, Souza FM, Jesus EC, Fernandes GW. 2012. The mosaic of habitats in the high-altitude Brazilian rupestrian fields is a hotspot for arbuscular mycorrhizal fungi. Applied Soil Ecology 52: 9-19.
- Cheung KC, Reis LK, Jesus CCC. 2016. Análise fitossociológica de um fragmento de Cerrado em Campo Grande, MS. Multitemas 21: 241-246.
- Clarke KR, Green RH. 1988. Statistical design and analysis for a biological effects study. Marine Ecology Progress Series 46: 213-226.
- Coelho MS, Fernandes GW, Pacheco P, et al. 2016. Archipelago of montane forests surrounded by rupestrian grasslands: new insights and perspectives. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 129-156.
- Conceição AA, Pirani JR, Meirelles ST. 2007. Floristics, structure and soil of insular vegetation in four quartzite-sandstone outcrops of "Chapada Diamantina", Northeast Brazil. Revista Brasileira de Botânica 30: 641-656.
- Conceição AA, Rapini A, Carmo FF, *et al.* 2016. Rupestrian grassland vegetation, diversity, and origin. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 105-123.
- Coutinho ES, Fernandes GW, Berbara RLL, Valério HM, Goto BT. 2015. Variation of arbuscular mycorrhizal fungal communities along an altitudinal gradient in rupestrian grasslands in Brazil. Mycorrhiza 25: 627-638.



Hernani Alves Almeida, Thaíse de Oliveira Bahia, Irene Gélvez-Zúñiga and Geraldo Wilson Fernandes

- Crawley MJ. 2007. The R book. 2nd. edn. New York, John Wiley and Sons. Fernandes GW. 2016. The megadiverse rupestrian grassland. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 3-14.
- Fernandes GW, Almeida HA, Nunes CA, et al. 2016. Cerrado to rupestrian grasslands: patterns of species distribution and the forces shaping them along an altitudinal gradient. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 345-377.
- Fernandes GW, Price PW. 1991. Comparisons of tropical and temperate galling species richness: the role of environmental harshness and plant nutritional status. In: Price PW, Lewinsohn TM, Fernandes GW, Benson WW. (eds.) Plant-animal interactions: evolutionary ecology in tropical and temperate regions. New York, Wiley. p. 91-115.
- Figueira JEC, Ribeiro KT, Ribeiro MC, *et al.* 2016. Fire in rupestrian grasslands: plant response and management. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 415-448.
- Flora do Brasil 2020 em construção. 2017. Flora do Brasil 2020 Algas, fungos e plantas Jardim Botânico do Rio de Janeiro. http://floradobrasil.jbrj.gov.br/. 08 Jan. 2017.
- Giulietti AM, Harley RM, Queiroz LP, Wanderley MG, Berg C. 2005. Biodiversidade e conservação das plantas no Brasil. Megadiversidade 1:52-61
- Giulietti AM, Menezes NL, Pirani JR, Meguro M, Wanderley MGL. 1987. Flora da Serra do Cipó, Minas Gerais: caracterização e lista de espécies. Boletim de Botânica 9: 1-151.
- Giulietti AM, Pirani J, Harley RM. 1997. Espinhaço range region. Eastern Brazil. In: Davis SD, Heywood VH, Herrera-MacBryde O, Villa-Lobos O, Hamilton AC. (eds.) Centres of plant diversity. A guide and strategies for the conservation. Cambridge, IUCN Publication Unity. p. 397-404.
- Hammer Ø, Harper DAT, Ryan PD. 2001. PAST: Paleontological statistics software package for education and data analysis. Palaeontologia Electronica 4. http://palaeo-electronica.org/2001_1/past/issue1_01. htm. 30 Sep. 2017.
- Haridasan M. 2008. Nutritional adaptations of native plants of the cerrado biome in acid soils. Brazilian Journal of Plant Physiology 20: 183-195.
- Hinsinger P. 2001. Bioavailability of soil inorganic P in the rhizosphere as affected by root-induced chemical changes: a review. Plant and Soil 237: 173-95.
- Hoffmann WA, Moreira AG. 2002. The role of fire in population dynamics of woody plants. In: Oliveira PS, Marquis RJ. (eds.) The Cerrados of Brazil: ecology and natural history of a neotropical savanna. New York, Columbia University Press. p. 159-177.
- Hopper SD. 2009. OCBIL theory: towards an integrated understanding of the evolution, ecology and conservation of biodiversity on old, climatically-buffered, infertile landscapes. Plant and Soil 322: 49-86.
- Klink CA, Machado RB. 2005. Conservation of the Brazilian cerrado. Conservation Biology 19: 707-713.
- Kuster VC, Castro SAB, Vale FHA. 2016. Morphological and physiological responses of three plant species occurring in distinct altitudes in the Neotropical savannah. Brazilian Journal of Botany 39: 1039-1049.
- Lenza E, Abadia AC, Menegat H, *et al.* 2017. Does fire determine distinct floristic composition of two Cerrado savanna communities on different substrates? Acta Botanica Brasilica 31: 250-259.
- Mengel K, Kirkby EA. 1982. Principles of plant nutrition. Bern, International Potash Institute.
- Messias MCTB, Leite MGP, Meira-Neto JAA, Kozovits AR. 2011. Life-form spectra of quartzite and itabirite rocky outcrops sites, Minas Gerais, Brazil. Biota Neotropica 11: 255-268.
- Messias MCTB, Leite MGP, Meira-Neto JAA, Kozovits AR, Tavares R. 2013. Soil-vegetation relationship in quartzite and ferruginous Brazilian rocky outcrops. Folia Geobotanica 48: 509-521.
- Mews HA, Pinto JRR, Eisenlohr PV, Lenza E. 2016. No evidence of intrinsic spatial processes driving Neotropical savanna vegetation on different substrates. Biotropica 48: 433-442.

- Miranda S, Sato MN, Andrade SMDA, Haridasan M, Morais HC. 2004. Queimadas de Cerrado: caracterização e impactos. In: Aguiar LMS, Camargo AJA. (eds.) Cerrado: ecologia e caracterização. Planaltina, Embrapa Cerrados. p. 69-123.
- Mota GDS, Luz GR, Mota NM, et al. 2018. Changes in species composition, vegetation structure, and life forms along an altitudinal gradient of rupestrian grasslands in south-eastern Brazil. Flora 238: 32-42.
- Mota NM, Rezende VL, Mota GDS, Fernandes GW, Nunes YRF. 2016. Forces driving the regeneration component of a rupestrian grassland complex along an altitudinal gradient. Brazilian Journal of Botany 39: 845-860.
- Moura IIOD, Gomes-Klein VL, Maria Felfili J, Ferreira HD. 2010. Diversidade e estrutura comunitária de cerrado sensu stricto em afloramentos rochosos no Parque Estadual dos Pireneus, Goiás. Brazilian Journal of Botany 33: 455-467.
- $\label{eq:Mueller-Dombois D, Ellenberg H. 1974. Aims and methods of vegetation ecology. New York, John Wiley \& Sons.$
- Negreiros D, Fernandes GW, Berbara RLL, Rodarte LHO, Barbosa NPU. 2011. Caracterização físico-química de solos quartzíticos degradados e áreas adjacentes de campo rupestre na Serra do Cipó, MG, Brasil. Neotropical Biology and Conservation 6: 156-161.
- Negreiros D, Moraes MLM, Fernandes GW. 2008. Caracterização da fertilidade dos solos de quatro leguminosas de campos rupestres, Serra do Cipó, MG, Brasil. Revista de la Ciencia del Suelo y Nutricion Vegetal 8: 30-39.
- Negreiros D, Stradic S, Fernandes GW, Rennó HC. 2014. CSR analysis of plant functional types in highly diverse tropical grasslands of harsh environments. Plant Ecology 215: 379-388.
- Neves SPS, Funch R, Conceição AA, Miranda LAP, Funch LS. 2016. What are the most important factors determining different vegetation types in the Chapada Diamantina, Brazil?. Brazilian Journal of Biology 76: 315-333.
- Nuland ME, Wooliver RC, Pfennigwerth AA, et al. 2016. Plant-soil feedback: connecting ecosystem ecology and evolution. Functional Ecology 30: 1032-1042.
- Oliveira RS, Galvão HC, Campos MCR, Eller CB, Pearse SJ, Lambers H. 2015. Mineral nutrition of *campos rupestres* plant species on contrasting nutriente-impoverished soil types. New Phytologist 205: 1183-1194.
- Olsen SR. 1972. Micronutrient interaction. In: Mortvedt JJ, Giordano PM, Lindsay WL. (eds.) Micronutrients in agriculture. Madison, Soil Science Society of America. p. 243-264.
- Pausas JG, Austin MP. 2001. Patterns of plant species richness in relation to different environments: an appraisal. Journal of Vegetation Science 12: 153-166.
- Pinto JRR, Oliveira EL, Pinto ADS. 2009. Composição florística e estrutura da vegetação arbustivo-arbórea em um cerrado rupestre, Cocalzinho de Goiás, Goiás. Revista Brasileira de Botânica 32: 23-32.
- R Core Team. 2014. R: A language and environment for statistical computing. R Foundation for Statistical Computing. Vienna, Austria.
- Rapini A, Ribeiro PL, Lambert S, Pirani JR. 2008. A flora dos campos rupestres da Cadeia do Espinhaço. Megadiversidade 4: 16-24.
- Reis-Junior RA, Martinez HEP. 2002. Adição de Zn e absorção, translocação e utilização de Zn e P por cultivares de cafeeiro. Scientia Agricola 59: 537-542.
- Robins NS, Pye K, Wallace H. 2013. Dynamic coastal dune spit: the impact of morphological change on dune slacks at Whiteford Burrows, South Wales, UK. Journal of Coastal Conservation 17: 473-482.
- Rocha NMWB, Carstensen DW, Fernandes GW, Buisson E, Morellato LPC. 2016. Phenology patterns across a rupestrian grassland altitudinal gradient. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 275-289.
- Schaefer CEGR, Cândido HG, Corrêa GR, Nunes JA, Arruda DM. 2016. Soils associated with rupestrian grasslands. In: Fernandes GW. (ed.) Ecology and conservation of mountaintop grasslands in Brazil. Cham, Springer. p. 55-69.
- Schaefer CEGR, Mendonça BAF, Ferreira-Junior WG, Valente EL, Correa GR. 2012. Relações solo-vegetação em alguns ambientes



Together yet separate: variation in soil chemistry determines differences in the arboreal-shrub structure of two contiguous rupestrian environments

- brasileiros: fatores edáficos e florística. In: Martins SV. (ed.) Ecologia de florestas tropicais do Brasil. Viçosa, Editora UFV. p. 253-293.
- Silva AC. 2005. Solos. In: Silva AC, Pedreira LCVS, Almeida-Abreu PA. (eds.) Serra do Espinhaço Meridional: paisagens e ambientes. Belo Horizonte, O Lutador. p.59-78.
- Silveira FA, Negreiros D, Barbosa NP, et al. 2016. Ecology and evolution of plant diversity in the endangered campo rupestre: a neglected conservation priority. Plant and Soil 403: 129-152.
- Sólymos P. 2009. Processing ecological data in R with the meta Package. Journal of Statiscal Software 29: 1-28.
- Stradic S, Buisson E, Fernandes GW. 2015. Vegetation composition and structure of some neotropical mountain grasslands in Brazil. Journal of Mountain Science 12: 864-877.
- Teixeira WA, Lemos-Filho JP. 2002. Fatores edáficos e a colonização de espécies lenhosas em uma cava de mineração de ferro em Itabirito, Minas Gerais. Revista Árvore 26: 25-33.
- Torres DM, Fontes MAL, Samsonas HDP. 2017. Soil-vegetation relationships in structuring cerrado *sensu stricto* communities in southern Minas Gerais, Brazil. Rodriguésia 68: 115-128.
- Vincent RC, Meguro M. 2008. Influence of soil properties on the abundance of plant species in ferruginous rocky soils vegetation, southeastern Brazil. Revista Brasileira de Botânica 31: 377-388.
- Wallace A, Muller RT, Alexander GV. 1978. Influence of phosphorus on zinc, iron, manganese and copper uptake by plants. Soil Science 126: 336-341.

