

Composition and structure of fish assemblage from Passa Cinco stream, Corumbataí river sub-basin, SP, Brazil

Carmassi, AL.^{a*}, Rondineli, G.^b, Ferreira, FC.^c and Braga, FMS.^a

^aDepartamento de Zoologia, Instituto de Biociências, Universidade Estadual Paulista – UNESP, Av. 24A, 1515, Bela Vista, CP 199, CEP 13506-900, Rio Claro, SP, Brazil

^bDepartamento de Produção Vegetal, Centro de Ciências Agrárias, Universidade Federal do Espírito Santo – UFES, Alto Universitário, s/n, Guararema, CP 16, CEP 29500-000, Alegre, ES, Brazil

^cDepartamento de Ecologia, Instituto de Biociências, Universidade Estadual Paulista – UNESP, Av. 24A, 1515, Bela Vista, CP 199, CEP 13506-900, Rio Claro, SP, Brazil

*e-mail: alberto.carmassi@gmail.com

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(With 10 figures)

Abstract

The aim of this work was to determine the composition of the fish assemblage of Passa Cinco stream and verify changes in their structure on the altitudinal gradient. Six samples were performed at five different sites in Passa Cinco stream (from the headwater, at order two, to its mouth, at order six), using an electric fishery equipment and gill nets in May, July, September and November of 2005 and January and March of 2006. The indices of Shannon's diversity, Pielou's evenness and Margalef's richness were quantified separately considering the different fishery equipment (nets versus electric fishery equipment). An ANOVA was used to compare samples collected in relation to values of abundance, diversity, evenness and richness. The representativeness of the species was summarised by their average values of abundance and weight. We captured 5082 individuals distributed into 61 species. We observed a trend of increasing diversity, richness and evenness of species from site 1 to 3, with further decrease in sites 4 and 5. The values found for habitat diversity also followed this pattern. Significant differences were found for all three indices considering the electric fishery samples. For individuals caught with nets, only the richness index showed a significant difference. *Characidium* aff. *zebra* was an important species in the headwater and transition sites and *Hypostomus strigaticeps* in middle-lower course sites. Despite the small extension of the Passa Cinco stream, environments structurally well defined were evidenced by the species distribution and assemblage composition along the gradient.

Keywords: ichthyofauna, species diversity, altitudinal gradient, river continuum concept.

Composição e estrutura da assembleia de peixes do Rio Passa Cinco, sub-bacia do rio Corumbataí, São Paulo, Brasil

Resumo

O objetivo desse trabalho foi determinar a composição da assembleia de peixes do Rio Passa Cinco, além de verificar alterações em sua estrutura em função do gradiente altitudinal. Foram realizadas seis amostragens em cinco pontos distintos no Rio Passa Cinco (contemplando trechos de ordem dois, três, quatro, cinco e seis), fazendo uso de pesca elétrica e redes de espera, nos meses de maio, julho, setembro e novembro de 2005 e janeiro e março de 2006. Os índices de diversidade de Shannon, equidade de Pielou e riqueza de Margalef foram quantificados separadamente, considerando as diferentes artes de pesca (redes versus pesca elétrica). Foi utilizada a ANOVA para comparar os pontos de coleta quanto aos valores médios de abundância, diversidade, equidade e riqueza. A representatividade das espécies foi resumida pelos seus valores médios de abundância e peso. Foram capturados 5082 indivíduos distribuídos em 61 espécies. Foi observada uma tendência de aumento de diversidade, riqueza e equidade das espécies do ponto 1 ao 3, com posterior queda nos pontos 4 e 5, indo ao encontro com os valores de diversidade de habitat, que também seguiram esse padrão. Foram encontradas diferenças significativas para os três índices considerando as amostras de pesca elétrica. Para os indivíduos capturados com redes, apenas o índice de riqueza apresentou diferença significativa. *Characidium* aff. *zebra* foi uma espécie importante nos trechos de cabeceira e transição, e *Hypostomus strigaticeps* nos trechos de médio-baixo curso. Apesar da pequena extensão do Rio Passa Cinco, ambientes estruturalmente bem definidos foram evidenciados pela distribuição das espécies e pela composição da assembleia ao longo do gradiente.

Palavras-chave: ictiofauna, diversidade de espécies, gradiente altitudinal, conceito do rio contínuo.

1. Introduction

Fish distribution in an environment is rarely caused by a single factor. Changes in the fish species composition from the headwaters to the lower parts is a common phenomenon and conceptual models based on temperate rivers seek to explain the mechanisms responsible for these processes (Matthews, 1998). Geomorphology is an important factor affecting the structure of fish communities in lotic environments (Allan, 1997), because from the headwaters to the mouth, the river goes through different terrain features, leading to changes in limnological characteristics and structural environment.

According to the river continuum concept (Vannote et al., 1980), we expect a gradual increase in species richness along the gradient, and the area of the middle section as the most diverse. These changes usually are associated with habitat changes along the gradient (Gorman and Karr, 1978). Matthews and Styron (1981) also suggested that the physical and chemical conditions in the headwaters are more stressful than in the lower portions, so that few fish species can colonize these areas.

Spatial patterns of species distribution have been addressed in several studies (Uieda, 1984; Garutti, 1988; Uieda and Barreto, 1999; Silvano et al., 2000; Mazzoni and Lobón-Cerviá, 2000; Uieda and Uieda, 2001; Pavanelli and Caramaschi, 2003; Castro et al., 2003; Suárez and Petreire, 2003; Suárez and Petreire, 2005; Castro et al., 2004; Casatti, 2005; Braga and Andrade, 2005; Cetra and Petreire, 2006) that, in general, reported an increase in species diversity since the headwaters to the mouth in fish communities at different locations. The relationship between habitat and community composition was presented by Montag et al. (1997), Benedito-Cecilio et al. (2004) and Gomiero and Braga (2006), which showed the correlations between habitat diversity and general conditions in the environment.

Among the main patterns of longitudinal variation in stream fishes, there are the species additions and replacements (Gilliam et al., 1993; Petry and Schultz, 2006). Species additions are generally correlated with less severe environmental gradients, leading to smooth changes in abiotic and/or structural factors, while the species replacements occur as a result of abrupt changes in stream geomorphology or are related to abiotic conditions (Balon and Stewart, 1983; Winemiller and Leslie, 1992; Edds, 1993; Jackson et al., 2001; Wilkison and Edds, 2001; Ferreira and Petreire, 2009).

The Passa Cinco stream is one of the main rivers of the Corumbataí river sub-basin, that belongs to the Piracicaba river basin, one of the last options to provide good water quality to several municipalities. Situated next to major urban, agricultural, technological and scientific centres of southeastern Brazil, it has been degraded over a century, through soil use and occupancy, and by excessive withdrawal of water for human consumption and agriculture. Given the regional importance of these water bodies, the aim of this study was to determine the composition of the fish assemblages in Passa Cinco stream and measure the changes in their structure along the longitudinal gradient.

2. Material and Methods

This study was accomplished in the Passa Cinco stream, with headwaters localised in the Serra da Cachoeira a component of the complex of Serra de Itaqueri, in the municipal district of Itirapina. The drainage area of Passa Cinco stream is 525 km², covering about 60 km from its headwaters (with about 1000 m) to its confluence with the Corumbataí river (to 480 m) (Garcia et al., 2004). Currently, it has 51.72% of its area occupied by pastures, 14.13% by sugar cane, 15.67% by native forest and 0.74% by savanna (Valente and Vettorazzi, 2002).

Six samplings were accomplished at five different sites in Passa Cinco stream, contemplating sections of order two, three, four, five and six, according to the Strahler (1952) classification, with the following geographic coordinates: site 1 (order 2) - 22° 23' 36" S and 47° 53' 08" W, site 2 (order 3) - 22° 22' 10" S and 47° 51' 22" W, site 3 (order 4) - 22° 21' 63" S and 47° 48' 48" W, site 4 (order 5) - 22° 24' 74" S and 47° 43' 34" W and site 5 (order 6) - 22° 30' 97" S and 47° 39' 49" W in the months of May, July, September and November of 2005 and January and March of 2006 (Figures 1 and 2).

In each sample site, the predominant substrate type was recorded, as well as presence/absence of riparian vegetation, degree of shading, type of current and mean depth, as shown by Rondineli and Braga (2009). From these data, the habitat heterogeneity at each site was estimated using the Shannon diversity index (Gorman and Karr, 1978).

The fishery equipment used was electric fishery and gill nets. The electric fishery equipment (which consists of a generator that provides power - 110 V - for a rectifier current that has the capacity to increase the voltage - up to 1500 V - and reduce the amperage - to 2 A) was used in the first three sites (1, 2 and 3). In these sites we performed a downstream-upstream pass at 50 m stretches without using contention nets. Gill nets (with mesh sizes varying from 3 to 9 cm between adjacent knots) were used in sites 3, 4 and 5. The sequence of gill nets was determined at random and placed in the afternoon (between 15 and 18 hours) remaining until the following morning. After each collection, fish were put into plastic bags, fixed in 10% formalin for 2 days and then transferred to 70% ethanol until the analysis was accomplished. In the laboratory, the fishes were identified to species level, measured for total length (cm) and weighed (g). Voucher specimens were deposited in the Ichthyology Laboratory, Department of Zoology of the Universidade Estadual Paulista, in Rio Claro.

The indexes of Shannon's diversity, Pielou's evenness and Margalef's richness were calculated for each sample site, separating the individuals captured by electric fishery from those captured by gill nets. To determine mean differences, ANOVAs were performed for each index to compare the sites sampled with the same fishery equipment (P1, P2 and P3 - electric fishery and P3, P4 and P5 - gill nets).

The representativeness of the species were summarised by their mean values of abundance ($\bar{N} = N/F_0$) and weight ($\bar{P} = P/F_0$), where N is the number of individuals, P the

weight and F_o the frequency of occurrence (Ferreira, 2007). Thus, \bar{N} and \bar{P} measure the local importance of each species sampled in each order. The species most and least important were highlighted. This analysis was performed for each sample site, considering the different fishery equipment.

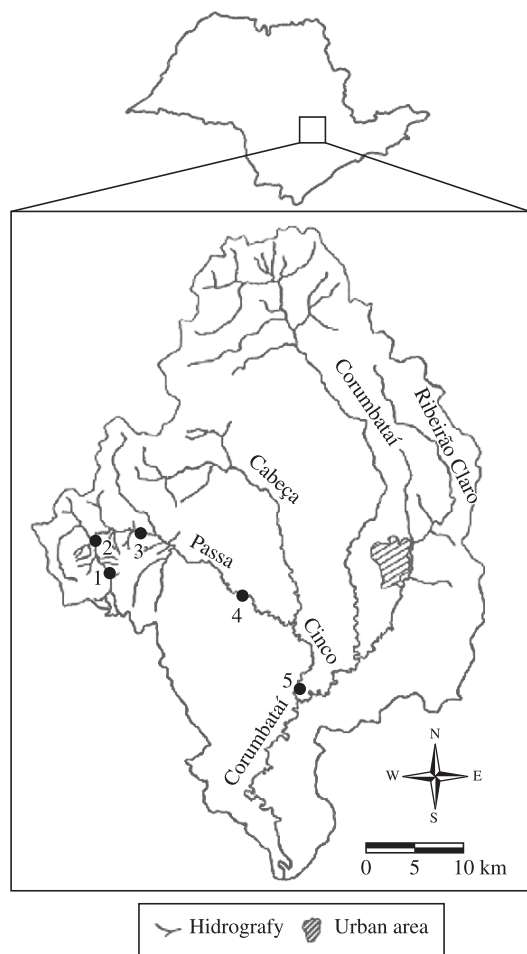


Figure 1. Hydrographic map of the Corumbataí river sub-basin, showing the sampled sites (modified from Garcia et al., 2004) 1) Sample site 1; 2) Sample site 2; 3) Sample site 3; 4) Sample site 4; 5) Sample site 5; Urban area – Rio Claro).

Table 1. Bottom type, marginal vegetation, degree of shading, current type and mean width found in each of the sampling sites of Passa Cinco stream.

| Sites | Bottom type | Marginal vegetation | Degree of shading | Current type | Mean width (m) |
|-------|--------------------|---------------------|-------------------|-----------------|----------------|
| 1 | pebble | absent | >76% | riffle | shallow |
| 2 | pebble gravel | present | 51-75% | riffle run | shallow |
| 3 | pebble gravel sand | present | 26-50% | riffle run pool | moderate |
| 4 | sand | present | 26-50% | run | Deep |
| 5 | sand | present | 0-25% | run | Deep |

3. Results

The substrate type changed gradually from site 1 to site 5, the marginal vegetation (presence of grasses along the streams bank) occurred in all sampling sites, except for site 1. The degree of shading decreased from the headwater to the mouth. The types of predominant current were: riffle in site 1, riffle and run in site 2, riffle, run and pool in site 3 and run in sites 4 and 5. The mean depth increased from upstream to downstream (Table 1 and Figure 2). Site 3 showed the highest diversity of habitats, followed by sites 5 and 4 (Figure 3).

We captured 5082 individuals belonging to 61 species. *Characidium* aff. *zebra* was the most abundant species (Table 2). The number of individuals, species richness and the indexes of Shannon's diversity, Margalef's richness and Pielou's evenness are shown in Table 3 for each site and fishing gear. The largest catches occurred in sites 3, 2 and 1 sampled with electric fishery. The greatest richness and diversity were obtained in site 3, regardless of the fishing gear used. The evenness index values for all sites were around 0.8.

For sites 1 to 3 (electric fishery), the ANOVA showed significant differences ($p < 0.001$, Table 4) in the indexes of richness, diversity and evenness. So we can accept that the Shannon diversity changed due to variation in either richness and evenness. For sites 3 to 5 (gill nets), only the richness index differed significantly ($p = 0.047$, Table 4).

Figure 4 presents the mean values, standard errors and 95% confidence intervals for the indexes of diversity, richness and evenness as well as for the number of individuals in the five sampling sites, considering separately the sites sampled with electric fishery (sites 1, 2 and 3) and gillnets (sites 3, 4 and 5). An increase in diversity, evenness and richness can be found from site 1 to site 3. Moreover, a decrease in richness was found from site 3 to site 5, with the lowest value found in site 4. Regarding the number of individuals, there is a greater abundance in places where electric fishery was used. However, the fishery equipment had more subtle effects on the indexes of diversity, evenness and richness, as shown in site 3, the only one in which it was possible to apply the two techniques.

When species composition was analysed in terms of mean weight and mean abundance per sample, it is possible to see which will best represent the sampled sites. When considering only the individuals captured with electric fishery *Characidium* aff. *zebra*, *Trichomycterus* sp.1, *Imparfinis mirini* and *Cetopsorhamdia iheringi*



Figure 2. View of the samples sites: a) site 1; b) site 2; c) site 3; d) site 4; e) site 5.

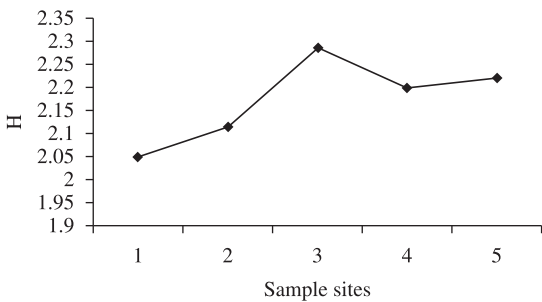


Figure 3. Values of Shannon diversity index for the habitat at each sample site.

were the most important species in site 1 (Figure 5); *Characidium* aff. *zebra*, *Trichomycterus* sp.1, *Bryconamericus stramineus* and *B. turiuba* in site 2 (Figure 6); and *C.* aff. *zebra*, *Hypostomus strigaticeps*, *Parodon nasus* and *Apareiodon ibitiensis* in site 3 (Figure 7). Whereas individuals caught with nets, *H. strigaticeps* was the most important species in site 3 (Figure 8); in site 4, *Odontostilbe* aff. *microcephala*, *Astyanax* sp.1, *A. altiparanae* and *H. strigaticeps* stood out (Figure 9); and in site 5, *H. strigaticeps* and *A. altiparanae* were the most important species (Figure 10).

Table 2. Fish species caught in the Passa Cinco stream for each sample site in decreasing order of abundance.

| Code | Species | Descriptor | P1 | P2 | P3 | P4 | P5 | Total |
|----------|---------------------------------------|--|-----|-----|-----|----|-----|-------|
| Czeb | <i>Characidium aff. zebra</i> | (Eigenmann, 1909) | 212 | 408 | 282 | 21 | 10 | 933 |
| Pcau | <i>Phalloceros harpagus</i> | Lucinda, 2008 | 202 | 275 | 99 | 0 | 3 | 579 |
| Astysp1 | <i>Astyanax</i> sp.1 | | 118 | 104 | 44 | 61 | 46 | 373 |
| Hstri | <i>Hypostomus strigaticeps</i> | (Regan, 1908) | 9 | 36 | 181 | 19 | 115 | 360 |
| Trichsp1 | <i>Trichomycterus</i> sp.1 | | 199 | 93 | 51 | 0 | 0 | 343 |
| Bstr | <i>Bryconamericus stramineus</i> | Eigenmann, 1908 | 4 | 129 | 44 | 90 | 1 | 268 |
| Omic | <i>Odontostilbe aff. microcephala</i> | Eigenmann, 1907 | 1 | 2 | 58 | 82 | 105 | 248 |
| Imir | <i>Imparfinis mirini</i> | Hasemann, 1911 | 110 | 69 | 47 | 1 | 4 | 231 |
| Btur | <i>Bryconamericus turiuba</i> | Langeani, de Lucena, Pedrini & Tarelho-Pereira, 2005 | 5 | 119 | 67 | 0 | 0 | 191 |
| Apar | <i>Astyanax scabripinnis paranae</i> | Eigenmann, 1914 | 83 | 97 | 6 | 0 | 0 | 186 |
| Cihe | <i>Cetopsorhamdia iheringi</i> | Schubart & Gomes, 1959 | 57 | 41 | 70 | 6 | 0 | 174 |
| Ccue | <i>Corumbataia cuestae</i> | Britski, 1997 | 2 | 112 | 39 | 5 | 0 | 158 |
| Aalt | <i>Astyanax altiparanae</i> | Garutti & Britskii, 2000 | 7 | 13 | 42 | 25 | 71 | 158 |
| Aibi | <i>Apareiodon ibitiensis</i> | Amaral Campos, 1944 | 1 | 38 | 71 | 0 | 0 | 110 |
| Pnas | <i>Parodon nasus</i> | Kner, 1859 | 0 | 9 | 86 | 1 | 2 | 98 |
| Aboc | <i>Astyanax bockmanni</i> | Vari & Castro, 2007 | 3 | 26 | 46 | 8 | 2 | 85 |
| Cfla | <i>Corydoras flaveolus</i> | Ihering, 1911 | 0 | 24 | 32 | 5 | 12 | 73 |
| Imbor | <i>Imparfinis borodini</i> | Mees & Cala, 1989 | 14 | 25 | 30 | 0 | 0 | 69 |
| Parg | <i>Piabina argentea</i> | Reinhardt, 1867 | 0 | 31 | 25 | 0 | 0 | 56 |
| Rque | <i>Rhamdia quelen</i> | (Quoy & Gaimard, 1824) | 15 | 15 | 15 | 3 | 6 | 54 |
| Hanc | <i>Hypostomus ancistroides</i> | (Ihering, 1911) | 6 | 4 | 12 | 7 | 20 | 49 |
| Hypsp1 | <i>Hypostomus</i> sp.1 | | 4 | 1 | 25 | 5 | 12 | 47 |
| Gbra | <i>Geophagus brasiliensis</i> | Kner, 1865 | 0 | 3 | 21 | 6 | 2 | 32 |
| Afas | <i>Astyanax fasciatus</i> | (Cuvier, 1819) | 0 | 0 | 11 | 0 | 13 | 24 |
| Aaff | <i>Apareiodon affinis</i> | (Steindachner, 1879) | 0 | 0 | 10 | 1 | 9 | 20 |
| Rlat | <i>Rineloricaria latirostis</i> | (Boulenger, 1900) | 1 | 3 | 7 | 2 | 5 | 18 |
| Hison | <i>Hisonotus</i> sp. | | 0 | 0 | 14 | 3 | 0 | 17 |
| Pimesp1 | <i>Pimelodella</i> sp.1 | | 0 | 0 | 1 | 0 | 12 | 13 |
| Ains | <i>Steidachnerina insculpta</i> | (Fernández-Yépez, 1948) | 0 | 0 | 1 | 1 | 10 | 12 |
| Hher | <i>Hypostomus cf. hermanni</i> | (Ihering, 1905) | 0 | 0 | 10 | 0 | 0 | 10 |
| Neop | <i>Neoplecostomus</i> sp. | | 7 | 0 | 3 | 0 | 0 | 10 |
| Microg | <i>Microglanis cf. garavelloi</i> | | 0 | 0 | 7 | 0 | 0 | 7 |
| Lfri | <i>Leporinus friderici</i> | (Bloch, 1794) | 0 | 0 | 1 | 0 | 5 | 6 |
| Gcar | <i>Gymnotus cf. carapo</i> | Linnaeus, 1758 | 0 | 0 | 6 | 0 | 0 | 6 |
| Hins | <i>Hisonotus insperatus</i> | Britski & Garavello, 2003 | 0 | 0 | 0 | 5 | 0 | 5 |
| Pret | <i>Poecilia reticulata</i> | Peters, 1859 | 0 | 0 | 2 | 3 | 0 | 5 |
| Evir | <i>Eigenmannia virescens</i> | (Valenciennes, 1836) | 0 | 2 | 0 | 3 | 0 | 5 |
| Isch | <i>Imparfinis schubarti</i> | (Gomes, 1956) | 0 | 0 | 3 | 1 | 0 | 4 |
| Lstr | <i>Leporinus striatus</i> | Kner, 1858 | 0 | 0 | 0 | 0 | 4 | 4 |
| Cgom | <i>Characidium gomesi</i> | Travassos, 1956 | 0 | 0 | 4 | 0 | 0 | 4 |
| Cmod | <i>Cyphocharax modestus</i> | (Fernández-Yépez, 1948) | 0 | 0 | 0 | 0 | 3 | 3 |
| Plin | <i>Prochilodus lineatus</i> | (Valenciennes, 1836) | 0 | 0 | 3 | 0 | 0 | 3 |
| Hequ | <i>Hypheobrycon eques</i> | (Steindachner, 1882) | 0 | 0 | 3 | 0 | 0 | 3 |
| Hypsp2 | <i>Hypostomus</i> sp.2 | | 0 | 0 | 0 | 1 | 2 | 3 |
| Bryc | <i>Bryconamericus</i> sp. | | 0 | 0 | 1 | 1 | 0 | 2 |

Table 2. Continued...

| Code | Species | Descriptor | P1 | P2 | P3 | P4 | P5 | Total |
|----------|----------------------------------|---------------------------------|------|------|------|-----|-----|-------|
| Hmal | <i>Hoplias cf. malabaricus</i> | (Bloch, 1794) | 0 | 0 | 0 | 1 | 1 | 2 |
| Cgob | <i>Cetopsis gobioides</i> | Kner, 1858 | 0 | 1 | 1 | 0 | 0 | 2 |
| Hypsp3 | <i>Hypostomus sp.3</i> | | 0 | 0 | 2 | 0 | 0 | 2 |
| Caen | <i>Corydoras aeneus</i> | (Gill, 1858) | 0 | 0 | 1 | 0 | 1 | 2 |
| Sspi | <i>Serrasalmus maculatus</i> | Kner, 1858 | 0 | 0 | 0 | 0 | 2 | 2 |
| Pbri | <i>Planaltina britski</i> | Menezes, Weitzman & Burns, 2003 | 0 | 0 | 0 | 0 | 2 | 2 |
| Loct | <i>Leporinus octofasciatus</i> | Steindachner, 1915 | 0 | 0 | 0 | 0 | 2 | 2 |
| Lpir | <i>Loricaria cf. piracicabae</i> | Ihering, 1907 | 0 | 0 | 0 | 0 | 1 | 1 |
| Cvan | <i>Cyphocharax vanderi</i> | (Britski, 1980) | 1 | 0 | 0 | 0 | 0 | 1 |
| Lvit | <i>Leporellus vittatus</i> | (Valenciennes, 1850) | 0 | 0 | 1 | 0 | 0 | 1 |
| Trichsp2 | <i>Trichomycterus sp.2</i> | | 0 | 0 | 0 | 0 | 1 | 1 |
| Cjag | <i>Crenicichla jaguarensis</i> | Haseman, 1911 | 0 | 0 | 0 | 0 | 1 | 1 |
| Hyph | <i>Hyphessobrycon sp.</i> | | 0 | 0 | 0 | 0 | 1 | 1 |
| Smar | <i>Synbranchus marmoratus</i> | Haseman, 1911 | 0 | 0 | 0 | 0 | 1 | 1 |
| Hlit | <i>Hoplosternum littorale</i> | (Hancock, 1828) | 0 | 0 | 0 | 0 | 1 | 1 |
| Aden | <i>Aphyocharax dentatus</i> | Eigenmann & Kennedy, 1903 | 0 | 0 | 0 | 1 | 0 | 1 |
| Total | | | 1061 | 1680 | 1485 | 368 | 488 | 5082 |

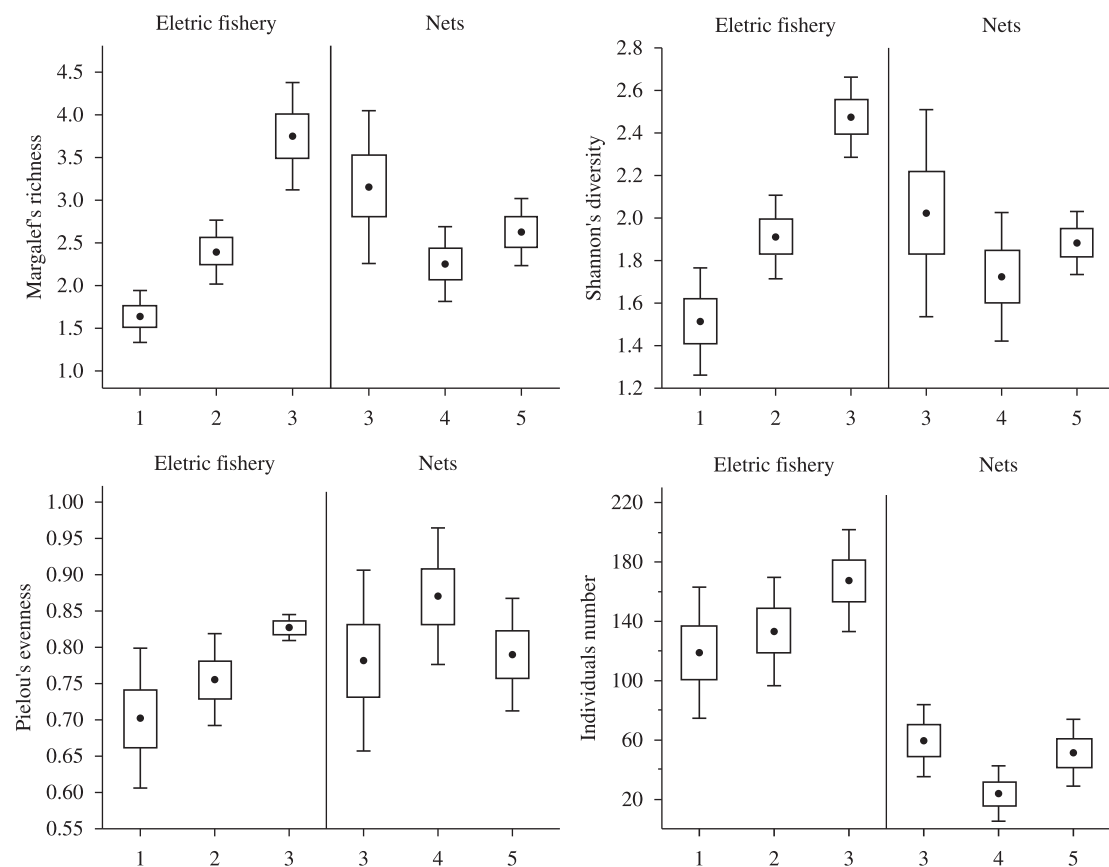


Figure 4. Mean values (●), standard error (□) and confidence interval 95% (bars) for the Shannon diversity index, Margalef's richness and Pielou's evenness and number of individuals, considering separately samples from electrofishing and nets.

Table 3. Number of individuals (N), richness (S) and indices of Shannon’s diversity (H), Margalef’s richness (MG) and Pielou’s evenness (J) for each sample site, considering the different fishery equipment used electric fishery equipment (P1, P2 and P3) and nets (P3, P4 and P5).

| | | N | S | H | MG | J |
|-----------------|----|-------|------|-----|-----|-----|
| Eletric fishery | P1 | 119.0 | 8.7 | 1.5 | 1.6 | 0.7 |
| | P2 | 133.3 | 12.7 | 1.9 | 2.4 | 0.8 |
| | P3 | 167.7 | 20.2 | 2.5 | 3.7 | 0.8 |
| Gill nets | P3 | 59.5 | 13.7 | 2.0 | 3.2 | 0.8 |
| | P4 | 24.0 | 7.7 | 1.7 | 2.2 | 0.9 |
| | P5 | 22.0 | 8.0 | 1.8 | 2.3 | 0.9 |

Table 4. Table ANOVA for indices of Margalef’s richness, Shannon’s diversity and Pielou’s evenness separately considering the catch with electric fishery (1, 2 and 3) and nets (3, 4 and 5).

| | FV | Eletric fishery (sites 1, 2 and 3) | | | | | Gill nets (sites 3, 4 and 5) | | | | |
|---------------------------|-------|------------------------------------|--------|-------|-------|--------|------------------------------|-------|-------|------|-------|
| | | df | SQ | QM | F | p | df | SQ | QM | F | p |
| Margalef’s richness index | Local | 2 | 12.58 | 6.29 | 23.49 | <0.001 | 2 | 35.17 | 17.58 | 3.78 | 0.047 |
| | Error | 15 | 40.169 | 2.68 | | | 15 | 69.66 | 4.64 | | |
| Shannon’s diversity index | Local | 2 | 2.12 | 1.06 | 35.04 | <0.001 | 2 | 0.76 | 0.38 | 3.12 | 0.073 |
| | Error | 15 | 0.45 | 0.03 | | | 15 | 1.83 | 0.12 | | |
| Pielou’s evenness index | Local | 2 | 0.030 | 0.015 | 10.01 | 0.001 | 2 | 0.010 | 0.005 | 0.51 | 0.612 |
| | Error | 15 | 0.022 | 0.001 | | | 15 | 0.151 | 0.010 | | |

FV) Sorce of variation; df) Degrees of freedom; SQ) Sum of squares; QM) Mean square, F) F statistic, p) Probability value.

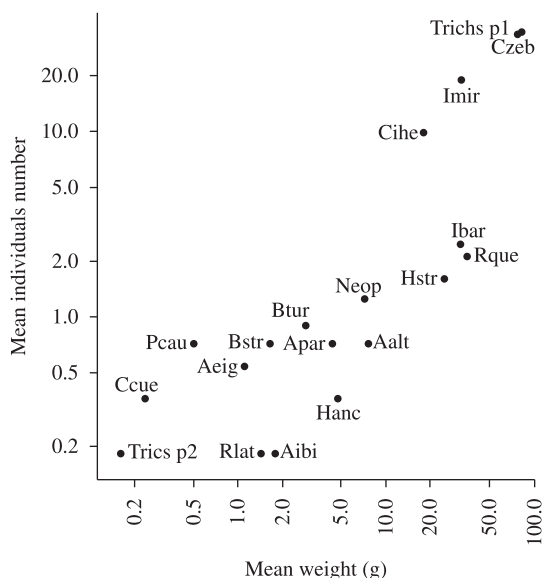


Figure 5. Relationship between weight and number of individuals caught in site 1, considering the individuals caught using electric fishing (for codes see Table 2).

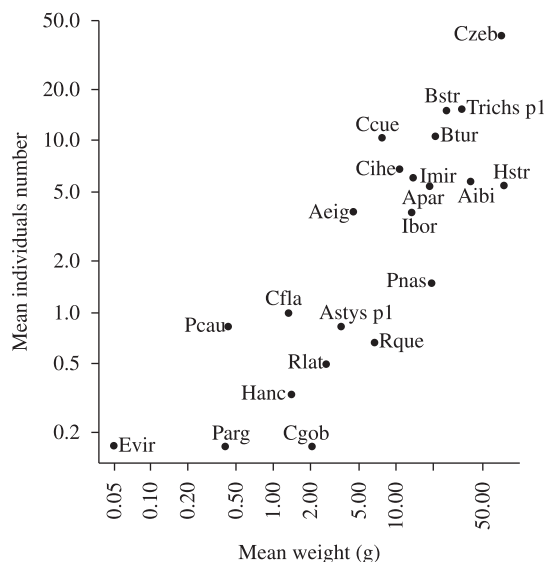


Figure 6. Relationship between weight and number of individuals caught in site 2, considering the individuals caught using electric fishing (for codes see Table 2).

4. Discussion

The Passa Cinco stream has a steep gradient, given its location in the *cuestas* of São Pedro and Analândia. In the headwaters, the water flow is faster, and the stream is more shallow and narrow. These characteristics directly influence the substrate composition. In such places, the

rocks are slightly larger and little suspended matter or background particulates are verified. All these features will be inverted along the channel. The large amount of water received in a short time, resulting from heavy rains in summer, makes the environment very susceptible to runoff, that quickly raises the water level. The substrate, depth and current are some of the most important physical

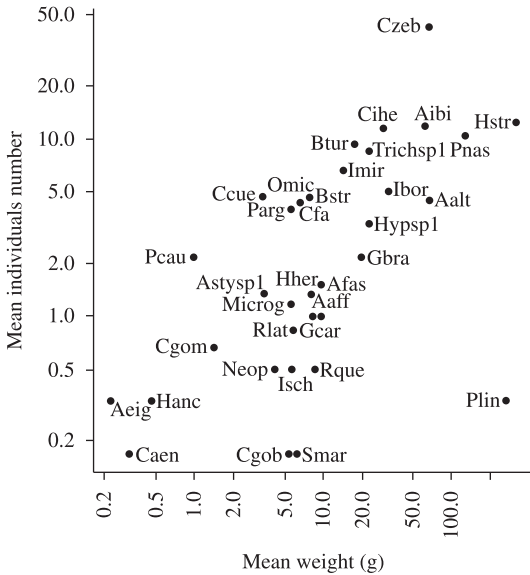


Figure 7. Relationship between weight and number of individuals caught in site 3, considering the individuals caught using electric fishing (for codes see Table 2).

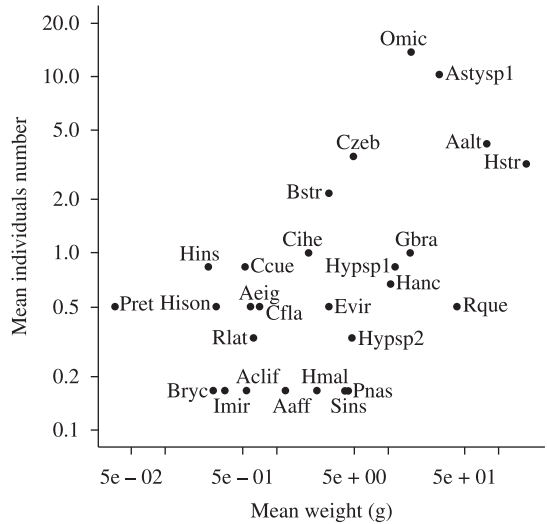


Figure 9. Relationship between weight and number of individuals caught in site 4, considering the individuals caught using nets (for codes see Table 2).

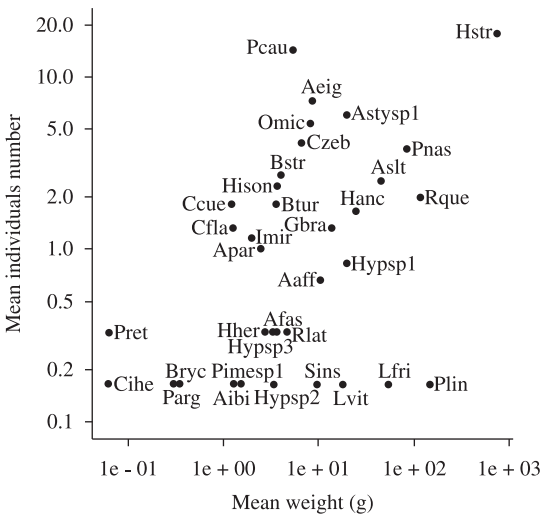


Figure 8. Relationship between weight and number of individuals caught in site 3, considering the individuals caught using nets (for codes see Table 2).

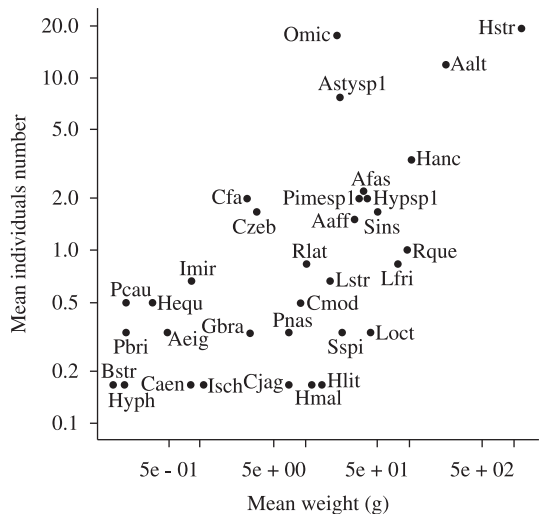


Figure 10. Relationship between weight and number of individuals caught in site 5, considering the individuals caught using nets (for codes see Table 2).

aspects in determining the distribution of stream fishes (Gorman and Karr, 1978; Argermeier and Karr, 1983). The combination of each aspects of these environmental characteristics produces a mosaic of microhabitats that changes along the gradient according to the physical conditions, which requires adjustments in the biological communities living there (Vannote et al., 1980).

In São Paulo state, the Alto Paraná system includes the major rivers and contains 38 families and 310 species of fish described (Langeani et al., 2007). In association with these major rivers, there is a large number of headwater streams inhabited mainly by species of small fishes with restricted

geographic distributions, such as *Bryconamericus turiuba*, *Astyanax bockmanni* and *Corumbataia cuetae*. The Passa Cinco stream is one of these streams with large numbers of small-sized species, some with restricted distributions.

Considering species diversity, richness and evenness, there was an increase from site 1 to 3, with further decrease in sites 4 and 5, agreeing with the values of the habitat's diversity. These patterns are not due exclusively to the fishery equipment employed, since site 3 was sampled with different equipments. In this site, all values had comparable magnitudes except number of individuals, which was remarkably higher in the electric fishery sites.

The ANOVA confirmed the increasing pattern of diversity, evenness and richness from site 1 to 3.

For individuals caught with nets, significant differences were observed only for the richness index. This was because sites 4 and 5 had the lowest number of species. The upstream reaches had lower habitat diversity than the downstream ones (Uieda and Barreto, 1999), which may explain the common pattern of higher species richness at lower portions (Gorman and Karr, 1978). Harrel et al. (1967) apud Peres-Neto (1995) suggest that an increase in species diversity along the river course can occur not only due to an increase in suitable habitats, but also a decrease in environmental fluctuations. The more severe and variable conditions in the headwaters requires from the organisms specific adaptations and higher energy to move against the current and to lessen the likelihood of being swept downstream (Allan, 1997). In more stable sites downstream, where the current is slower or even at a standstill, less energy expenditure is necessary from the individuals to stay in their positions (Allan, 1997).

Although there are fewer species in headwaters sites, there are larger populations, whereas in the downstream sections, the number of species increases and their population densities decrease. The species abundance distribution becomes more homogeneous leading to a higher evenness. This pattern can be seen for sites 1 to 3 (electric fishery) and sites 3 to 4 (gillnets). The same patterns would be expected following the increase in water volume (Garutti, 1988), but in this case, it would be necessary that the environment had a variety of physical structures to provide suitable habitats (Gorman and Karr, 1978; Argermeier and Karr, 1983). On the other hand, when the physical structures inside the stream channel are simplified, the increase in water volume can lead to an environmental homogenisation, as was observed for sites 4 to 5 in the Passa Cinco Stream.

Besides the occurrence and number of individuals, the species importance can also be defined by weight. These three additional measures provide different information about the fish assemblage (Ferreira, 2007). When considering only the species captured with electric fishery, *Characidium* aff. *zebra* was the most important species in the first three sample sites, occurring in large numbers and high weight. In addition to *C. aff. zebra*, other species were important for these sites (*Trichomycterus* sp.1, *Imparfinis mirini*, *Cetopsorhamdia itheringi*, *Bryconamericus stramineus*, *B. turiuba*, *Aperiodon ibitiensis*, *Parodon nasus* e *Hypostomus strigaticeps*). Headwater stream fishes from different families have morphological traits that allow them to better explore these environments (Braga, 2004). When considering individuals caught with nets (middle-lower portions), *H. strigaticeps* was the most important species in site 3, and remained important downstream, joining *Odontostilbe* aff. *microcephala*, *Astyanax* sp.1 and *A. altiparanae*. We suggest that the species mean abundance and weight are more or less related to population densities at each site. If so, we agree with Mazzoni (1998) that the structural components of the stream channel promote differential resource availability along the longitudinal gradient that, in turn, are correlated to these densities.

5. Conclusions

Considering the number of individuals and biomass, as well the changes in diversity, evenness and richness, it was found that, despite the short size of the Passa Cinco stream, the environments are well defined, and structured in headwater, transition (medium-low) and mouth portions. This was evidenced by the species distribution and assemblage composition along the gradient.

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