

## Edge effects on the blowfly fauna (Diptera, Calliphoridae) of the Tijuca National Park, Rio de Janeiro, Brazil

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(With 4 figures)

### Abstract

In this contribution we examine the diversity, abundance and species richness of Calliphoridae in the Tijuca National Park, Rio de Janeiro, RJ. Specimens were collected monthly between September 2009 and August 2010, using sardines as bait, in three points: A, on the forest edge (22°56'846''S 43°17'496''W), B, 700m from it (22°57'073''S 43°17'832''W) and C, 1,200m into the forest (22°57'321''S 43°18'031''W), evaluating the indicator species at each point and the anthropogenic influences and abiotic factors that determine species distribution. A total of 16,364 Calliphoridae were collected and 17 species were identified. Species abundance was strongly positively correlated with temperature. The greatest number of flies was collected at A, but in this point, diversity was lower and most individuals collected belonged to the dominant species. Point C, conversely, had the lower abundance and the highest diversity. *Chrysomya megacephala*, an urban and synanthropic species, was dominant at point A, whereas *Laneela nigripes* and *Mesembrinella peregrina*, typically forest species, were considered indicators at points B and C, respectively, showing that the anthropogenic influence is more intense at the forest edge.

**Keywords:** Atlantic Forest, diversity, environmental preservation, insect fauna.

## Efeitos de borda sobre a fauna de califorídeos (Diptera, Calliphoridae) no Parque Nacional da Tijuca, Rio de Janeiro, Brasil

### Resumo

O presente estudo visou analisar a diversidade, abundância e riqueza de espécies de Calliphoridae no Parque Nacional da Tijuca, Rio de Janeiro, RJ. Foram realizadas coletas mensais de setembro de 2009 a agosto de 2010, utilizando sardinha como isca, em três pontos: A, na borda (22°56'846''S 43°17'496''W); B, a 700m (22°57'073''S 43°17'832''W) e C, a 1.200m para o interior da mata (22°57'321''S 43°18'031''W), avaliando as espécies indicadoras em cada ponto, além da influência antrópica e os fatores abióticos que determinam a distribuição das espécies. Foram coletados 16.364 califorídeos, de 17 espécies. A abundância teve correlação extremamente forte e positiva com a temperatura. O ponto A foi o mais abundante, porém o menos diverso e com maior dominância. O ponto C, no entanto, foi o menos abundante e o mais diverso. *Chrysomya megacephala*, espécie sinantrópica e urbana, foi a espécie mais abundante no ponto A, enquanto *Laneela nigripes* e *Mesembrinella peregrina*, espécies caracteristicamente florestais, foram consideradas indicadoras para os pontos B e C, respectivamente, demonstrando que a influência antrópica atua de forma mais intensa no ponto da borda.

**Palavras-chave:** Mata Atlântica, diversidade, preservação ambiental, fauna de insetos.

### 1. Introduction

The Tijuca National Park (PARNATijuca), with an area of 35.51 km<sup>2</sup>, is located on the metropolitan area of the city of Rio de Janeiro. The vegetation in the reserve is part of the Atlantic Forest biome, which is known for

its high rates of endemism (ca 50%). Unfortunately, for many years the Atlantic Forest has been subjected to a growing and irreversible process of fragmentation; the original vegetation has been extensively destroyed, and

humans occupy most areas surrounding the few fragments left. As in with the remaining of the Atlantic Forest, the fauna of the PARNA Tijuca has suffered great impact as a result of human occupation, and the destruction of corridors uniting the forest fragments with other preservation areas in the state of Rio de Janeiro have caused the disappearance of many species (ICMBio, 2006).

In fact, human activities are largely responsible for the fragmentation of natural habitats, and most areas of natural vegetation that remain in the Atlantic forest are near humans (Janzen, 1983). Forest fragmentation dramatically increases edge effects, causing a number of immediate consequences, including a reduction in the habitat's size. Habitat modification has become a major cause of species extinction and biodiversity loss (Primack and Rodrigues, 2001).

Amplification of the edge effect, as defined by Forman and Gordon (1986), is one of the most important outcomes of habitat fragmentation, resulting in changes in structure, composition and/or relative abundance of species in the marginal part of a fragment. Other important modifications introduced by increasing the edges of forests are: more light, higher temperatures and winds, and decreasing of humidity (Rodrigues, 1998). The effects of these modifications are evident up to 500m into a forest (Laurance, 1991), but may vary according to the biological group under study. Because plant and animal species are adapted realistic to the actual conditions in there respective environment, the kinds of changes may cause them to be eliminated from a fragment, either directly, or indirectly, by allowing the establishment of alien species, which may be more adapted to the conditions that prevail in the changed environment and therefore more able to compete for resources there (Laurance et al., 2001).

From above, studies trying to understand how environmental changes affect different species have been conducted (Sousa, 2008). According to Brown (1997), insects are excellent environmental indicators because they are small, sensitive to changes, reproduce fast and their populations have high densities (Pais, 2003).

Human-insect associations are more common in urban areas, where waste, excrements of humans and domestic animals attract insects and serve as a substrate for the development of certain species (Mello et al., 2004). Humans often create ideal conditions for the proliferation of insects considered important to medical and veterinary medicine.

In the order Diptera, the family Calliphoridae is highlighted for the ability that its component species have to colonize new environments (Greenberg and Szyska, 1984). Furthermore, calliphorids have diverse feeding habits, are highly dispersive, and are efficient in locating resources, even when these are scarce or distant (Prado and Guimarães, 1982). Calliphoridae species have colonized different environments such as forests, rural or urban areas, and their prevalence and densities are determined not only by geography, but also by climate and food availability (Vianna et al., 1998). In addition, according Nuorteva

(1963), some Calliphoridae species are closely associated with anthropomorphic habitats.

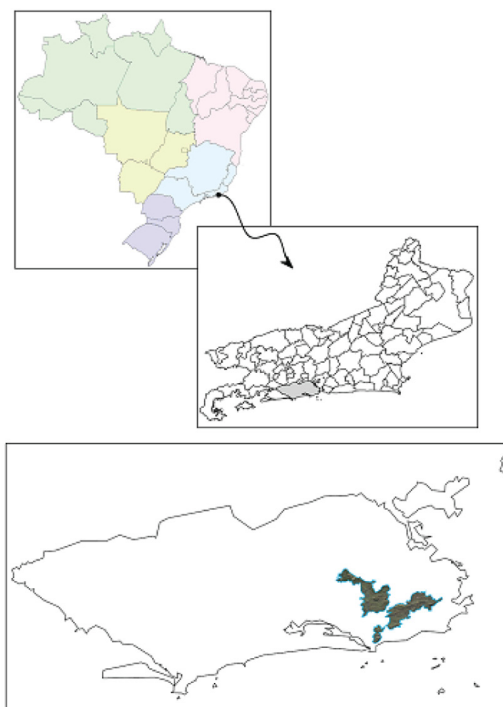
Studies on the diversity of Calliphoridae in protected areas of the state of Rio de Janeiro are restricted to the Reserva Biológica do Tinguá (hereinafter referred to as REBIO Tinguá) (Marinho et al., 2006; Mello et al., 2007; Ferraz et al., 2010b). For the PARNA Tijuca, only the publication by D'Almeida and Lopes (1983), which focused on synanthropic flies, is available. Because this study has been conducted over twenty years ago, and the PARNA Tijuca has suffered increased human pressure since then, we consider interesting it necessary to determine which species currently colonize this unit of urban forest.

The objectives of the present study were: (1) analyze the species in the family Calliphoridae at the PARNA Tijuca, (2) compare structural parameters such as abundance and species richness at the border and inside of the forest, (3) verify the presence of indicator species in the study area, and (4) from analysis of the species found, consider the possible influences of anthropogenic and abiotic factors on the blowfly fauna.

## 2. Material and Methods

We collected flies at the PARNA Tijuca (Figure 1) on a monthly basis between September 2009 and August 2010, on the first half of each month, with the exception of April, when we had to collect in the last two weeks because the park was closed during the first two, due to heavy rains.

Six traps were distributed following Mello et al. (2007), using sardines as baits exposed for 48 hours. The traps



**Figure 1.** Tijuca National Park, Rio de Janeiro, Brazil.

were arranged in pairs at three locations, one at the edge of the park (S22° 56' 846" W43° 17' 496"), where there is a big parking lot and people tend to congregate during the day, another 700m from it (S22° 57' 073" W43° 17' 832") and a third 1,200 m into the forest (S22° 57' 321" W43° 18' 031"), following the Pico do Bico do Papagaio trail. They were designated as points A, B and C respectively.

The specimens captured were euthanized and taken to the Laboratório de Estudos de Dípteros, UNIRIO. The samples were identified using the key of Mello (2003), and compared with specimens from the entomological collection of the National Museum/ UFRJ.

The Pearson's Correlation Test, using BioStat 5.0, was used to ascertain the correlation between abundance and environmental variables (temperature, relative humidity and rainfall) and the main species collected with climatic parameters. The latter were obtained from the National Institute of Meteorology (INMET), Jacarepaguá, Rio de Janeiro. The data were tested for normality using the D'Agostino test and the values for rainfall and blowfly abundance were logarithmized in order to be used in parametric tests. The Jaccard coefficient was used to assess the similarity of populations between sampling points.

Diversity was calculated with the programs Past and DivEs 2.0 from the following indices: Shannon-Wiener ( $H'$ ), which gives greater weight to rare species; equitability (Shannon  $J'$ ), used to describe the distribution of populations in the community; and Simpson's Dominance Index ( $D$ ), which is not significantly influenced by the rare species, measuring the probability of randomly finding two individuals of the same species in a population. The Indicator Species Test (Dufrêne and Legendre, 1997) was used to verify the

relationships among the species with each of the collection points. The results were subjected to a Monte Carlo test with 1,000 permutations using the program PC-ORD 4.1 to verify its significance given  $p < 0.05$ .

Following Krüger (2006), we defined a species as rare, intermediate, or common, according to the number of individuals collected: species with one or two individuals were considered rare; from three up to 51 individuals, intermediate; with 52 or more individuals, common. In order to determine whether a species was accidental, accessory or constant in each collecting point, the formula of constancy of occurrence was used –  $C = nx100/N$ , where  $n$  = number of samples containing the species under study,  $N$  = total number of samples collected (Dajoz, 1983).

### 3. Results and Discussion

During the sampling period, 16,364 blowflies were captured, distributed in 10 genera and 17 species. Table 1 shows the absolute and relative abundance of Calliphoridae species collected at each point during the study period. Point C had the lowest species richness (11), whereas A and B had 13 species each. On Table 1, *C. megacephala*, *L. nigripes*, *M. peregrina* and *H. semidiaphana* combined amount to more than 90% of the total abundance found, demonstrating the importance of these decomposers in the ecological balance of this ecosystem. The classification of species as rare, intermediate and common (Krüger, 2006) and accidental, accessory or constant (Dajoz, 1983) can be seen in Table 2.

We noticed the prevalence of females (87.06%) in baits when compared with males (12.94%), which may be due to the fact that females need substrates for oviposition

**Table 1.** Relative and absolute abundance of the species of Calliphoridae captured at each point \* at the Tijuca National Park, Rio de Janeiro, RJ, between September, 2009 and August, 2010.

Species	A	%	B	%	C	%	TOTAL	%
<i>Chloroprocta idioidea</i> (Robineau-Desvoidy, 1830)	04	0.04	01	0.03	00	0.00	05	0.03
<i>Cochliomyia hominivorax</i> (Coquerel, 1858)	00	0.00	01	0.03	00	0.00	01	0.01
<i>Cochliomyia macellaria</i> (Fabricius, 1805)	02	0.02	00	0.00	00	0.00	02	0.01
<i>Chrysomya albiceps</i> (Wiedemann, 1830)	312	3.38	16	0.40	14	0.44	337	2.06
<i>Chrysomya megacephala</i> (Fabricius, 1805)	7,029	76.06	187	4.72	330	10.43	7,546	46.11
<i>Chrysomya putoria</i> (Wiedemann, 1830)	02	0.02	01	0.03	00	0.00	03	0.02
<i>Eumesebrinella besnoiti</i> (Séguy, 1925)	00	0.00	00	0.00	01	0.03	01	0.01
<i>Eumesebrinella pauciseta</i> (Aldrich, 1922)	01	0.01	03	0.08	02	0.06	06	0.04
<i>Hemilucilia segmentaria</i> (Fabricius, 1805)	86	0.93	79	2.00	108	3.42	273	1.67
<i>Hemilucilia semidiaphana</i> (Rondani, 1850)	627	6.78	352	8.89	747	23.62	1,726	10.55
<i>Huascaromusca aeneiventris</i> (Wiedemann, 1830)	00	0.00	02	0.05	00	0.00	02	0.01
<i>Laneela nigripes</i> Guimaráes, 1977	318	3.44	1940	49.01	634	20.04	2,892	17.67
<i>Lucilia eximia</i> (Wiedemann, 1819)	779	8.43	1,129	28.52	682	21.56	2,590	15.83
<i>Mesembrinella bellardiana</i> Aldrich, 1922	76	0.82	211	5.33	346	10.94	633	3.87
<i>Mesembrinella peregrina</i> Aldrich, 1922	05	0.05	36	0.91	298	9.42	339	2.07
<i>Paralucilia borgmeieri</i> (Mello, 1969)	02	0.02	00	0.00	00	0.00	02	0.01
<i>Paralucilia pseudolyrcea</i> (Mello, 1969)	00	0.00	00	0.00	01	0.03	01	0.01
TOTAL	9,243		3,958		3,163		16,364	

\*Points: A= forest edge; B= 700m from forest edge; C = 1,200m into the forest.

**Table 2.** Classification of species as rare (1 or 2 individuals), intermediate (3 up to 51 individuals), common (more than 52 individuals) and accidental, accessory or constant ( $C = nx100 / N$ , where  $n$  = number of samples containing the species study,  $N$  = total number of sampling points), at each point and in general.

Species	A	B	C	General
<i>Chloroprocta idioidea</i>	intermediate and accidental	rare and accidental	absent	intermediate and accidental
<i>Cochliomyia hominivorax</i>	absent	rare and accidental	absent	rare and accidental
<i>Cochliomyia macellaria</i>	rare and accidental	absent	absent	rare and accidental
<i>Chrysomya albiceps</i>	common and accessory	intermediate and accessory	intermediate and accessory	common and accessory
<i>Chrysomya megacephala</i>	common and constant	common and constant	common and constant	common and constant
<i>Chrysomya putoria</i>	rare and accidental	rare and accidental	absent	intermediate and accidental
<i>Eumesebrinella besnoiti</i>	absent	absent	rare and accidental	rare and accidental
<i>Eumesebrinella pauciseta</i>	rare and accidental	intermediate and accidental	rare and accidental	intermediate and accessory
<i>Hemilucilia segmentaria</i>	common and constant	common and constant	common and constant	common and constant
<i>Hemilucilia semidiaphana</i>	common and constant	common and constant	common and accessory	common and constant
<i>Huascaromusca aeneiventris</i>	absent	intermediate and accessory	intermediate and accessory	intermediate and accessory
<i>Laneela nigripes</i>	common and constant	common and constant	common and constant	common and constant
<i>Lucilia eximia</i>	common and constant	common and constant	common and constant	common and constant
<i>Mesebrinella bellardiana</i>	common and constant	common and constant	common and constant	common and constant
<i>Mesebrinella peregrina</i>	intermediate and accessory	intermediate and constant	common and constant	common and constant
<i>Paralucilia borgmeieri</i>	rare and accidental	absent	absent	rare and accidental
<i>Paralucilia pseudolyrcea</i>	absent	absent	rare and accidental	rare and accidental

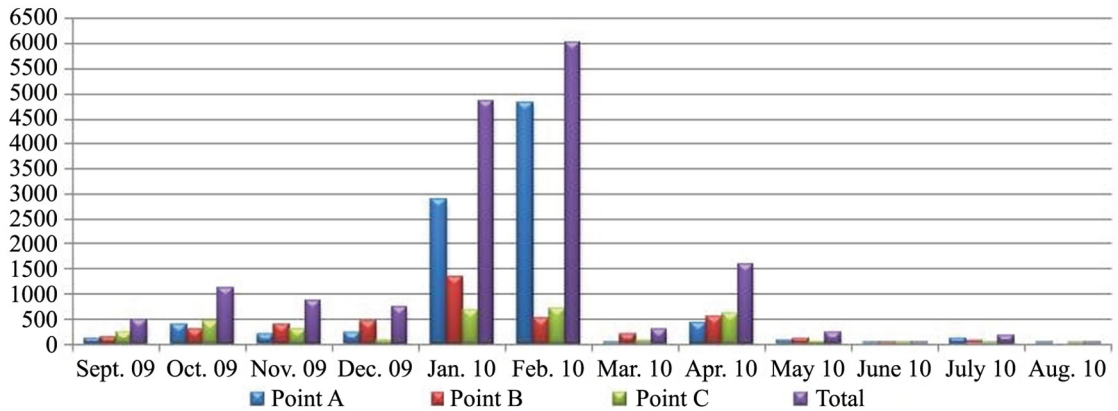
and maturation of their ovarian follicles (Avancini, 1988). This sex bias was also observed by Paraluppi and Castellón (1994), Marinho et al. (2006) and Ferraz et al. (2010a), and according to Sousa et al. (2010), may also be related to the position of the trap, suspended 1.5m above the ground.

The collecting of February 2010 resulted in the greatest numbers of specimens (Figure 2). Most specimens collected were *C. megacephala*, at point A. Contrasting with this number, the collections conducted in May to August 2010 contained much less calliphorids. Since the abundance of this family was strongly and positively correlated with temperature ( $r=0.9142$ ,  $p<0.0001$ ) in our data, we suppose that the low temperatures recorded in these months are responsible for the few numbers of individuals collected (Figure 3). According to Vogt and Woodburn (1982), blowflies peaked in the warmer seasons of the year.

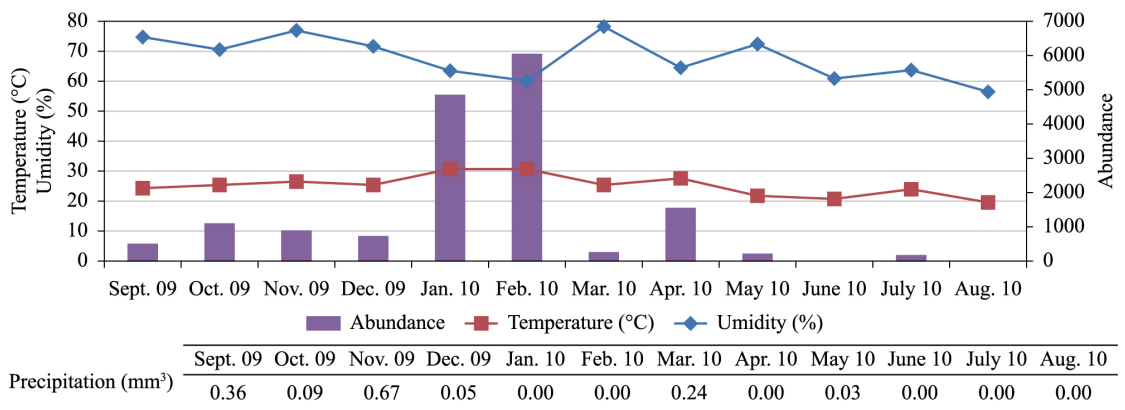
According to D'Almeida and Fraga (2007), prolonged rains can influence negatively on the abundance of Calliphoridae because the soil becomes soaked, which kills the pupae. However, in our data, no significant correlations were found between humidity ( $r=0.2781$ ,  $p=0.2150$ ) and rainfall ( $r=-0.1502$ ,  $p=0.6412$ ). The correlation between

blowfly abundance and climatic factors is extremely important, since Dajoz (1983) and Vianna et al. (2004) stated that climate variations are more important in the balance of Calliphoridae than biotic factors, which exert a secondary role.

The following diversity values were found in the study area: Shannon-Wiener index ( $H'=0.673$ ), equitability  $J$  ( $J'=0.5469$ ) and Simpson's dominance ( $D=0.8818$ ). The variations of these indices over the sampling period can be seen in Figure 4, highlighting the collection in September 2009, with the greatest diversity according to all indices. Contrasting with this, we observed the lowest value of the Shannon-Wiener diversity in August 2010, when only *M. peregrina* and *L. eximia*, two species considered constant, were captured. As a reminder, this index assigns greater weight to rare species. Point C was the most diverse ( $H'=1.843$ ,  $J'=0.7687$ ), followed by B ( $H'=1.382$ ,  $J'=0.5387$ ) and A ( $H'=0.926$ ,  $P=0.3612$ ). These results are in agreement with those of Ferraz et al. (2010b), who also found greater diversity in the farthest point from the edge (1,000m), and are also corroborated by the studies of Esposito (1999) in the Amazon, who found



**Figure 2.** Monthly distribution of Calliphoridae collected from September 2009 to August 2010, at points A (edge), B (700m from edge) and C (1,200m from edge) and total value, at Tijuca National Park, Rio de Janeiro, RJ, Brazil.



**Figure 3.** Calliphoridae population fluctuation and environmental variables (Temperature, Relative Humidity and Precipitation), during the collecting period, Tijuca National Park, Rio de Janeiro, RJ, Brazil.

that the level of diversity in anthropogenic environments tends to be lower than in forested areas.

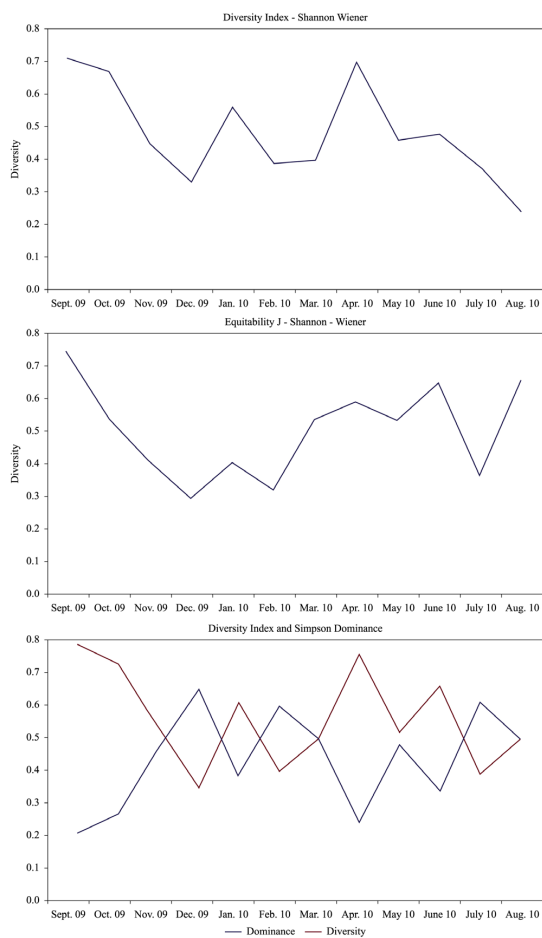
Studies have been developed to determine how edge effects are acting on the dipterous fauna in forests. McGeoch and Gaston (2000) observed differences in mortality and prevalence rates of Agromyzidae in suburban woods in UK. Penariol and Madi-Ravazzi (2013) observed a difference in abundance and richness of Drosophilidae in a fragment of semideciduous forest in Brazil. Ferraz (2011) reviewed this topic in Calliphoridae, underscoring the need for implementation of ecological studies to verify these edge effects in the population and highlight species as bioindicators. Thus, comparing our findings with those of the above authors, we suggest that there is a greater diversity of species of blowfly in environments where edge effects, usually caused by human impact, are less extensive.

The results obtained for species dominance were the opposite of those described above: point A has mostly dominant species ( $D=0.5925$ ), represented by *C. megacephala*, followed by point B ( $D=0.3351$ ), and

point C ( $D=0.1754$ ). The dominant species found in our results differ from Ferraz et al. (2009), who found that *C. albiceps* predominates in another area of the Atlantic Forest of Rio de Janeiro. This contrast is important because it shows that two areas in the same biome provide different overviews of the distribution and dominance of different species of calliphorids.

*Laneela nigripes* and *M. peregrina* were considered indicator species in B and C points, respectively ( $p=0.040$  and  $p=0.001$ ). This is very important, since both species belong to the subfamily Mesembrinellinae, described in the literature as being exclusively Neotropical and resident of dense forests and wetlands (Mello, 1967; Toma and Carvalho, 1995). Therefore, the PARNA Tijuca has the conditions for the establishment of forest species in the more internal areas. Mesembrinellinae has been identified by Gadelha et al. (2009) as possible indicators of preserved forest environments.

The values obtained by the Jaccard coefficient were high ( $A \times B=73.33\%$ ,  $A \times C=60\%$ ,  $B \times C=60\%$ ), and according to Mantovani (1987), this index rarely exceeds 60%, which



**Figure 4.** Monthly variation of the diversity index of the Calliphoridae (Shannon-Wiener ( $H'$ ), Shannon  $J'$  and Simpson's Dominance), during the collecting period at the Tijuca National Park, Rio de Janeiro, RJ, Brazil.

would indicate very similar populations. Thus, we can think PARNA Tijuca as a preserved area. Ferraz et al. (2010b) and Gadelha (2009) also observed this high similarity in studies at the REBIO Tinguá, demonstrating it is common when it comes to collecting sites that are near each other and in the same environment.

*Chrysomya* Robineau-Desvoidy 1830 species were more abundant in January and February 2010. This exotic genus was introduced in Brazil in the 1970s (Guimarães et al., 1978) and is highly synanthropic. *Chrysomya megacephala* was the most abundant species at point A (76.06%), and considered common and accessory at all collecting points. Its distribution is strongly influenced by temperature ( $r=0.7634$ ,  $p=0.0038$ ), as it occurs in greater abundance in the warmer months, as observed by D'Almeida and Fraga (2007). It was also the most abundant species in studies conducted by Paraluppi and Castellón (1994) in a region of the Amazon rainforest; by Rodrigues-Guimarães et al. (2004) in the metropolitan area of Rio de Janeiro; and by Mello et al. (2004) and Dias et al. (2009) next to garbage

dumps in Rio de Janeiro and São Paulo, respectively. Such high incidence of this species in different areas was explained by Prado and Guimarães (1982), who considered it as r-strategist with generalist feeding habits, a combination that results in high adaptability to diverse environments such as forest fragments. According to Vianna et al. (1998), this species has a preference for sites with anthropogenic influence, so its presence was stronger in the border area.

*Chrysomya albiceps*, on the other hand, was not as abundant as *C. megacephala*. This finding contrasts with the results obtained by Ferraz et al. (2009), who found that it was the third most often collected species, and by Costa et al. (1992), who found that *C. albiceps* was the species of *Chrysomya* most often collected in the urban area of Rio Grande do Sul. In the present study, *C. albiceps* was considered common only at the edge, intermediate in the remaining points, and accessory according to the constancy index. The abundance of *C. albiceps* was strongly correlated with temperature ( $r=0.7637$ ,  $p=0.0035$ ), contrasting with the results of Ferraz et al. (2010a), who did not find a correlation between the presence of *C. albiceps* and this parameter. In the results of D'Almeida and Lopes (1983), from the PARNA Tijuca, *C. albiceps* was highly synanthropic and was the second most frequent species in samples. We believe that the dominance of *C. megacephala*, especially at the area most impacted by humans (A), is in part responsible for the low counts of *C. albiceps*. The competition between these species is related to coexistence in ephemeral food resources as castings (Ullyett, 1950). When food resources are limited there may be competition, predation or cannibalism (Aguiar-Coelho and Milward-de-Azevedo, 1995; 1998).

The dominance of *C. megacephala* may have also been responsible for the few numbers of *C. macellaria*, previously observed by Rodrigues-Guimarães et al. (2004), D'Almeida and Fraga (2007) and Sousa et al. (2010). D'Almeida and Lopes (1983) recorded the highest abundance of *C. macellaria* in a rural area, contrasting with previous reports on the distribution of this species, which is considered urban, reflecting the shift that this species has undergone as a result of competition with *Chrysomya*.

The impact of the introduction of *Chrysomya* in Brazil also extends to the distribution of *L. eximia*. D'Almeida and Lopes (1983) found that this species was highly synanthropic in Rio de Janeiro; Paraluppi and Castellón (1994) also collected it more often from an urban area. However, as shown by Mello et al. (2007), Ferraz et al. (2010a) and in this study, the distribution of *L. eximia* has shifted to areas more inside the forest. In our data, it was more abundant at the point located 700m from the edge (B), and was common and constant in all sampling points. This may reflect the ability of this species to adapt to different environments, with more or less human pressure, as highlighted by Furusawa and Cassino (2006).

According to D'Almeida and Lopes (1983), species of *Hemilucilia* Brauer, 1895 are essentially Neotropical and prevail in forests. In this study, *H. segmentaria* was considered common and constant in all collection points,

and its presence was strongly correlated with temperature ( $r=0.8289$ ,  $p=0.0009$ ). This differs from the results of Ferraz et al. (2010a), who found a strong positive correlation of its presence with temperature, and a strong negative correlation with humidity and precipitation. In Sousa et al. (2010) in the Amazon, *H. segmentaria* was more abundant where there was some degree of preservation of the environment, or in an advanced process of recovery forest area. This corroborates the hypothesis of D'Almeida and Lopes (1983) that this species prefers uninhabited forest areas.

*Hemilucilia semidiaphana*, the most abundant species in the studies of Ferraz et al. (2010a), was considered common and constant at all sampling points except at point C, where it was accessory because it have not occurred in large number of samples, despite the greater abundance of this species there. Its distribution was strongly influenced by temperature ( $r=0.7688$ ,  $p=0.0035$ ) and had no significant relationship with the other variables. This contrasts with the results of Ferraz et al. (2010a), who found a negative correlation with moisture and precipitation. According to D'Almeida and Lopes (1983) this species avoids inhabited areas, and was collected exclusively in forest areas by Paraluppi and Castellón (1994), confirming that *H. semidiaphana* is a synanthropic, a habit also confirmed by the results of Furusawa and Cassino (2006).

Regarding Mesembrinellinae, Sousa (2008) stated that some environmental factors may limit the distribution of some species of this subfamily, which in turn may be more sensitive to environmental changes. In the present study, we collected six species of Mesembrinellinae, being *E. besnoiti* and *H. aeneiventris* were considered rare and incidental, whereas *E. pauciseta* was intermediate at point B, rare at A and C, and accessory when analyzed for all points combined.

The presence of *L. nigripes* in samples was strongly influenced by temperature ( $r=0.7838$ ,  $p=0.0025$ ), contrasting with the results of Gadelha (2009) that showed no such correlation. This species was the most abundant at the midpoint, and was considered common and constant at all collection points, as also observed by Gadelha (2009) and Ferraz et al. (2010b). Mello et al. (2007) reported this species as the most abundant at the REBIO Tinguá, demonstrating that it is distributed in areas of the Atlantic Forest of Rio de Janeiro, especially in the range between 500 and 700m.

The presence of *M. bellardiana* in traps was strongly influenced by temperature ( $r=0.762$ ,  $p=0.004$ ), differing from Gadelha (2009), who found no significant correlation. The synanthropic index of this species was -100 in the studies of D'Almeida and Lopes (1983), which reflects this species' complete aversion to anthropic environments. In the present study, the species was considered common and constant at all sampling points, but its distribution was more concentrated inside the forest. This result is in agreement with those results obtained by Ferraz et al. (2009), in which this species was the most abundant in the innermost point (1,000m) of REBIO Tinguá. However, *M. bellardiana* seems to be more concentrated in the range

of 1,200 to 1,000m inside the forest, since specimens were less abundant in the point located at 2,000m in the study of Gadelha (2009).

*Mesembrinella peregrina* was considered common and constant only at point C, intermediate and constant at point B and intermediate and accessory at point A, demonstrating its preference for areas inside the forest and away from human activity. The presence of this species in our samples was not significantly influenced by any of the environmental variables analyzed (temperature  $r=0.4556$ ,  $p=0.1366$ ; humidity  $r=0.0121$ ,  $p=0.9702$ ; rainfall  $r=-0.1328$ ,  $p=0.6807$ ), and its abundance was probably influenced by biotic factors such as competition with other species of flies that share the same food resources, food availability, attractiveness of the bait, among other factors.

There are few studies involving ecology and taxonomic identification of Mesembrinellinae, complicating the analysis of the results presented here. The case of *M. peregrina* is even more complex, since it was not recorded in any of the studies conducted on Calliphoridae in the State of Rio de Janeiro or in other forest areas of the country.

The species *C. hominivorax*, *C. macellaria*, *E. besnoiti*, *H. aeneiventris*, *P. borgmeieri* and *P. pseudolyrcea* were considered rare and accidental in our results, whereas *C. idioidea* and *C. putoria* were intermediate and accidental, and *E. pauciseta* was intermediate and accessory when considering all points combined. Thus, it is not possible to draw conclusions regarding the distribution and population fluctuation of these species. However, it is noteworthy that we had not expected to collect *C. hominivorax* because of the biontophagous habits of this species. Also, *C. idioidea*, an accidental species in relation to our data, was the most abundant species in the work of Sousa et al. (2010) in the Amazon, totaling over 80% of Calliphoridae collected. It seems that this species, also considered accidental in the results of Ferraz et al. (2010b), is more adapted to the Amazon forest biome than to the Atlantic forest.

Thus, we can conclude that 17 species of Calliphoridae are distributed along the forest in the PARNA Tijuca, including species considered urban and forest. The edge effect was felt in diversity, richness and abundance of species of Calliphoridae, where the lowest point of forest fragmentation, more internalized located at 1,200 m (point C) the edge was the most diverse, while the A and B points, with greater fragmentation and located 700m on the edge, respectively, were marked by the presence of dominant species. The Calliphoridae were influenced by human presence, ranging distribution along gradients of forest, according to the characteristics of each species, eg, *Chrysomya megacephala*, a synanthropic species, was the most abundant species in the anthropogenic point (edge), whose richness was the lowest among the points studied. While at points of lower human action, more internalized areas predominated forest species, highlighting *Laneela nigripes* and *Mesembrinella peregrina* that were indicator species of the points B and C, respectively. The distribution of Calliphoridae was influenced by humans along the forest gradient, according to the characteristics of each species, and was also strongly influenced by temperature.

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