

# Estimated richness and environmental correlates of miniature fish assemblages in the rio Jacundá, Brazil

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South America is home to more miniature fishes (<26 mm in standard length) than any other continent. Despite this diversity, the ecology of miniature fishes is poorly studied. To promote the study of miniature fish ecology, we investigated patterns in total richness, assemblage structure and environmental correlates for miniature fishes in the rio Jacundá drainage of the Lower Amazon River basin, Pará State. Based on multi-pass dip-netting of leaf litter at 20 locations distributed across two sites, we collected miniature species and used rarefaction to estimate 9 to 14 species might be present. The miniature fish assemblage at the upstream site was a nested subset of the downstream site, and water pH and canopy cover, two features known to be altered by deforestation, correlated most strongly with assemblage variation. Our work represents one of the first quantitative assessments of environmental correlates with miniature fish assemblages and highlights research topics that should be investigated further to promote conservation and preservation of the overlooked and understudied Amazonian diminutive freshwater fish fauna.

**Keywords:** Community, Conservation, Miniaturization, Ordination, Rarefaction.

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A América do Sul abriga o maior número de peixes miniaturas (<26 mm de comprimento padrão) do que qualquer outro continente. Apesar dessa diversidade, a ecologia dos peixes miniaturas é pouco estudada. Visando promover estudos de ecologia de peixes miniaturas, investigamos padrões de riqueza total, estrutura da assembleia e fatores ambientais correlacionados para peixes miniaturas no rio Jacundá, drenagem do baixo rio Amazonas, Pará. Com base em múltiplas passagens de redes no sedimento em 20 pontos distribuídos em dois locais, coletamos espécies miniaturas e usamos rarefação para estimar que 9 a 14 espécies podem estar presentes. A assembleia de peixes miniaturas no local à montante foi um subgrupo aninhado na assembleia no local à jusante, e pH da água e cobertura de copas, dois fatores sabidamente alterados por desmatamento, foram os mais correlacionados com a variação na assembleia. Nosso trabalho representa um dos primeiros estudos quantitativos de fatores ambientais correlacionados às assembleias de peixes miniaturas e ressalta um tópico de pesquisa que deveria ser melhor investigado para promover a conservação da pouco conhecida fauna de peixes diminutos da Amazônia.

**Palavras-chave:** Comunidade, Conservação, Miniaturização, Ordenação, Rarefação.

## INTRODUCTION

The evolution of extremely small adult body size is known as miniaturization and is a common theme in animal evolution with examples from all major metazoan taxa (Hanken, Wake, 1993; Rundell, Leander, 2010). The process of miniaturization is particularly notable among freshwater fishes in South America, including species  $\leq 26$  mm standard length (SL) at maximum size (reviewed by Weitzman, Vari, 1988; Toledo-Piza *et al.*, 2014). More than 200 species of miniature freshwater fishes are known in South America, a number that exceeds that of any other continent (Toledo-Piza *et al.*, 2014) and continues to grow through the description of new species (*e.g.*, Abrahão *et al.*, 2019; Carvalho, Reis, 2020; Mattox *et al.*, 2020, 2021; Rodrigues, Netto-Ferreira, 2020). In their description of “the almost invisible league”, Carvalho *et al.* (2006) described communities of miniature fishes that occupy the benthos of leaf-littered, black-water streams in the Amazon. These miniature fishes tend to occupy benthic (Henderson, Walker, 1990) and floating (Carvalho *et al.*, 2013) litter banks, suggesting some level of specialized habitat requirements. However, the ecology of miniature fishes is not well studied, having remained underexplored in the same way as most of the freshwater fish diversity in South America (Reis *et al.*, 2016).

Miniature fishes are overlooked and understudied for a number of reasons. The capture of miniature species requires specialized sampling gear with mesh sizes that are finer than what is typically used, and gears must be deployed in a manner that involves digging into the substrate or benthic debris (*e.g.*, Carvalho *et al.*, 2014). When they are captured, miniature fishes are sometimes dismissed as juvenile forms of other, larger species or are disregarded altogether in favor of focus on larger species (Weitzman, Vari,

1988). Confident identification requires voucher specimens be returned to the laboratory because identification in the field can be extremely challenging (Van der Sleen, Albert, 2018). Moreover, an overall similarity in general appearance between congeners and uncertainty regarding the taxonomy of miniature fishes can complicate identification and phylogenetic placement (e.g., Britz *et al.*, 2014). Despite these challenges, miniature fishes represent a significant and growing component of global biodiversity with new species descriptions outpacing our understanding of their ecology (Costa, Le Bail, 1999; Toledo-Piza *et al.*, 2014).

In their review of miniaturization in South American freshwater fishes, Weitzman, Vari (1988) noted that the majority of miniature species inhabit lentic or slow flowing waters. Although the authors noted many streams inhabited by miniature fishes were characterized by low pH, Weitzman, Vari (1988) were careful not to identify specific environmental parameters as potential selective agents for miniaturization and called for further investigations on miniature fishes, especially studies that would shed light on potential physical environmental parameters shaping the ecology and evolution of miniaturization. Subsequent reviews of miniature fishes in other regions (Kottelat, Vidthayanon, 1993; Conway, Moritz, 2006; Bennett, Conway, 2010) also highlight the link between small body size and lentic or slow flowing habitats, especially swamps that may experience annual periods of drought in which they are reduced to a small series of pools (Kottelat, Vidthayanon, 1993; Kottelat *et al.*, 2006). Additional research on the ecology of miniature fishes is needed to advance our understanding of the mechanisms that generate and maintain miniature fish diversity. This is especially true in the Amazon, where anthropogenic alterations to natural rainforest land cover may be altering the structure and function of the riverscapes that contributed to the impressive diversity of miniature fishes in this region of the Neotropics. For example, Ríos-Villamizar *et al.* (2017) found that accumulated total deforestation in the rio Purus correlated with a decrease in water pH, and Carvalho, Uieda (2010) found deforestation contributed to open canopies over streams and ultimately reduced the leaf-litter habitats commonly occupied by miniature fishes (Henderson, Walker, 1990; Carvalho *et al.*, 2006). Deforestation is also linked to changes in stream fish body size at both the individual and assemblage levels (Ilha *et al.*, 2018), meaning natural fish body size gradients are affected by anthropogenic activities. The effects of deforestation on larger-bodied Amazonian stream fishes point to high importance of local habitat factors (Montag *et al.*, 2019), but local habitat correlates of miniature fish assemblages have not been the focus of quantitative studies to our knowledge. Consequently, research pertaining to relationships between local habitat variables and miniature fish assemblages is necessary to improve our understanding of how future changes to Neotropical river systems might influence these remarkable diminutive fishes (Albert *et al.*, 2020).

This study aimed to investigate patterns in total richness, assemblage structure and environmental correlates for miniature fishes in the rio Jacundá, a tributary of the Lower Amazon River Basin. We chose this tributary because it houses a large number of endemic and miniature fishes and because of ongoing and rapid habitat loss in the region (Claro-García, Shibatta, 2013). First, we used rarefaction and species accumulation curves to estimate miniature and non-miniature fish richness sampled with specialized gear targeting miniature species. We expected that the number of miniature fish species would plateau with less effort compared to non-miniature fishes

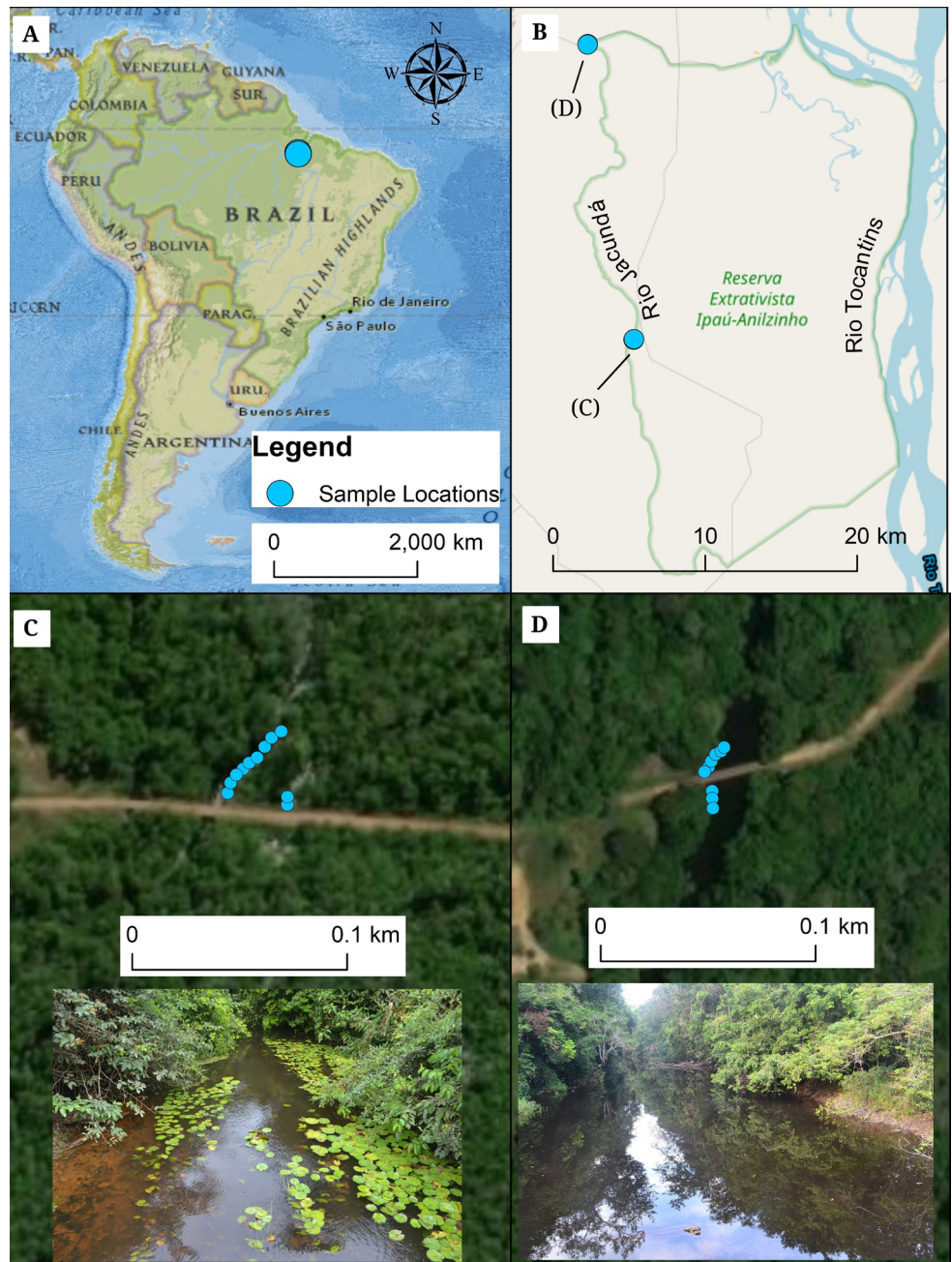
given that sampling methods favored collection of smaller fishes. Second, we assessed differences in assemblage composition at two sites located within the same tributary stream. We expected that either nesting or turnover in the identity of miniature fishes would contribute to differences in assemblage composition between the two sites and that species would partition habitats such that they rarely occurred together (Henderson, Walker, 1990). Collectively, our analyses fill an important knowledge gap concerning the ecology of miniature freshwater fishes and provide insight into habitat features that might be important for conservation.

## MATERIAL AND METHODS

We collected fishes from the rio Jacundá, a stream in the lower Amazon (Pará State) that flows into the Furo de Santa Maria, which empties into the Atlantic Ocean with the rio Tocantins. The rio Jacundá is an oligotrophic, black-water, floodplain river characterized by low pH and high density of leaf litter. During the dry season in August 2019, we visited two sites 20 km apart along the rio Jacundá on the western edge of Reserva Extrativista Ipaú-Anilzinho, Pará, Brazil and sampled 20 locations (11 at the upstream site and nine at the downstream site; Fig. 1). We distributed sampling locations along gradients of habitat alteration associated with road crossing construction, particularly clear-cutting at the road edge transitioning to natural forest cover with distance from road crossings. At the center of each sampling location, we measured water temperature ( $^{\circ}\text{C}$ ), pH, conductivity ( $\mu\text{s}/\text{cm}$ ), total suspended solids (parts per million), dissolved oxygen ( $\text{mg}/\text{L}$ ), and water depth (cm). We visually estimated percent of benthos covered by leaf litter, emergent vegetation, detritus, and open water, and we used a concave densitometer to measure percent canopy cover above each sampling location. We collected the global position system (GPS) coordinates for each sampling location to develop spatial variables that could be used to assess spatial relationships among the sampling locations, specifically asymmetric eigenvector maps (Blanchet *et al.*, 2008). We first measured the hydrographic distance (m) between each point along the river course and created a graph theoretic representation of the locations denoted as nodes with edges (lines) connecting these nodes. We used the inverse distance between nodes to weight the edges (Dray *et al.*, 2006) and directed edges according to the flow of water. We then used the 'aem' function from the 'adespatial' package in R to create multivariate vectors representing spatial relationships among fish sampling locations (Dray *et al.*, 2020).

We used a combination of fine-mesh (0.8 mm) dip-netting and seining to collect fishes from a 2 m by 2 m area within wadeable (*i.e.*, <1 m deep) portions along the stream bank at each sampling location. Our standardized sampling protocol at each location included an initial seine haul through the 2 m x 2 m area, followed by three rounds of dip netting (by two netters) to scoop leaf litter from the benthos, followed by a second seine haul. Dip netters covered the entire benthos of the sampling area by digging handheld nets into the leaf litter and retrieving both litter and fishes. Leaf litter collected via dip netting was inspected for fishes prior to being discarded back into the stream. Fishes collected via seine and dip nets at each location were euthanized immediately in a solution of river water and clove oil (eugenol). Euthanized fishes were

**FIGURE 1** | Study area map illustrating **A**. The location of the rio Jacundá in the lower Amazon Basin, **B**. The two areas sampled on the western edge of the Reserva Extrativista Ipaú-Anilzinho, Pará, Brazil, **C**. The distribution of 11 sampling locations and a site photograph at the upstream area, and **D**. The distribution of nine sampling locations and a site photograph at the downstream area.



fixed in solution of a 4% buffered formalin and transferred to a solution of 70% ethanol for preservation. In the laboratory, specimens were identified to the lowest possible taxonomic level (*i.e.*, species) using appropriate taxonomic keys (*e.g.*, van der Sleen, Albert, 2018) and measured for SL (mm). Specimens were classified as miniature if the known adult size was  $\leq 26$  mm SL (*sensu* Weitzman, Vari, 1988) based on Toledo-Piza *et al.* (2014). Voucher specimens are deposited at Laboratório de Ictologia de Sorocaba (LISO), UFSCar, *campus* Sorocaba.

We used a subsampling-based statistical analysis to assess the richness of the miniature fish assemblage in the sampled area of the rio Jacundá. For our first objective, we used the rarefaction and extrapolation techniques of Chao *et al.* (2014) to assess accumulation

of species (separated by miniature *vs.* non-miniature) across our samples and projected richness to two points, including twice our sampling effort (*i.e.*,  $n = 40$  samples) and asymptotic richness. We used the 'iNEXT' function from the 'iNEXT' package in R (Hsieh *et al.*, 2020) to generate curves for miniature and non-miniature fishes independently. Preliminary results from this analysis illustrated complete sampling of miniature fishes but incomplete sampling of non-miniature fishes, so we proceeded with analysis of only miniature fish assemblages. We reviewed the conservation status listed by the International Union for the Conservation of Nature (IUCN; <https://www.iucnredlist.org/>) for each of the miniature species identified.

To examine our second objective, we used non-metric multidimensional scaling (NMDS) only on miniature fish assemblages to visualize differences in assemblage structure between the two sampling sites and among sampling locations ( $n = 20$ ). The NMDS was implemented with the 'vegan' package in R (Oksanen *et al.*, 2019) and based on the Bray–Curtis distance metric (Bray, Curtis, 1957) calculated from Hellinger-transformed abundance data (López-Delgado *et al.*, 2019). We used the 'envfit' function from the 'vegan' package to identify environmental and spatial variables that significantly correlated with assemblage structure. We quantified species co-occurrence by calculating all pairwise Pearson correlation coefficient between species using the 'corrplot' function from the 'corrplot' package in R (Taiyun, Viliam, 2017). Prior to analyses, we log-transformed all continuous environmental variables that were measured on a non-bound scale, logit-transformed all continuous environmental variables measured on a percentage scale, and then standardized all variables to a mean of zero and standard deviation of one (*i.e.*, *z*-score transformation) to better approximate normal distributions. Following the NMDS, we tested for a significant difference in assemblage structure between the two sampling locations using permuted multivariate analysis of variance (pmanova) implemented with the 'adonis2' function from the 'vegan' package using the transformed miniature fish assemblage matrix as the response variable and a two-level factor representing the two sites. All analyses were conducted in R version 3.6.2 (R Development Core Team, 2019).

## RESULTS

We collected nine miniature and 44 non-miniature species across the 20 sampling locations, yielding a total of 717 miniature and 1,199 non-miniature specimens. One of the miniature and one of the non-miniature species we collected are not formally described (GMTM, pers. obs.), and two specimens were represented only by juveniles that could not be identified below the level of genus (Tab. 1). None of the miniature fishes collected were evaluated by the IUCN (Tab. 2). Accumulation of miniature fish richness across sampling effort leveled out between 10 and 20 samples, while accumulation of non-miniature fishes continued past our 20 samples and was still increasing at twice our sample size (Fig. 2). The asymptotic richness estimation (95% confidence range) was 9 (9–14) species for miniatures and 59 (49–96) species for non-miniatures. Average SL was 12.1 mm (range = 8.3–20.1) for miniature fishes and 24.1 mm (range = 4.5–151.1) for non-miniature fishes.

**TABLE 1** | Fishes collected from the rio Jacundá in the Amazon basin, Brazil, during August 2019, including the number of specimens collected and mean standard length of specimens. Miniature fishes are shown in bold text. Lengths were not recorded for two juvenile non-miniature specimens (“-”). Vouchers are deposited at Laboratório de Ictiologia de Sorocaba (LISO).

Order	Family	Species	Number collected	Mean standard length (mm)	Voucher number
Characiformes	Acestrorhynchidae	<i>Gnathocharax steindachneri</i>	7	21.90	374
	Characidae	<b><i>Hemigrammus aff. tridens</i></b>	14	11.39	381
		<i>Hemigrammus bellottii</i>	110	22.53	378
		<i>Hemigrammus cf. orthus</i>	103	13.62	375
		<i>Hemigrammus cf. schmardae</i>	190	14.28	377
		<i>Hemigrammus levis</i>	26	27.66	376
		<i>Hemigrammus ocellifer</i>	54	24.90	379
		<i>Hemigrammus rhodostomus</i>	12	28.10	380
		<i>Hyphessobrycon heterorhabdus</i>	1	18.21	382
		<i>Hyphessobrycon rosaceus</i>	106	17.08	383
		<i>Moenkhausia collettii</i>	10	34.21	384
		<i>Pristella maxillaris</i>	1	18.54	385
		<b><i>Tyttobrycon sp.</i></b>	207	11.27	386
		Crenuchidae	<b><i>Ammocryptocharax minutus</i></b>	5	15.56
	<i>Crenuchus spilurus</i>		11	28.03	388
	<b><i>Elachocharax pulcher</i></b>		9	13.99	389
	<b><i>Microcharacidium weitzmani</i></b>		164	12.35	390
	Curimatidae	<i>Curimatopsis crypticus</i>	112	22.71	391
	Erythrinidae	<i>Hoplerethrinus unitaeniatus</i>	1	117.83	392
		<i>Hoplias malabaricus</i>	1	96.37	393
	Iguanodectidae	<i>Iguanodectes rachovii</i>	15	41.35	394
	Lebiasinidae	<i>Copella arnoldi</i>	37	20.45	395
		<i>Nannostomus beckfordi</i>	110	24.73	396
<i>Nannostomus eques</i>		4	25.56	397	
<i>Nannostomus harrisoni</i>		20	24.74	398	
<i>Nannostomus limatus</i>		10	26.29	399	
Serrasalminidae	<i>Serrasalmus sp.</i> (juvenile)	1	-	400	
Siluriformes	Aspredinidae	<i>Bunocephalus coracoideus</i>	19	45.67	401
	Callichthyidae	<i>Megalechis picta</i>	2	66.80	402
	Doradidae	<b><i>Physopyxis ananas</i></b>	54	15.06	403
	Heptapteridae	<i>Gladioglanis conquistador</i>	51	23.06	404
	Loricariidae	<i>Otocinclus aff. mura</i>	1	25.50	405
	Trichomycteridae	<b><i>Ammoglanis amapaensis</i></b>	2	13.44	406



TABLE 1 | (Continued)

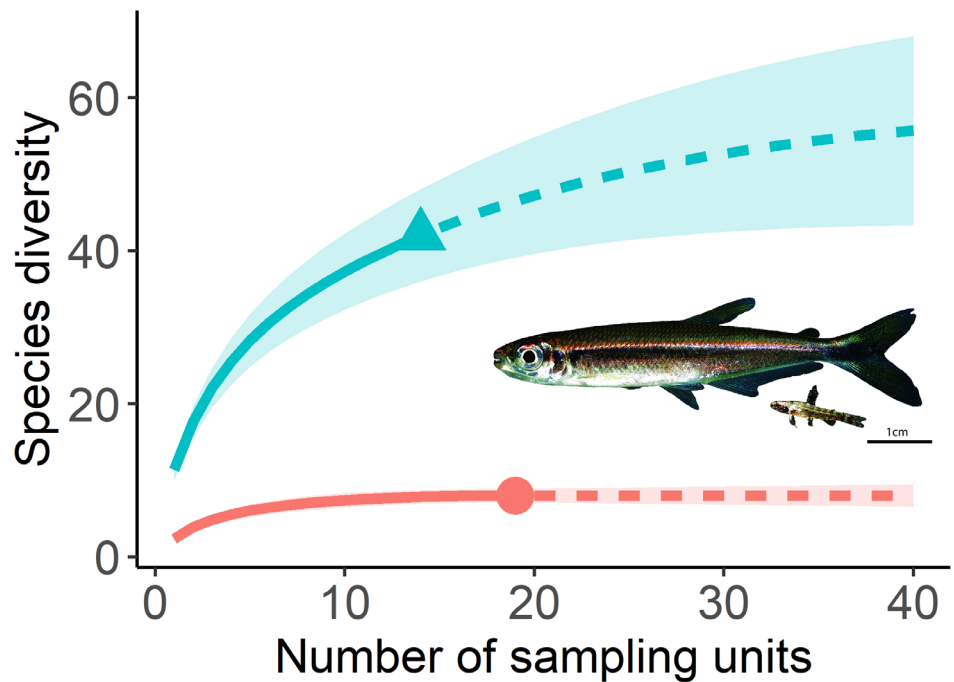
Order	Family	Species	Number collected	Mean standard length (mm)	Voucher number
		<i>Potamoglanis hasemani</i>	48	11.91	407
Gymnotiformes	Hypopomidae	<i>Brachyhypopomus</i> sp.	2	64.36	408
		<i>Microsternarchus bilineatus</i>	42	73.74	409
	Gymnotidae	<i>Gymnotus coropinae</i>	6	50.24	410
	Rhamphichthyidae	<i>Gymnorhamphichthys rondoni</i>	1	108.94	411
		<i>Steatogenys duidae</i>	4	53.95	412
Beloniformes	Belonidae	<i>Potamorhaphis guianensis</i>	3	140.60	413
Cyprinodontiformes	Fluviophylacidae	<i>Fluviophylax palikur</i>	216	12.56	414
	Rivulidae	<i>Kryptolebias campelloi</i>	1	12.94	415
		<i>Kryptolebias marmoratus</i>	5	20.97	416
Cichliformes	Cichlidae	<i>Aequidens</i> sp. (juvenile)	1	–	417
		<i>Aequidens tetramerus</i>	1	58.15	418
		<i>Apistogramma agassizii</i>	92	17.96	419
		<i>Apistogramma gossei</i>	3	21.20	420
		<i>Crenicichla inpa</i>	1	59.11	421
		<i>Crenicichla regani</i>	14	38.72	422
		<i>Hypselacara coryphaenoides</i>	1	43.78	423
Perciformes	Polycentridae	<i>Monocirrhus polyacanthus</i>	1	73.98	424
Synbranchiformes	Synbranchidae	<i>Synbranchus marmoratus</i>	4	82.85	425

The NMDS analysis showed clear separation of fish assemblages within and among the two sites, with a two-dimensional stress value of 0.13 (Fig. 3). The upstream site (Fig. 1C) was a nested subset of the downstream site (Fig. 1D) with four of the nine miniature species occurring at only the downstream site, including *Ammocryptocharax minutus* Buckup, 1993, *Elachocharax pulcher* Myers, 1927, *Hemigrammus* aff. *tridens*, and *Tyttobrycon* sp. (Tab. 2). With the exception of *Ammoglanis amapaensis* Mattos, Costa & Gama, 2008, all species that occurred upstream had higher abundances upstream relative to downstream. These patterns contributed to a significant difference between the upstream and downstream sites based on the pmanova test ( $F_{1,18} = 9.7$ ,  $P < 0.01$ ,  $R^2 = 0.35$ ). The environmental variables that correlated with assemblage structure based on the 'envfit' analysis (Tab. 3) were pH ( $R^2 = 0.33$ ,  $P = 0.03$ ), water temperature ( $R^2 = 0.32$ ,  $P = 0.04$ ), conductivity ( $R^2 = 0.31$ ,  $P = 0.04$ ), percent canopy cover ( $R^2 = 0.30$ ,  $P = 0.04$ ), and total suspended solids ( $R^2 = 0.29$ ,  $P = 0.04$ ). The only spatial variable that correlated with assemblage structure was the first axis of the asymmetric eigenvector map (i.e., AEM1,  $R^2 = 0.50$ ,  $P < 0.01$ ). Pairwise correlations of species abundance at the 20 sampling locations showed low correlations in species occurrence, with only *Microcharacidium weitzmani* Buckup, 1993 and *Tyttobrycon* sp. having an absolute correlation coefficient  $>0.50$  (Fig. 4).

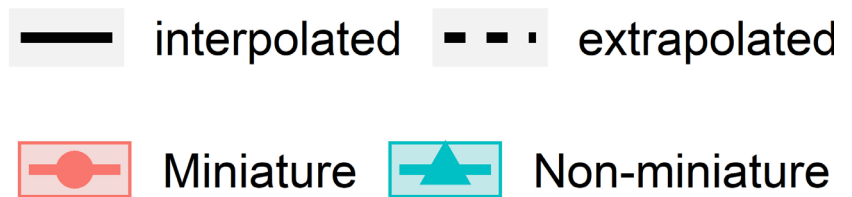


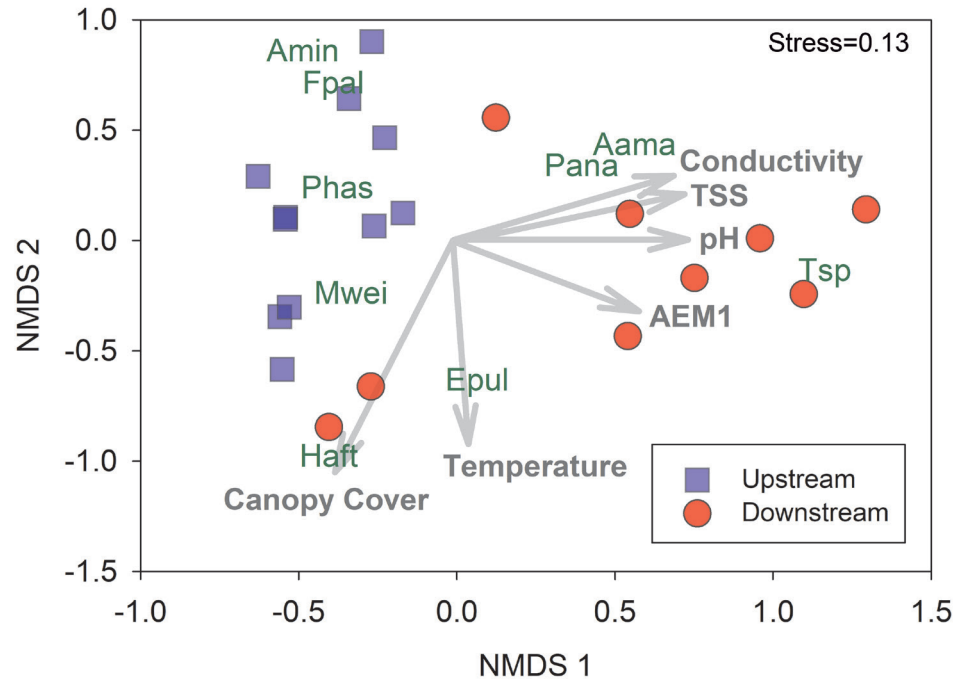
**TABLE 2** | Abundances of miniature fish species collected at the upstream and downstream sampling sites (see Fig. 1 for locations) and conservation status according to the International Union for the Conservation of Nature (IUCN). NE = Not Evaluated.

Species	Upstream	Downstream	IUCN
<i>Ammocryptocharax minutus</i>	2	3	NE
<i>Ammoglanis amapaensis</i>	1	1	NE
<i>Elachocharax pulcher</i>	0	9	NE
<i>Fluviphylax palikur</i>	178	38	NE
<i>Hemigrammus aff. tridens</i>	0	14	NE
<i>Microcharacidium weitzmani</i>	108	56	NE
<i>Physopyxis ananas</i>	39	15	NE
<i>Potamoglanis hasemani</i>	43	5	NE
<i>Tyttobrycon</i> sp.	0	207	-



**FIGURE 2** | Accumulation (solid lines), extrapolation (dashed lines), and 95% confidence intervals (shaded areas) for diversity of miniature (red) and non-miniature (blue) freshwater fishes collected from the rio Jacundá, rio Amazonas basin, Brazil.





**FIGURE 3** | Non-metric multidimensional scaling (NMDS) plot illustrating 20 sampling locations (points) at upstream (blue boxes) and downstream (red circles) sampling sites. Miniature fish scores are shown as green text, including Aama (*Ammoglanis amapaensis*), Amin (*Ammocryptocharax minutus*), Epul (*Elachocharax pulcher*), Fpal (*Fluviphylax palikur*), Haft (*Hemigrammus aff. tridens*), Mwei (*Microcharacidium weitzmani*), Pana (*Physopyxis ananas*), Phas (*Potamoglanis hasemani*), and Tsp (*Tyttobrycon* sp.). Environmental and spatial correlates with miniature fish assemblage structure are shown as gray vectors and text, including water overhead canopy cover (%), water temperature (C), pH, total suspended solids (TSS, ppm), conductivity ( $\mu\text{s}/\text{cm}$ ), and the spatial relationship among sites from the first axis of an asymmetric eigenvector map (AEM1).

## DISCUSSION

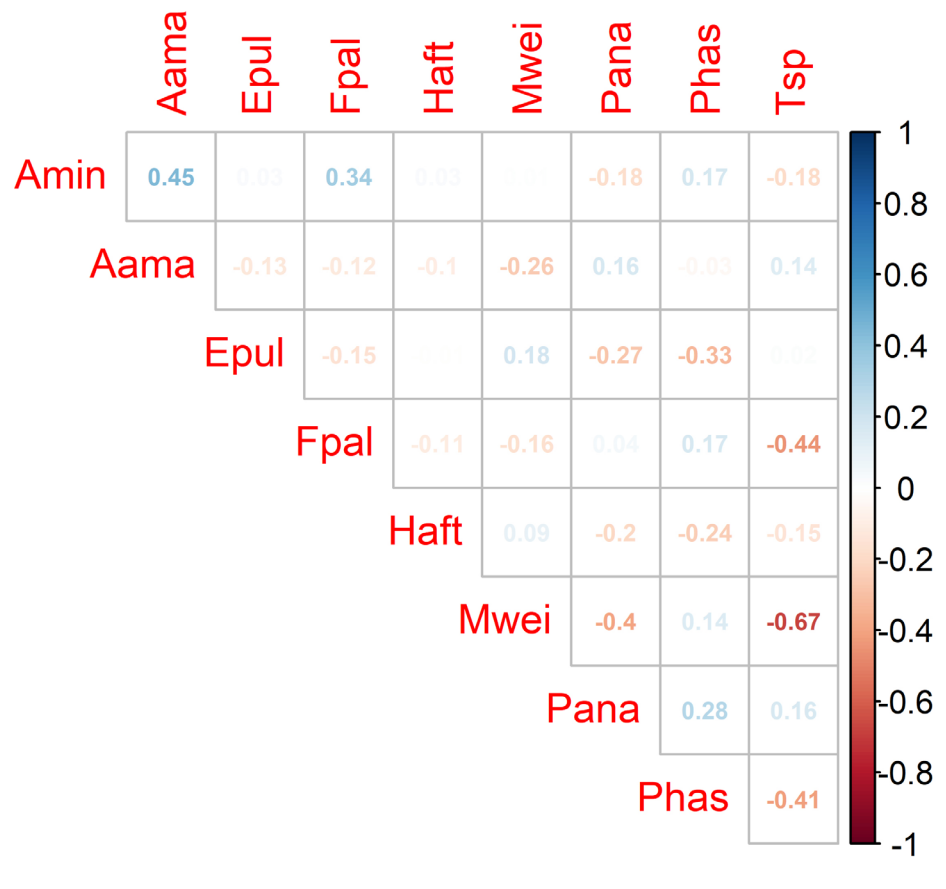
Few studies have quantitatively assessed miniature fish local richness or relationships with environmental variables. Henderson, Walker (1986, 1990) used hand nets to study the fish assemblages in litter banks of the small rio Tarumã-Mirim. Henderson, Walker (1990) used an early form of rarefaction to illustrate species accumulation across sampling efforts and challenged earlier notions that leaf litter banks were depauperate by collecting 20 species that apparently spatially segregated their habitats. However, no quantitative assessment of habitat was presented despite Henderson, Walker (1986) reporting water quality parameters. Carvalho *et al.* (2006) conducted behavioral observations of miniature fishes and noted the apparent use of transparency and camouflage to avoid predation, but the authors provided no quantitative assessment of habitats used by fishes. Carvalho *et al.* (2013) used fine-mesh hand nets to collect fishes from floating litter banks in four rivers of the Amazon basin. The authors used rarefaction analysis to compare richness across four river systems, but did not differentiate miniature fishes from juvenile forms of non-miniature fishes. Our work adds to this body of literature by revealing richness

specific to miniature fishes as well as local habitat variables that correlate with benthic assemblage structure.

We found that assemblage composition was not uniform between the two sampling sites. The upstream site was a nested subset of the downstream site, yet abundances were greater upstream among the majority of miniature species present at both sites. Although our inference is limited because of only two sampling sites, this pattern

**TABLE 3 |** Environmental variables that showed significant correlation with miniature fish assemblages in the non-metric multidimensional scaling plot.

Variable	NMDS1	NMDS2	R <sup>2</sup>	P-value
Temperature (°C)	0.03	-0.99	0.32	0.04
pH	0.99	0.02	0.33	0.03
Conductivity (µs/cm)	0.94	0.34	0.31	0.04
Total Suspended Solids (ppm)	0.94	0.32	0.29	0.04
Canopy Cover (%)	-0.37	-0.93	0.30	0.04
AEM1	0.89	-0.46	0.50	<0.01



**FIGURE 4 |** Pairwise Pearson correlations in species abundance across the 20 sampling locations showing positive correlations in blue and negative correlations in red. Species codes are Aama (*Ammoglanis amapaensis*), Amin (*Ammocryptocharax minutus*), Epul (*Elachocharax pulcher*), Fpal (*Fluviophylax palikur*), Haft (*Hemigrammus aff. tridens*), Mwei (*Microcharacidium weitzmani*), Pana (*Physopyxis ananas*), Phas (*Potamoglanis hasemani*) and Tsp (*Tyttobrycon* sp.).

should be explored in future research targeting miniature fishes across gradients of stream size, habitat types, and anthropogenic alterations. Our fine scale study uncovered two environmental variables that are likely important at broader scales. First, water pH strongly correlated with differences in assemblage structure between the two sites. Ríos-Villamizar *et al.* (2017) noted that pH increased in a downstream direction in the rio Purus, but pH decreased with greater accumulated upstream total deforestation. The environmental or ecological factors driving the evolution of miniaturization in Neotropical freshwater fishes have yet to be identified (Weitzman, Vari, 1988). Because miniature fishes are frequently encountered in still or slow-flowing waters with low pH, anthropogenic alterations to water pH caused by deforestation could alter miniature fish assemblages over ecological and evolutionary time scales. A second environmental variable that strongly correlated with miniature fish assemblage structure was percent canopy cover. Previous work established the link between deforestation and reductions in leaf litter density within streams (Carvalho, Uieda, 2010). The NMDS analysis showed a trajectory towards convergence in assemblage structure at sampling locations with greater canopy cover and lower pH (negative values along NMDS 1 and 2) regardless of upstream versus downstream sampling site identity. These locations were dominated by *Hemigrammus aff. tridens* and *Elachocharax pulcher* at the downstream site and *Microcharacidium weitzmani* at both upstream and downstream sites. Henderson, Walker (1990) noted that *E. pulcher* was most abundant at locations with newly fallen leaves that had not yet begun to decompose, suggesting a link between this species and high canopy cover. Our fine-scale results combined with patterns from previous studies suggest future research on miniature fishes across broader spatial scales and abiotic gradients should consider pH and canopy cover during hypothesis development.

Our results support previous suggestions that miniature fishes spatially partition habitats, a finding that highlights future research avenues. Henderson, Walker (1990) suggested miniature fishes in the litter banks of the rio Tarumã-Mirim spatially partitioned their habitats, perhaps as a means of reducing competition for limited resources. Our analyses of miniature fish assemblages advance this assertion in two ways. First, miniature fish species showed clear separation in multivariate space in our NMDS analysis, highlighting that different species dominated different sampling locations distributed across gradients of primarily pH and canopy cover. Olden *et al.* (2006) used functional traits of fishes to suggest that the filling of “vacant” multivariate space by non-native species was evidence of niche segregation that ultimately allowed for co-occurrence of native and non-native species in the Colorado River basin in North America. Hypotheses regarding niche segregation by miniature fishes could be tested as functional trait information becomes available in the future. Second, results from pairwise correlations quantifying co-occurrence showed generally low or negative similarities in species abundances among sampling locations. Strong negative relationships might be expected if local assemblages were structured by competition (*e.g.*, Peres-Neto, 2004), whereas strong positive relationships might be expected if facilitation structured assemblages (*e.g.*, Peoples, Frimpong, 2016). This is a promising avenue for future research as hypotheses could easily be tested in small mesocosms given the minute adult size of the species involved.

Recent evidence from fishes suggests that miniaturization may arise from an acceleration of development rate (Marinho, 2017) and/or truncated ontogeny (Britz,

Conway, 2009; Mattox *et al.*, 2016). Miniaturization of South American freshwater fishes is also hypothesized to arise in response to biotic factors such as enhanced survival (as species) associated with minute body size and year-round reproduction (Roberts, 1984). Thus, body size likely interacts with life history traits to track species-specific optima in tradeoffs among growth, mortality, and reproduction (Kozłowski, Gawelczyk, 2002). We suggest priority research areas pertaining to miniature fish ecology include life history traits, experiments involving competition, and tracking movements. Enhancing our understanding of the ecology of these understudied and overlooked fishes will benefit both basic research on their evolutionary history and applied research pertaining to conservation. The former is necessary to provide context to the repeated and widespread evolution of miniature body size in freshwater fishes (Rüber *et al.*, 2007), while the latter is important given impending and on-going changes to Amazonian streams inhabited by miniature fishes (*e.g.*, Winemiller *et al.*, 2016; Albert *et al.*, 2020).

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#### ETHICAL STATEMENTS

Fishes collected via seine and dip nets at each location were euthanized immediately in a solution of river water and clove oil (eugenol). Fish collections were conducted with permission from Sistema de Autorização e Informação em Biodiversidade (SISBio/MMA 45429–3 and 70186–1) and Texas A&M University – TAMU IACUC (IACUC 2019–0042).

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The authors declare no competing interests.

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