

The influence of environmental characteristics on the distribution of ciliates (Protozoa, Ciliophora) in an urban stream of southeast Brazil

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(With 8 figures)

Abstract

The aim of this research was to study the ciliated protozoa community at three sampling stations that receive different levels of domestic sewage along the São Pedro Stream in the municipality of Juiz de Fora, Minas Gerais, Brazil, in order to determine the influence of organic pollution on this community and to assess the feasibility of using ciliates as water quality indicators. Four physical-chemical parameters of the water samples were evaluated: dissolved oxygen concentration, electrical conductivity, pH and temperature. The sediment was obtained manually, using dredges with capacity of 300 mL, at each collection point. Point 1 was located in a rural region that receives a low sewage load, while Points 2 and 3 were located in populated regions receiving high sewage loads. We found 22 ciliate species, of which 18 are included in the saprobic system and are considered bioindicators. These showed beta-mesosaprobic environments at Point 1 and alfa-mesosaprobic to polisaprobic environments at Points 2 and 3. The low levels of dissolved oxygen and the high electrical conductivity values at Points 2 and 3, together with the strong similarity between the ciliate taxocenoses of these points and the weak similarity between Point 1 and the other two, confirm the high sewage loads received at the latter two points. The combination of the biological indicators and physical-chemical analyses therefore proved itself to be an efficient method of evaluating water quality, and has excellent potential to support decisions on the conservation of headwaters and recuperation of degraded environments in lotic systems.

Keywords: ciliated protozoa, organic pollution, water quality, watercourses.

Influência das características ambientais sobre distribuição dos ciliados (Protozoa, Ciliophora) em um córrego urbano no sudeste brasileiro

Resumo

O objetivo deste trabalho foi estudar a comunidade de protozoários ciliados em três estações amostrais, que recebem diferentes níveis de lançamento de esgoto doméstico, ao longo do córrego São Pedro, Juiz de Fora, MG, a fim de se determinar a influência da poluição orgânica sobre a composição e distribuição desta comunidade e avaliar a viabilidade da utilização dos ciliados como indicadores da qualidade da água. Foram mensurados quatro parâmetros físico-químicos da água amostrada: teor de oxigênio dissolvido, condutividade elétrica, pH e temperatura. Amostras do sedimento foram obtidas manualmente com o auxílio de dragas, com capacidade de 300 mL, em cada ponto de coleta. O ponto 1, localizado em uma região rural, recebe baixa carga de esgoto, enquanto os pontos 2 e 3, localizados em regiões com ampla ocupação humana, recebem altas cargas de esgoto. Foram registradas 22 espécies de ciliados, sendo que 18 estão incluídas no sistema sapróbio e são consideradas bioindicadoras de ambientes beta-mesossapróbio no ponto 1 e de alfa a polissapróbio nos pontos 2 e 3. Os baixos valores de oxigênio dissolvido e os altos valores de condutividade elétrica registrados nas estações 2 e 3, juntamente com a similaridade entre a taxocenose de ciliados destas estações e a baixa similaridade entre a estação 1 e as demais, confirmaram as altas cargas de esgoto recebidas nestas estações. A união do método biológico com a análise físico-química mostrou-se, portanto, um eficiente método na avaliação da qualidade da água, e apresenta grande potencial de utilização em tomadas de decisões relativas à conservação de nascentes e recuperação de ambientes degradados em sistemas lóticos.

Palavras-chave: águas correntes, poluição orgânica, protozoários ciliados, qualidade da água.

1. Introduction

Studies on pollution effects in freshwater environments were traditionally based on observing the physical and chemical characteristics of the water (Norris and Thoms, 1999). However, these parameters, isolated from the analysis of the biotic community, do not provide enough evidence to completely evaluate water quality. The current trend is to analyze not only changes in the physical-chemical characteristics, but also the responses to these changes by the organisms that live in these environments. Biological data, together with physical and chemical data represent an important tool to evaluate water quality in rivers and streams and have contributed to efforts to control emissions of organic pollutants in urban lotic systems (Suehiro and Tezuca, 1981; Grolière et al., 1990; Sparagano and Grolière, 1991; Madoni, 1993; Fernandez-Leborans and Novillo, 1996; Sola et al., 1996; Madoni and Bassanini, 1999; Madoni, 2005; Madoni and Braghiroli, 2007).

In this context, ciliated protozoa are very important because the disturbances caused by pollution can significantly alter the aquatic food chain, and thus the composition and abundance of these protozoa (Czapik, 1982; Primc, 1988; Primc-Habdija et al., 1998; Madoni and Bassanini, 1999; Madoni and Braghiroli, 2007).

The use of ciliated protozoa as bioindicators has advantages over the use of other organisms. The high sensitivity of these protists to changes in their surroundings, allied with their short generation time, enables them to reveal the response to environmental contamination much more quickly. Besides this, they are widely distributed geographically, being essential components of nearly all environments and can be easily maintained in the laboratory (Sparagano and Grolière, 1991; Piccinni and Gutiérrez, 1995; Fernandez-Leborans and Novillo, 1996).

Their great sensitivity to physical and chemical factors can be explained by the fact that many protozoa have specific demands in relation to the characteristics of the medium in which they live, such as the quantity of dissolved organic matter, temperature, pH, electric conductivity and dissolved oxygen concentration (Noland, 1925; Kudo, 1966; Sleight, 1988). Among these characteristics, the quantity of organic matter and dissolved oxygen in the water define pollution zones that are associated with particular species of protozoan indicators (Foissner and Berger, 1996). Four main zones of pollution are distinguished: polysaprobity (very heavily polluted), alfa-mesosaprobity (heavily polluted), beta-mesosaprobity (moderately polluted) and oligosaprobity (clean or very slightly polluted) (Streble and Krauter, 1987; Foissner and Berger, 1996).

In recent years many studies have examined the composition, distribution and dynamics of ciliate communities in lotic systems, as well as their value as water quality indicators (Madoni and Ghetti, 1981; Suehiro and Tezuca, 1981; Wiackowski, 1981; Czapik, 1982; Primc, 1988; Grolière et al., 1990; Sparagano and Grolière,

1991; Madoni, 1983; 1984; 1993; Fernandez-Leborans and Novillo, 1996; Sola et al., 1996; Primc-Habdija et al., 1998; Madoni and Bassanini, 1999; Madoni, 2005; Madoni and Braghiroli, 2007).

The present research studies the community of ciliated protozoa at points with different raw domestic sewage levels along São Pedro Stream (Southeast Brazil) to determine the influence of organic pollution on the composition and distribution of this community and to evaluate the use of ciliates as water quality indicators.

2. Material and Methods

2.1. Study area and sampling stations

The watershed of São Pedro Stream, located in the southwest part of the urban area of the municipality of Juiz de Fora, Minas Gerais, Brazil (Figure 1), includes São Pedro Dam, which is responsible for 9% of the city's water supply. The streamlet itself extends for 13,250 m (watercourse) and 10,750 m (straight line), traversing various districts of the municipality. Its headwaters are located at an altitude of 875 m above sea level (Latuf, 2004).

The collections were done monthly, always in the morning, at three points (Figure 1) along the stream, to obtain samples with different levels of untreated sewage. Point 1 (23K, 663036, 7590303) is located in a rural region and receives a low sewage load, while Points 2 (23K, 668307, 7591772) and 3 (23K, 668645, 7592804) are located in considerably populated regions, receiving high sewage loads.

At each station, samples were collected monthly from August 2002 to June 2003. Benthic ciliates were collected from the substrate by using dredges with capacity of 300 mL, at each collection point. At each station three samples were collected along the line of the river cross-section and near the border. Glass jars were lowered to the bottom, recovered and capped. The samples were then put in thermal containers to maintain their physical conditions and taken to the laboratory.

2.2. Physical and chemical parameters

Monitoring of the physical and chemical properties of the water at the sampling points was done with portable equipment at the collection sites, recording the following parameters: dissolved oxygen concentration, electrical conductivity, pH and water temperature.

2.3. Processing the samples and identifying the ciliates

Of the samples taken to the laboratory, we placed five fractions of approximately 20 mL in Petri dishes to observe the organisms under a stereoscopic microscope with transmitted light, on the day of collection. The ciliates was sorted from these portions with micropipettes (made in the laboratory) and observed in vivo (Tuffrau, 1959) until eight hours after their collection, under an optical microscope (bright field and phase contrast) in order to make preliminary identifica-

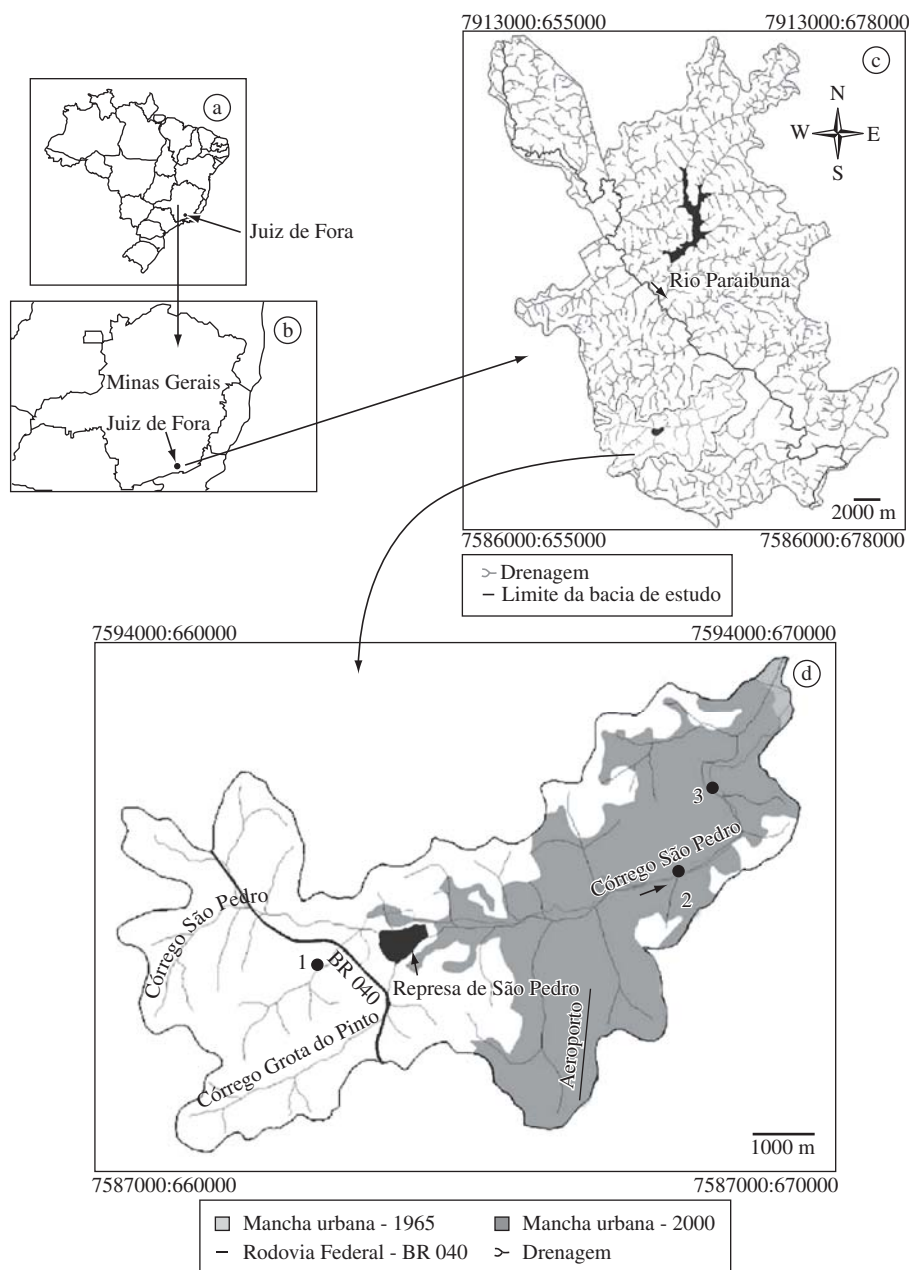


Figure 1. Map of the São Pedro streamlet showing the sampling stations (P1 = Station 1; P2 = Station 2; P3 = Station 3) in the urban area of Juiz de Fora, Minas Gerais, Brazil (from Latuf, 2004).

tions. Observation of live ciliates is of great taxonomic importance in classifying genera and species (Foissner and Berger, 1996). Cultures were prepared to carry out further identification techniques, adding crushed rice in fractions of other samples. The rice grains served as a source of carbon for the bacteria present, encouraging excystment of the ciliates. The following silver impregnation techniques were carried out: protargol (Dieckmann, 1995) and “dry” silver (Klein, 1958; Foissner, 1991).

2.4. Statistical treatment

The Mann-Whitney nonparametric test was used to verify differences between the mean values of the physical and chemical parameters among the three collection stations.

Cluster analysis, with Euclidian distance and UPGMA, was used to examine the similarity among these parameters at the three collection points, as well as the data showing presence or absence of the ciliate community at the points. To determine the degree of similar-

ity among the ciliate communities of the three sampling stations, the qualitative similarity indices of Jaccard and Sorensen (Magurran, 1988) was calculated.

3. Results and Discussion

3.1. Physical and chemical analysis

Dissolved oxygen is one of the most important gases in the dynamics and characterization of aquatic ecosystems (Esteves, 1988; Von Sperling, 1996). The low average oxygen values obtained at Points 2 (3.02 mg.L⁻¹) and 3 (2.04 mg.L⁻¹), when compared with the high average values at Point 1 (6.49 mg.L⁻¹) (Figure 2), indicate high organic pollution loads at the stations in the urban area (Points 2 and 3). There was a significant difference (Mann-Whitney test, $p < 0.05$) between the mean O₂ values from Point 1 in relation to Points 2 and 3, and no significant difference (Mann-Whitney test, $p < 0.05$) between the mean O₂ values of Points 2 and 3.

The electrical conductivity of a solution is considered an important variable in characterizing water bodies and detecting sources of pollution (Esteves, 1988). The low conductivity levels at Point 1 and high values at Points 2 and 3 confirm the high sewage loads received at the latter two collection sites (Figure 3). There was a significant difference (Mann-Whitney test, $p < 0.05$) between the mean electrical conductivity values from Point 1 (78.6 $\mu\text{S}\cdot\text{cm}^{-1}$) in relation to the sampling stations 2 (286.4 $\mu\text{S}\cdot\text{cm}^{-1}$) and 3 (293.0 $\mu\text{S}\cdot\text{cm}^{-1}$), and no significant difference (Mann-Whitney test, $p > 0.05$) between these average values from Points 2 and 3.

The water temperature was nearly the same at all three collection points: 1 (22.4 °C), 2 (22.5 °C) and 3 (23.2 °C), with no significant difference (Mann-Whitney test, $p > 0.05$) among them (Figure 4).

Point 2 (7.41) and Point 3 (7.46) had lower pH values than Point 1 (8.11) (Figure 5). In lakes where there is an upper aerobic layer and another anaerobic layer at

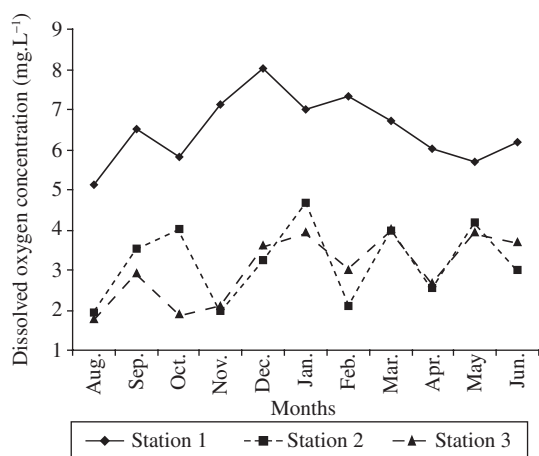


Figure 2. Variation of dissolved oxygen concentration from three points along São Pedro stream, from August 2002 to June 2003.

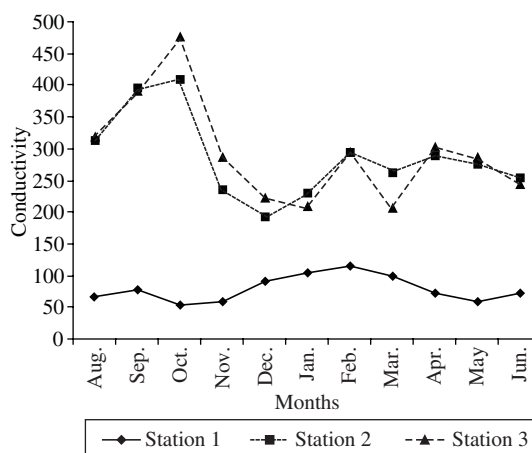


Figure 3. Variation of conductivity from three points along São Pedro stream, from August 2002 to June 2003.

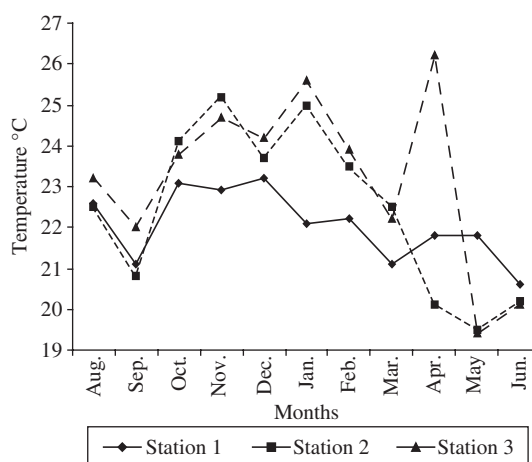


Figure 4. Variation of temperature from three points along São Pedro stream, from August 2002 to June 2003.

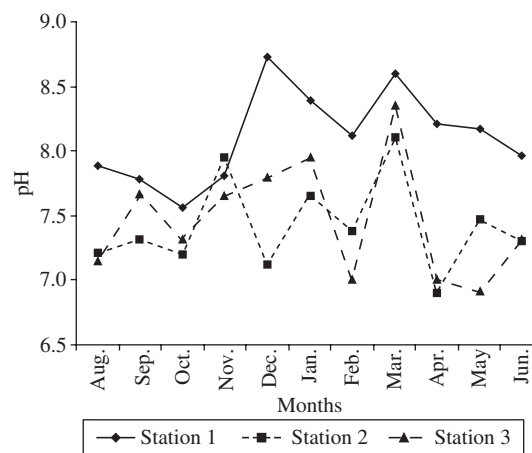


Figure 5. Variation of pH from three points along São Pedro stream, from August 2002 to June 2003.

the bottom, anaerobias can cause elevation of the pH as a consequence of photosynthetic action. It is interesting to note that phenomena like those that keep the pH slightly base in polluted waters can also be occurring in lotic systems, such as in the São Pedro Stream, in stretches that receive high domestic sewage loads. These effluents flow in constantly, forming a layer of activated mud on the bed. This layer can be washed out during extremely rainy periods, but it accumulates again in short order (Branco, 1986). There was a significant difference (Mann-Whitney test, $p < 0.05$) between the mean pH values for Point 1 in relation to Points 2 and 3, and no significant difference (Mann-Whitney test, $p < 0.05$) between these mean values for Points 2 and 3.

The similarity between collection stations 2 and 3, regarding physical and chemical parameters, is shown in Figure 6.

3.2. Biological analysis

Table 1 presents the ciliated protozoa taxa that were identified from São Pedro Stream during the study and the occurrence frequency of each species at the three sample stations along eleven months sampling. A total amount of 22 species of ciliated protozoa belonging to 20 genera were identified. The species *Coleps* sp., *Spirostomum minus* Roux, 1901, *Frontonia leucas* (Ehrenberg, 1833), *Urocentrum turbo* Ehrenberg, 1838, *Stylonychia pustulata* Müller, 1786 and *Vorticella* sp. occurred at all the sampling stations. The species occurring at only one station were: *Halteria* sp. at Station 1, *Podophrya fixa* Müller, 1786 at Station 2; and *Stentor coeruleus* Ehrenberg, 1830, *Tokophrya lenarum* Stein, 1859 and *Blepharisma sinuosum* Sawaya, 1940 at Station 3. Of the 22 taxa recorded, 17 occurred at Stations 2 and 3.

The smaller number of species registered at Station 1 (6 species) in comparison with Stations 2 (18 species) and 3 (20 species) (Table 1) can be related to the low abundance of species typical of clear water and headwaters, making them harder to find in samples collected from that station. Foissner (1997), studying the ciliate fauna of four clear water rivers in Germany, attributed the low abundance of ciliates encountered to a possible shortage of nutrients in these environments.

The degree of similarity among the ciliate taxocenoses at the three sampling stations studied is shown

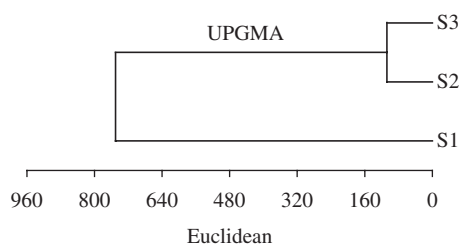


Figure 6. Dendrogram of cluster analysis, showing the similarities among stations (S1, S2 and S3) on the basis of physical and chemical qualities.

in Table 2 and Figure 7. As expected, there was a close similarity between Stations 2 and 3 and low similarity between these two and Station 1.

The species *Stentor coeruleus* and *Blepharisma sinuosum* were recorded only in the cultures kept in the laboratory (Table 1), because they possibly were encysted in the samples analyzed on the collection day, due to the unfavorable environmental conditions. In the cultures, with favorable food supply, the excystment of these species occurred. However, it was not possible to investigate these species in the ecological study because they had not been registered on the collection day. The excystment mechanism is important because it permits the ciliated protozoan populations to resist predation, unfavorable physical and chemical conditions and food shortages (Taylor, 1983).

The saprobic system for evaluating water quality, and more specifically for assessing organic pollution, developed by Kolkwitz and Marsson (1908, 1909) and Kolkwitz (1950), is widely used in biological classification of flowing waters. The original lists of indicator species has been revised and extended by various authors (Sladeček, 1973; Sladeček et al., 1981; Foissner, 1988, Foissner et al., 1995), who have added data on the water quality and revised taxonomic errors (Paiva and Silva-Neto, 2004).

Among the ciliate species found in São Pedro Stream, 18 are included in the saprobic system and are considered bioindicators: *Caenomorpha* sp., *Lagynus* sp., *Loxodes striatus* (Eugelmann, 1862) Penard, 1917, *Metopus* sp., *Paramecium caudatum* Ehrenberg, 1838, *Plagiopyla* sp. and *Spirostomum teres* Clapèrè et Lachmann, 1958 are indicators of polysaprobic environments; *Euplotes eurystomus* Wrzesmowski, 1870, *Podophrya fixa*, *Stentor coeruleus* and *Tokophrya lenarum* of alpha-mesosaprobic environments; *Coleps* sp., *Spirostomum minus* and *Stentor polymorphus* Ehrenberg, 1830 of alpha-mesosaprobic to beta-mesosaprobic environments; *Frontonia leucas*, *Urocentrum turbo* and *Stylonychia pustulata* of beta-mesosaprobic environments; and *Halteria* sp., which indicates oligosaprobic to alpha-mesosaprobic environments (Table 1).

Figure 8 shows the presence of indicator species of alpha-mesosaprobic and polysaprobic environments at Stations 2 and 3 and the occurrence of an indicator species of an oligosaprobic environment only at Station 1. Among the eight species included in the saprobic system that occurred exclusively at Stations 2 and 3 (Table 1), only *E. eurystomus* is an indicator of an alpha-mesosaprobic environment. The others are indicators of a polysaprobic environment. These saprobity levels correspond to class III and IV waters, with low or no oxygen content, characterizing polytrophic environments, meaning extremely polluted (Streble and Krauter, 1987).

At Station 1, where six ciliate species occurred, three of them indicate a beta-mesosaprobic environment, whereas the other two indicate alpha to beta-mesosaprobic environments (Table 1). Thus, this site has a moderate to

Table 1. List of ciliate species at three stations (S1, S2, S3), with number of positive samples throughout the 11 collecting months, in which each species was found, saprobity levels (S) and main food source (Mf).

Protist species	S1	S2	S3	S*	Mf
<i>Caenomorpha</i> sp.	-	5	5	p-i	Ba, Sb
<i>Lagynus</i> sp.	-	1	3	p-i	O
<i>Loxodes striatus</i> (Engelmann, 1862)	-	3	4	p	Al, Ki, Cy
<i>Metopus</i> sp.	-	5	5	p-i	Ba, Al, Fl
<i>Paramecium caudatum</i> Ehrenberg, 1833	-	11	11	p-a	Ba, Al
<i>Plagiopyla</i> sp.	-	4	2	p-i	Ba, Sb, Al, Fl
<i>Spirostomum teres</i> Claparède and Lachmann, 1858	-	5	5	p	Ba, Sb, Al, Ki
<i>Euplotes eurystomus</i> Wrzesniowski, 1870	-	3	4	a	O
<i>Podophrya fixa</i> (Müller, 1786)	-	1	-	a	R
<i>Stentor coeruleus</i> (Pallas, 1766)	-	-	+	a	O
<i>Tokophrya lenarum</i> (Stein, 1859)	-	-	1	a	R
<i>Coleps</i> sp.	3	4	5	a-b	-
<i>Spirostomum minus</i> Roux, 1901	5	6	6	a-b	Ba
<i>Stentor polymorphus</i> (Müller, 1773)	-	3	1	a-b	Ba, Al, Ki
<i>Frontonia leucas</i> Ehrenberg, 1833	8	5	5	b	O
<i>Urocentrum turbo</i> Müller, 1786	4	7	9	b	Ba, Ki
<i>Stylonychia pustulata</i> (Müller, 1786)	2	7	7	b	O
<i>Halteria</i> sp.	2	-	-	o-a	Al
<i>Blepharisma sinuosum</i> Sawaya, 1940	-	-	+	-	-
<i>Epistylis</i> sp.	-	3	1	-	-
<i>Prorodon</i> sp.	-	2	3	-	-
<i>Vorticella</i> sp.	1	5	9	-	-

p = polisaprobity; i = isosaprobity; a = alfa-mesosaprobity; b = beta-mesosaprobity; o = oligosaprobity; Al = algae; Ba = bacteria; Cy = cyanobacteria; Fl = heterotrophic flagellates; Ki = diatoms; O = omnivorous; R = predator; Sb = sulphur bacteria. + = species recorded only in the cultures kept in the laboratory. *Foissner and Berger (1996).

Table 2. Qualitative similarity indices of Jaccard and Sorensen between the station pairs 1 and 2 (S1-S2), 1 and 3 (S1-S3) and 2 and 3 (S2-S3).

Stations	Qualitative similarity index	
	Jaccard	Sorensen
S1-S2	0.32	0.48
S1-S3	0.29	0.44
S2-S3	0.81	0.89

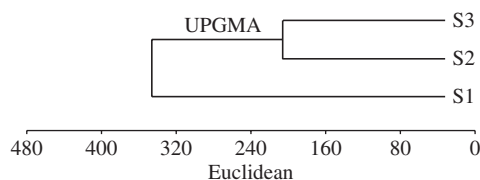


Figure 7. Dendrogram of cluster analysis, showing the similarities among stations (S1, S2 and S3) on the basis of ciliate taxocenoses.

critical grade of pollution. Among the species that occurred at all the collection stations (Table 1), *S. minus*, *F. leucas* and *U. turbo* have been reported along rivers with different saprobic levels (Wiackowski, 1981;

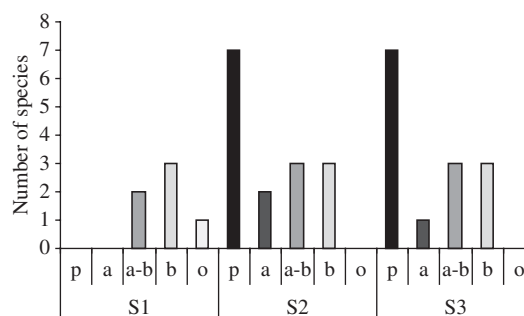


Figure 8. The number of species indicators of polisaprobity (p), alfa-mesosaprobity (a), alfa to beta-mesosaprobity (a-b), betamesosaprobity (b) and oligosaprobity (o) environments at the three stations (S1, S2 and S3).

Czapik, 1982; Madoni and Ghetti, 1982; Sparagano and Grolière, 1991; Sola et al., 1996; Madoni and Bassanini, 1999), which shows the ample ecological valence that is characteristic of this species.

Several authors, studying rivers in France (Sparagano and Grolière, 1991), Spain (Sola et al., 1996), Italy (Madoni and Ghetti, 1981; Madoni, 1993; Madoni and

Bassanini, 1999; Madoni, 2005; Madoni and Braghiroli, 2007) and Poland (Wiackowski, 1981), have demonstrated an increased proportion of polysaprobic and alpha-mesosaprobic species and a decrease in beta-mesosaprobic and oligosaprobic species with increasing pollution loads caused by anthropic impacts. Madoni (1993) demonstrated the efficiency of using ciliate protists as water quality indicators in the Parma River, comparing data collected in 1978, when the system received untreated sewage, and in 1989, after a treatment station was operating. The marked change in the ciliate community was accompanied by an increase in organic load, which caused the disappearance of some typical species and dampened the diversifying effects of other environmental factors, producing comparatively similar community structures in that whole zone.

The species *Coleps* sp., *Epistylis* sp. and *Vorticella* sp. have not been classified as to saprobity because they have only been identified at the genus level. On the lists of indicators contained in Foissner (1988) and Foissner and Berger (1996), the genera *Coleps* Nitzsch, 1817, *Epistylis* Ehrenberg, 1830 and *Vorticella* Linnaeus, 1767 indicate environments that vary from oligosaprobic to alpha-mesosaprobic. The presence of *Epistylis* sp. only at Stations 2 and 3 and the greater frequency of *Vorticella* sp. at these stations as well (Table 1), corroborate that peritrichous ciliates are strongly related to organic pollution (Antipa, 1967; Burbank and Spoon, 1967; Henebry and Ridgeway, 1979).

The indicator species of polysaprobic environments recorded at Stations 2 and 3, located in the urban area of São Pedro Stream, have been reported by various authors in eutropic lakes (Finlay, 1981, 1982; Laybourn-Parry et al., 1990), at sewage treatment stations (Salvadó et al., 1995) and in polluted rivers and streams (Madoni and Ghetti, 1981; Suehiro and Tezuca, 1981; Wiackowski, 1981; Czapik, 1982; Sola et al., 1996; Madoni and Bassanini, 1999; Madoni, 2005; Madoni and Braghiroli, 2007).

In his guide to freshwater protozoa, Patterson (1996) classified the species of the genera *Metopus* (Claparède et Lachman, 1850), *Caenomorpha* Perty, 1852, *Plagiopyla* Stein, 1860, *Spirostomum* Ehrenberg, 1833 and the species *L. striatus* as anoxic benthonic ciliates. Foissner and Berger (1996) included *Caenomorpha* spp., *Metopus* spp., *Plagiopyla nassuta*, *Loxodes* spp. and *Lagynus elegans* in a community called "Metopetum", composed mainly of strictly anaerobic organisms, which do not have mitochondria and present sulphurous bacteria as symbionts, being "infallible" indicators (Czapik, 1982) of the presence of sulphurated hydrogen in the water. Heterotrophic sulphurous bacteria occur principally at the surface of sediment, where the reducing conditions are more favorable, and appear in great densities in highly polluted environments (Esteves, 1988).

The availability of food is an important biotic factor that controls the distribution of ciliated protozoan popu-

lations in the various ecosystems (Noland, 1925; Sleigh, 1988). Several studies have shown the effect of pollution on the trophic relationships of the community of ciliated protozoa in flowing waters (Wiackowski, 1981; Czapik, 1982; Primc, 1988; Primc-Habdija et al., 1998; Madoni and Bassanini, 1999; Madoni, 2005; Madoni and Braghiroli, 2007).

In oligosaprobic waters, ciliates mainly feed on algae, while only a few feed on bacteria and cyanobacteria. The trophic relationships are significantly altered when the level of saprobity increases. These changes start with a decrease in the number of algivorous ciliates and an increase in the number of bacterivorous species (Wiackowski, 1981; Czapik, 1982; Madoni and Bassanini, 1999; Madoni, 2005). Organic pollutants cause an increase in phosphates and other nutrients, altering the structure of bacteria communities, inducing changes in the ciliate fauna, which depend directly on these bacteria for food (Primc, 1988).

Among the species that occurred at all the collection stations (Table 1), only *S. minus* is exclusively bacterivorous. *F. leucas* and *S. pustulata* are omnivorous and *U. turbo* can feed on both bacteria and diatoms. The species of the genus *Coleps* are mostly omnivorous and some species of the genus *Vorticella* can feed on algae and bacteria (Foissner and Berger, 1996). The omnivorous feeding habits of some ciliate species enable them to occur both in headwaters and polluted stretches of rivers and streams, as demonstrated in various studies (Czapik, 1982; Sparagano and Grolière, 1991; Packroff and Zwick 1996; Sola et al. 1996; Madoni and Bassanini 1999). In the present work, *Halteria* sp., which occurred only at Station 1, is algivorous, indicating an environment with little impact. This corroborates the data obtained by Czapik (1982), who recorded species with this feeding habit at well preserved spots along a river polluted by domestic sewage.

Among the ten species recorded exclusively at Stations 2 and 3 (Table 1), 70% are indicators of polysaprobic environments, and the majority preferentially bacterivorous. Among the bacterivorous ciliates, the most specialized are those that feed on sulphurous bacteria, indicating highly degraded environments (Czapik, 1982). The species *Caenomorpha* sp., *Lagynus* sp., *Metopus* sp. and *Plagiopyla* sp. registered in the present paper are examples of this group of ciliates.

The study of the physical and chemical characteristics of São Pedro Stream, such as low levels of dissolved oxygen and high electrical conductivity values recorded at Stations 2 and 3 (Figures 2 and 3), as well as the high similarity among the ciliate taxocenoses at these stations and the low similarity between Station 1 and the other two (Figure 7), confirm the high sewage loads at these stations, along with the high sensitivity of the ciliate populations to changes in organic pollution and the saprobic condition in running waters.

The combination of the biological method (ciliate community) with the traditional method (physical and

chemical data), therefore, is an efficient way to evaluate water quality and has great potential for use in reaching decisions on the preservation and recuperation of degraded lotic systems.

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