

NITROGEN LOSS BY EROSION FROM MECHANICALLY TILLED AND UNTILLED SOIL UNDER SUCCESSIVE SIMULATED RAINFALLS

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ABSTRACT

The description of the fate of fertilizer-derived nitrogen (N) in agricultural systems is an essential tool to enhance management practices that maximize nutrient use by crops and minimize losses. Soil erosion causes loss of nutrients such as N, causing negative effects on surface and ground water quality, aside from losses in agricultural productivity by soil depletion. Studies correlating the percentage of fertilizer-derived N (FDN) with soil erosion rates and the factors involved in this process are scarce. The losses of soil and fertilizer-derived N by water erosion in soil under conventional tillage and no tillage under different rainfall intensities were quantified, identifying the intervening factors that increase loss. The experiment was carried out on plots (3.5 × 11 m) with two treatments and three replications, under simulated rainfall. The treatments consisted of soil with and soil without tillage. Three successive rainfalls were applied in intervals of 24 h, at intensities of 30 mm/h, 30 mm/h and 70 mm/h. The applied N fertilizer was isotopically labeled (¹⁵N) and incorporated into the soil in a line perpendicular to the plot length. Tillage absence resulted in higher soil losses and higher total nitrogen losses (TN) by erosion induced by the rainfalls. The FDN losses followed another pattern, since FDN contributions were highest from tilled plots, even when soil and TN losses were lowest, i.e., the smaller the amount of eroded sediment, the greater the percentage of FDN associated with these. Rain intensity did not affect the FDN loss, and losses were greatest after less intense rainfalls in both treatments.

Keywords: ¹⁵N, rainfall simulation, water erosion.

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RESUMO: PERDA DE NITROGÊNIO POR EROSÃO EM SOLOS MOBILIZADO POR PREPARO MECÂNICO E SEM MOBILIZAÇÃO, SOB CHUVAS SIMULADAS SUCESSIVAS

A caracterização do destino do N proveniente do fertilizante em um sistema é uma ferramenta essencial para a melhoria de práticas de manejo que visem a máxima utilização do nutriente pela cultura e o mínimo de perdas. A erosão do solo ocasiona perda de nutrientes, como o N, causando impactos negativos na qualidade das águas de mananciais superficiais e subterrâneas, além da queda de produtividade agrícola em razão do empobrecimento do solo. Estudos que correlacionam o percentual de N proveniente do fertilizante com as taxas de erosão do solo, bem como os fatores intervenientes nesse processo, são escassos. Nesse sentido, foram quantificadas as perdas de solo e N provenientes do fertilizante por erosão hídrica em solo mobilizado pelo preparo convencional e solo sem mobilização, sob diferentes intensidades de chuva, identificando os fatores que interferem no aumento dessa perda. O experimento foi conduzido em parcelas experimentais de 3,5 m de largura e 11 m de comprimento, com dois tratamentos e três repetições, sob chuva simulada. Os tratamentos utilizados foram solos com mobilização e sem mobilização. Foram aplicadas três chuvas sucessivas, com intervalo de 24 h entre elas, com intensidades de 30 mm/h, 30 mm/h e 70 mm/h. O fertilizante nitrogenado utilizado foi marcado isotopicamente (^{15}N) e incorporado ao solo em linha transversal ao comprimento das parcelas. A ausência de preparo do solo condiciona as maiores perdas de solo por erosão e maiores perdas de nitrogênio total (NT) ao longo das chuvas aplicadas. As perdas do nitrogênio proveniente do fertilizante (NPF) seguiram padrão distinto; as maiores perdas de NPF foram verificadas nas parcelas com mobilização do solo, mesmo que tenham apresentado as menores perdas de solo e NT; ou seja, quanto menor a quantidade de sedimentos erodidos maior o percentual de NPF associado a esses. As chuvas de maior intensidade não influenciaram a perda de NPF; as maiores perdas ocorreram em decorrência das chuvas de menor intensidade nos dois tratamentos utilizados.

Palavras-chave: ^{15}N , chuva simulada, erosão hídrica.

INTRODUCTION

Soil erosion leads to nutrient losses (Favaretto, 2002; Bertol et al. 2003; Guadagnin, 2003; Guadagnin et al, 2005), with potentially negative impacts on water quality of surface and ground water sources, as well as on the air (Follett and Walker, 1989; Antweiler et al., 1996; Favaretto, 2002; Delgado et al., 2008). Losses of soil and associated elements can damage the environment, e.g., by siltation and eutrophication of waterways, and affect farmers and society by economic losses (Cardoso et al., 2012).

Runoff is responsible for the contamination of water bodies, by the transfer of colloidal sediments containing nutrients at generally high concentrations. The amount of transported nutrients is influenced by the soil management system. In general, nutrient concentrations in runoff are higher in systems of soil conservation management, while total nutrient losses are higher in conventional systems (Guadagnin et al., 2005).

Water erosion from soil is influenced by the rain, soil type, topography, soil cover and management, and by conservation practices. Among these factors, soil cover and management have most influence on water erosion (Cogo, 1981); the influence of soil management was described in numerous studies (Cassol and Lima, 2003; Cogo et al., 2003; Bertol et al., 2007; Panachuki et al., 2011).

According to Cogo et al. (2003), although a series of studies showed the effectiveness of conservation tillage in the control of erosion, water losses are extremely variable and may sometimes be higher in no-tillage, sometimes in reduced tillage systems, or sometimes under conventional tillage, or even similar between the different methods of soil tillage methods, determined by conditions such as rain intensity, soil type, topography and crop sequence/rotation used in farm management systems.

Runoff is the most important climatic factor for erosion, with a harmful action; it occurs when the stability established over time between soil, vegetation and soil organisms is upset, usually by human intervention (Mehl, 2000). Beutler et al. (2006) claimed that rain and runoff together are the main active agents of water erosion. The combination of these agents with terrain-related factors determines the intensity of soil degradation.

Among the essential nutrients for high yields in tropical agriculture, nitrogen is one of the most important (Döbereiner, 1997), with the highest crop requirement and the most exported by crops, being the first nutrient to become deficient in a system. This justifies the need to study the influence of the different factors that influence the loss of this nutrient by water erosion in detail.

The characterization of fertilizer N in agricultural systems is an essential tool for improving management practices aimed at maximum nutrient exploitation by crops and minimal losses (Fenilli et al., 2008).

The use of isotope-enriched fertilizer (^{15}N) is an excellent tool for this type of research (Reichardt and Bacchi, 2004).

Many studies using this tool address the percentage of fertilizer-derived N (FDN) taken up by crops and/or, FDN losses in runoff water (Carranca et al., 1999; Choi et al., 2003; Bertol et al., 2005; Mead et al., 2008; Giacomini et al., 2010; Woodward et al., 2012; Vallano and Sparks, 2013). Studies correlating the percentage of FDN with soil erosion rates and with the factors involved in this process are scarce.

In this context, the study objective was to quantify losses of soil and fertilizer-derived N by water erosion from tilled and untilled soil under different rainfall intensities, identifying the intervening factors that increase loss.

MATERIAL AND METHODS

Description of experimental plots

This field study was conducted in an experimental area of Sudeste (Fazenda Canchim), UTM coordinates 206219, 7569671, in São Carlos, São Paulo. The soil at the experimental site was classified as *Latossolo Vermelho-Amarelo álico*, according to the Brazilian System of Soil Classification - SiBCS (Embrapa, 2013). The horizon has moderate, medium texture and 3 % slope, under long-standing use in experiments with no-tillage oat.

The experimental area was structured in downslope plots (width 3.5 m \times length 11 m). Each plot was delimited at the far ends by plywood boards (width 0.20 m, driven halfway into the soil). At the lower end a hole was spared, to which a PVC pipe was fixed to channel the runoff to a 1.0 m³ collector box.

Two treatments were used: plots with tilled soil (TS) (plowing and disking) and with untilled soil (US), with three replications. The soil was prepared in mid-February, and in both treatments the soil was maintained completely bare during the experiment, removing even the residues of the previous crop. The experiment was conducted between February 13 and 15, 2007. The replications were considered in this way since all plots were statistically similar in terms of soil physical characteristics prior to the experiment (Table 1).

Rain was applied with a rotating boom simulator (Lombardi Neto et al., 1979), covering two plots simultaneously. Three successive rainfalls, in 24 h intervals, were simulated at the following intensities: 1st rain 30 mm/h, rain 2nd 30 mm/h and 3rd 70 mm/h. The total precipitation of the first two rainfalls was 6 mm and of the last 12 mm, corresponding to a total of 24 mm.

Application of ^{15}N

Each plot was fertilized with a rate of 50 kg ha⁻¹ N as ammonium sulfate [(NH₄)₂SO₄], with isotopic concentration of 5.09 % of ^{15}N atoms, 24 h before the first simulated rain. The quantity and form of application followed the concept of conventional application of [(NH₄)₂SO₄] in sugarcane cultivation, applied in rows spaced 0.90 m apart. In this way, 13.5 g of fertilizer was applied by hand to a depth of 3 cm, in a 3.5 m wide strip (Figure 1).

Sample collection and analysis

^{15}N in the soil: Samples were collected in metal cylinders after the end of the experiment, i.e., after simulation of the three rainfalls. The cylinders were divided into representative samples of the 0-5 and 5-10 cm layer in the fertilizer application line. The remaining samples were collected only from the 0-5 cm layer, at distances of 3 and 6 m from the fertilizer application line (Figure 1).

Table 1. Soil physical characteristics prior to the experiment (surface layer of 10 cm)

Plot	Clay	Silt	Sand	Moisture	Bd ⁽¹⁾
Untilled soil					
1	33.80	7.60	58.60	0.24	1.59
3	33.20	7.30	59.50	0.25	1.50
5	28.00	8.10	63.80	0.24	1.50
Tilled soil					
2	31.90	15.50	52.60	0.25	1.55
4	30.80	9.60	59.60	0.23	1.55
6	28.80	6.40	64.80	0.22	1.79

⁽¹⁾ Bd: bulk density. Source: Bramorski et al. (2012).

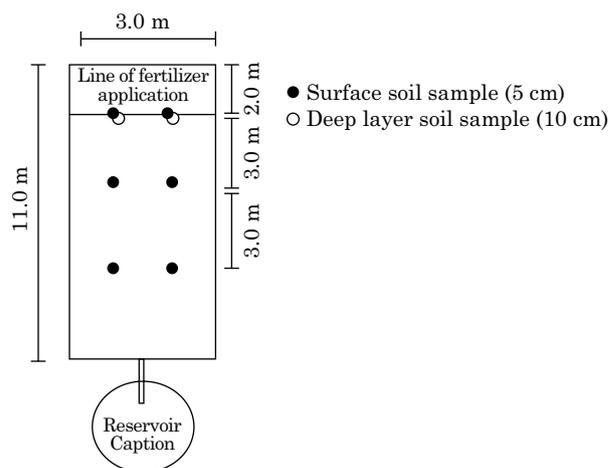


Figure 1. Diagram of the fertilizer application line and soil sampling.

¹⁵N in sediment and water: after each simulated rainfall event the material of surface runoff (water + sediment) was collected.

The samples were filtered and an aliquot of each resulting solution was subjected to Kjeldahl digestion, as described by Bremner and Mulvaney (1982), to determine N concentration. The sediment retained on the filter was dried at 60 °C for 24 h and then ground to determine total N and ¹⁵N abundance.

Nitrogen-15 abundance was determined by mass spectrometry (Hydra 20-20, coupled to an ANCA-GSL automatic analyzer, of SerCon Co., Crewe, UK).

Determination of fertilizer-derived N (FDN%)

The percentage of FDN and quantity of fertilizer-derived N (QFDN) for each sample type (soil-sediment and water) were calculated based on IAEA (1983).

The FDN% was calculated as the relationship between the percentage of excess ¹⁵N atoms in the sample (atom% ¹⁵N exc. samp.) and the percentage of excess ¹⁵N atoms in the fertilizer (atom% ¹⁵N exc. fert.):

$$FDN\% = \frac{(\% \text{ }^{15}\text{N exc. samp.})}{(\% \text{ }^{15}\text{N exc. fert.})} \times 100 \quad \text{Eq. 1}$$

The QFDN (kg ha⁻¹ N) was calculated from the result of FDN% and the total amount of N in the sample (QTN, kg ha⁻¹ N):

$$QFDN = \frac{(\% \text{ FDN})}{100} \times QTN \quad \text{Eq. 2}$$

The data were subjected to analysis of variance MANOVA (version 2.2) to detect statistically significant differences between the treatments and also between the ranges analyzed in each plot. Cluster analysis was applied in the ranges observed within plots.

RESULTS AND DISCUSSION

Distribution of fertilizer-derived nitrogen (FDN) in the soil

After simulation of cumulative rains of 24 mm, the FDN that was not transferred in the soil in relation to the application in Untilled soil (US) varied from 1.94 to 16.63 %. In Tilled soil (TS), in this variation was between 0.26 and 15.16 %.

The FDN carried up to 3 m away from the application range ranged from 0.10 to 2.21 % in US and 0.33 to 1.10 % in TS. In the uttermost position, corresponding to 6 m away from the tracer application, the values were between 0.25 and 1.29 % in US and 0.08 and 0.61 % in TS. The MANOVA analysis of

variance indicated significant differences in FDN transfer from application in the line, according to each treatment, in the three replications (p-value between 0.042 and 0.049) (Table 2).

Similarity groups were formed (Figure 2) for each distance, consisting of plots with the same tillage type, most evident at a distance of 3 m from the fertilizer line.

The plots 5 and 6, consisting of US and TS respectively, had a clearly different pattern from the others. The formation of a similarity group between these plots was observed in the fertilizer line, as shown above (Figure 2a). These plots had a lower capacity of retaining FDN in the soil matrix, as demonstrated by the low concentration in the fertilizer line, with smaller decreases in the values with increasing distance from the line of application.

In the other plots, aside from the higher FDN concentration in the fertilizer line, the values found at a distance of 6 m, in US, were around 13 and 49 times lower than those in the fertilizer line, while in TS they were around 24 and 100 times lower, indicating this type of soil tillage can delay the horizontal FDN transfer throughout the soil surface. Schick et al. (2000) demonstrated that in untilled soil, nutrients are accumulated in the soil surface, resulting from the surface application, or in shallow depth, of the soil fertilizer. According to Cogo (1981), in soils tilled conventionally along the full extension and to considerably deeper depths, surface roughness instantly becomes high, and sediment mobility during erosion is therefore

Table 2. Percentage of fertilizer-derived N (%FDN) recovered from the fertilization line in Untilled (P1, P3 and P5) and Tilled plots (P2, P4 and P6)

Distance from the fertilizer line	Untilled soil		Tilled soil	
	Plot 1		Plot 2	
m	%FDN			
0	16.63		15.16	
3	2.21		1.11	
6	1.29		0.62	
	p=0.042 ⁽¹⁾			
	Plot 3		Plot 4	
0	12.38		14.18	
3	0.12		0.33	
6	0.25		0.14	
	p=0.049			
	Plot 5		Plot 6	
0	1.94		0.26	
3	0.11		0.37	
6	0.43		0.09	
	p=0.043			

⁽¹⁾ Value of p in relation to the level of significance.

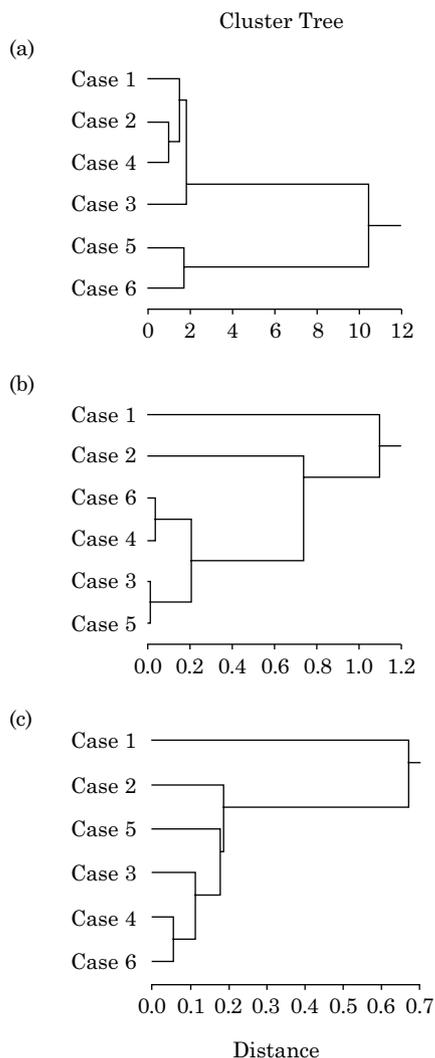


Figure 2. Grouping of the values found in the fertilizer line (a), at a distance of 3 m from the fertilizer line (b) and at a distance of 6 m from the fertilizer line (c), in all plots.

reduced. Giacomini et al. (2010) reported similar findings, i.e., a higher percentage of ^{15}N in plots with minimum tillage than under conventional tillage.

In general, the concentrations of FDN transferred vertically in the soil were higher in deeper layers than those in the soil surface in the line of the tracer fertilizer. In the US plots (Figure 3), the FDN concentrations were higher in the plots than in deeper layers TS (Figure 3).

The plots 5 and 6 (US and TS, respectively) apparently had distinct pattern, with high vertical FDN transfer in the soil profile, as also observed before for the horizontal transfer.

Loss of FDN in surface runoff

In relation to sediment yield at the end of each rain, the highest intensity induced highest losses

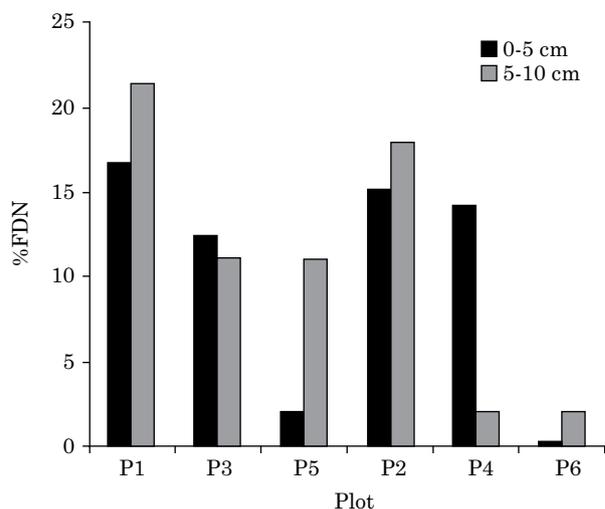


Figure 3. Percentage of fertilizer-derived nitrogen (%FDN) in the 0-5 and 5-10 cm layers in Untilled plots (P1, P3 and P5) and Tilled plots (P2, P4 and P6).

from all plots, independent of the treatment, with exception of plot 2 (TS) (Table 3).

The rain intensity led to higher TN loss from all plots and soil tillage did not influence the losses, analyzing each rain separately. Berg et al. (1988) also stated an increase in N losses with increase in runoff volume, in tilled (conventional) as well as in untilled soil (no-tillage). Analyzing the total losses, the US plots tended to highest TN losses, as also stated by Giacomini et al. (2010) in plots under minimum tillage (without mechanical soil tillage), compared to plots under conventional tillage.

It can be concluded that the higher the soil loss, the higher is the TN loss.

However, when analyzing the fertilizer-derived N quantity (FDN), the FDN values were highest in TS plots. In other words, no direct relation between erosion rate and FDN transport was observed.

Bertol et al. (2003) stated that the increase in nutrient loss rates is generally inversely proportional to soil loss. Alberts and Moldenhauer (1981) explained that as erosion decreases, the proportion of minor particles in the eroded material increases. That is, the sediments become nutrient-richer.

Bellanger et al. (2004) also claimed that bare, untilled soils tend to lose more coarser particles, compared to cultivated or simply tilled soils. The coarsest sediments have a lower capacity of N transfer, which is mainly linked to the finer sediments.

The results confirmed these statements (Table 3). In general, the FDN concentrations decreased with each rainfall, whereas the quantity of eroded sediment increased.

Table 3. Total eroded sediment, quantity of total nitrogen (TN) and fertilizer-derived N (FDN) and percentage of FDN in relation to TN (%FDN) in each treatment, at the end of each rain, in Untilled and Tilled soils

Rain intensity mm/h	Total eroded sediment			TN			FDN			FDN		
				g						%		
	Replication											
	1	2	3	1	2	3	1	2	3	1	2	3
	Untilled soil											
30	10.28	7.20	1.50	0.024	0.017	0.005	0.00043	0.00030	0.00018	1.80	1.79	3.38
30	11.70	9.60	2.40	0.036	0.029	0.001	0.00106	0.00022	0.00002	2.95	0.76	1.87
70	79.80	38.00	15.60	0.211	0.133	0.058	0.00034	0.00030	0.00015	0.16	0.23	0.26
Total	101.78	54.80	19.50	0.271	0.179	0.064	0.00183	0.00082	0.00035	0.68	0.46	0.55
	Tilled soil											
30	6.90	0.78	1.28	0.016	0.002	0.007	0.00062	0.00010	0.00089	3.88	4.27	12.83
30	7.36	1.35	6.90	0.027	0.004	0.023	0.00091	0.00010	0.00007	3.43	2.68	0.31
70	5.94	13.77	28.05	0.028	0.027	0.110	0.00005	0.00002	0.00045	0.24	0.08	0.41
Total	20.20	15.90	36.23	0.071	0.033	0.140	0.00158	0.00022	0.00141	2.33	0.67	1.01

Loss of fertilizer-derived N (FDN) in runoff water

The low pressure of (N₂) gas in the mass spectrometer to determine total N and ¹⁵N abundance of the organic and mineral forms in the water (nitrate + nitrite + ammonium) revealed a low concentration of the element.

Within the scope of this experiment, water, as an important pathway of FDN loss in dissolved form, was not addressed in particular.

Recovery of fertilizer-derived nitrogen (FDN) in the system

The absence of crops in this experiment led to the high FDN concentrations in the soil (16.63 % in US and 15.16 % in TS, in the fertilizer line and 0-5 cm layer and 21.36 and 17.85 % in US and 5.49 % in TS in the 5-10 cm layer). Basanta et al. (2003) found only about 10 % FDN incorporated in soil under sugarcane cultivation, after 128 mm rain in 100 days. Fenilli et al. (2008) found only 12.6 % FDN in soil under coffee.

Several authors (Giacomini et al., 2010; Thomsen and Christensen, 2007; Rieger et al., 2008), compared the dynamics of N fertilizers in the soil-plant system under conventional and minimum tillage (without mechanical soil tillage) and stated no significant differences between the tillage methods, in terms of N plant uptake.

However, the dynamics of soil N and its soil-sediment relations (after erosion) show that the tillage methods induce different situations.

The highest retention of FDN in soil, in the fertilizer line and in the deeper layers was observed in the US plots. In these plots, the horizontal FDN transfer along the soil profile was also highest, and

the losses associated to sediments were lowest, even though the soil losses by erosion were highest.

This could indicate that under this tillage type, most soil was lost in another way than by leaving the system via eroded sediment. Indices of higher FDN leaching to deeper soil layers, to beneath the 10 cm studied here, were observed. In addition, soil was possibly lost in runoff water. However, although the runoff volume was highest from these plots in all rain simulations, no FDN concentration was detected in the water, which may be the result of the low precipitation volume or of the volatilization of the more soluble N after the first rain.

With regard to FDN, exportation by erosion was more efficiently minimized from untilled soils, even though this efficiency was not observed in terms of soil loss.

Guadagnin et al. (2005) stated that systems of soil conservation management were more efficient than those of conventional management in reducing losses of soil, water and mineral N. To some extent, this confirms our study results, with exception of soil losses, which were higher from US, most likely due to the absence of a residue cover, which was not the case in the study of the said authors.

CONCLUSIONS

The absence of soil tillage led to higher soil erosion losses by the applied rainfalls, and to higher losses of total nitrogen, found mainly in organic form, associated to the finer soil minerals in the soil.

The losses of fertilizer-derived N were however higher from plots with tilled soil, indicating that the lost, highly soluble fertilizer, is not directly associated with the quantity of eroded sediment.

The rain intensity influenced soil and TN losses; the most intense rainfall induced highest losses, in US as well as in TS. However, rain intensity had no influence on FDN losses; losses were highest under the least intense rainfalls in both applied treatments.

The plots with untilled soil retained FDN more efficiently, horizontally as well as vertically and, less efficiently in terms of minimizing total soil and nitrogen loss.

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