

## Severe population decline of marsh deer, *Blastocerus dichotomus* (Cetartiodactyla: Cervidae), a threatened species, caused by flooding related to a hydroelectric power plant

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**ABSTRACT.** Balancing power production and environmental conservation can be problematic. The objective of this study was to investigate the abundance of marsh deer in the Paraná River Basin, above the Sergio Motta (Porto Primavera) Dam, before and after the impact of the dam closure. A fixed-wing, flat window aircraft was used to survey study transects. Observations were recorded based on the distance sampling line transect method, assuming that the detection probability decreases with increased distance. The abundance of marsh deer in the survey region prior to flooding was estimated to be 974 individuals (CV = 0.23). The overall abundance dropped from 974 to 444 (CV = 0.26) individuals after flooding, an overall reduction of 54%. This reduction can be attributed to the direct impact of the flooding process, but it was likely exacerbated by indirect effects, such as increased disease, hunting, and reduction in food availability. Prior to flooding, the marsh deer was distributed widely throughout the dam’s catchment area; however, the marsh deer habitat was almost completely destroyed by the flooding process. This situation highlights the need to implement management strategies that ensure the survival of the remaining fragmented population.

**KEY WORDS.** Abundance; anthropogenic impact; cervids; distance sampling; distribution.

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Concerns regarding wildlife conservation increased dramatically during the twentieth century because of the extinction, or near-extinction, of several species, which was caused by human activities (e.g., hunting and fishing, and habitat degradation) or natural events (e.g., catastrophic events and disease outbreaks).

The production of hydroelectric power depends on artificially built water reservoirs. The resulting inundation can cause a major loss of habitat near the source river and around the reservoir itself (BAXTER & GLAUDE 1980, REITAN & THINGSTAD 1999, ELLIS *et al.* 2009). Only a few studies have documented the direct effect of artificial flooding on wildlife, including investigations of the effects on birds (REITAN & THINGSTAD 1999, ELLIS *et al.* 2009) and mammals (SZABÓ *et al.* 2003). The loss or degradation of habitat due to reservoir inundation can cause a decline in animal populations and species richness (GALETTI *et*

*al.* 2010). The extent to which inundations affect wildlife habitats is dependent upon the timing, severity, and length of the inundation (WARREN & TURNER 1975, REITAN & THINGSTAD 1999).

The Sergio Motta (Porto Primavera) Dam (22°25’S, 52°58’W) is located in the Paraná River basin, between the states of São Paulo and Mato Grosso do Sul, Brazil. With the closing of the sluices, an estimated 200,000 ha of land were inundated, thus dramatically reducing the floodplain habitat for many species. One species that utilizes floodplain areas for habitat is the marsh deer, *Blastocerus dichotomus* Illiger, 1815; this species lives in lowland areas and has been threatened with extinction (DUARTE *et al.* 2008, PIOVEZAN *et al.* 2010). The population considered in this study was constrained between the Sergio Motta and Jupia dams; other than the habitat flooded by the lake, these animals do not have suitable alternative habitat in the region.

The marsh deer is the largest South American deer. Originally much more widely distributed throughout South America (NOWAK & PARADISO 1991), the marsh deer now occurs in east-central and northeastern Argentina, west-central and southern Brazil, Paraguay, southeastern Peru, and eastern Bolivia. The species has been extirpated from Uruguay (PIOVEZAN *et al.* 2010). Populations' sizes are declining throughout the species' range due to excessive hunting and conversion of wetlands for agriculture, tree plantations, and dams. In Brazil, hydroelectric plant projects have eliminated floodplain habitats along many large rivers, including the Tiete, Paraná, and Rio Grande, and cattle ranching has severely reduced and fragmented floodplain habitats (MÁRQUEZ *et al.* 2006, DUARTE *et al.* 2008). The marsh deer is a marsh-obligated animal that is integrated with marsh vegetation ecology. The marsh environment is dynamic and recycles naturally when hydrologic and geomorphic features are balanced.

Aerial surveys of marsh deer have been conducted since 1976 (SCHALLER & VASCONCELOS 1978, PINDER 1994, 1996, MAURO *et al.* 1995, MOURÃO & CAMPOS 1995, ANDRIOLO *et al.* 2001, 2005). PINDER (1994) estimated the Brazilian population of marsh deer to be 50,950 individuals, with only 1,300 animals in protected areas.

Aerial surveys using fixed-wing aircraft and following a line transect distance sampling methodology were conducted specifically to estimate marsh deer abundance because marsh deer tend to occupy flat low lands and are widely distributed on floodplain areas (ANDRIOLO *et al.* 2001, 2005, 2010a).

The formation of lakes associated with hydroelectric dams causes significant environmental change at various ecological levels. However, there is widespread ignorance concerning the negative effects of hydroelectric plants on wildlife, especially deer species. The potential impact of increasing hydroelectric energy development has prompted concern amongst conservationists, scientists and managers about the future of the Brazilian marsh deer population. Our objective was to investigate the marsh deer abundance and distribution in the Paraná River Basin above the Sergio Motta (Porto Primavera) Dam, before and after dam closure, to provide information that can support future strategies for species conservation and management.

## MATERIAL AND METHODS

The Paraná River flows southward from central Brazil to Paraguay and Argentina and is the site of two major hydroelectric power plants (Itaipu and Jupiá). The surveyed area corresponds to the stretch of the Paraná River between the Sergio Motta (Porto Primavera) Dam in the south and the Jupiá Dam in the north. The surveyed area was divided into seven blocks: Aguapeí, Peixe, Lagoa SP, Paraná N, Pardo, Paraná M and Paraná S (Fig. 1). The study area is not entirely homogeneous and the northern area is more affected by human activities.

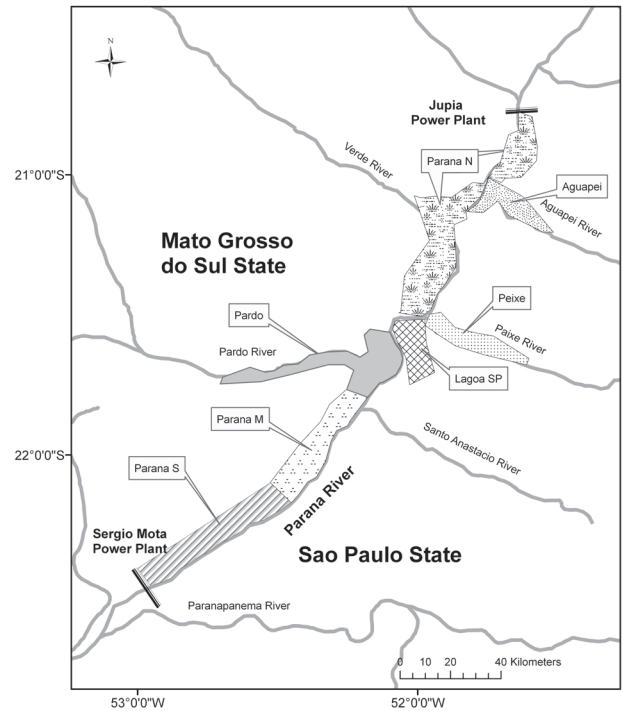


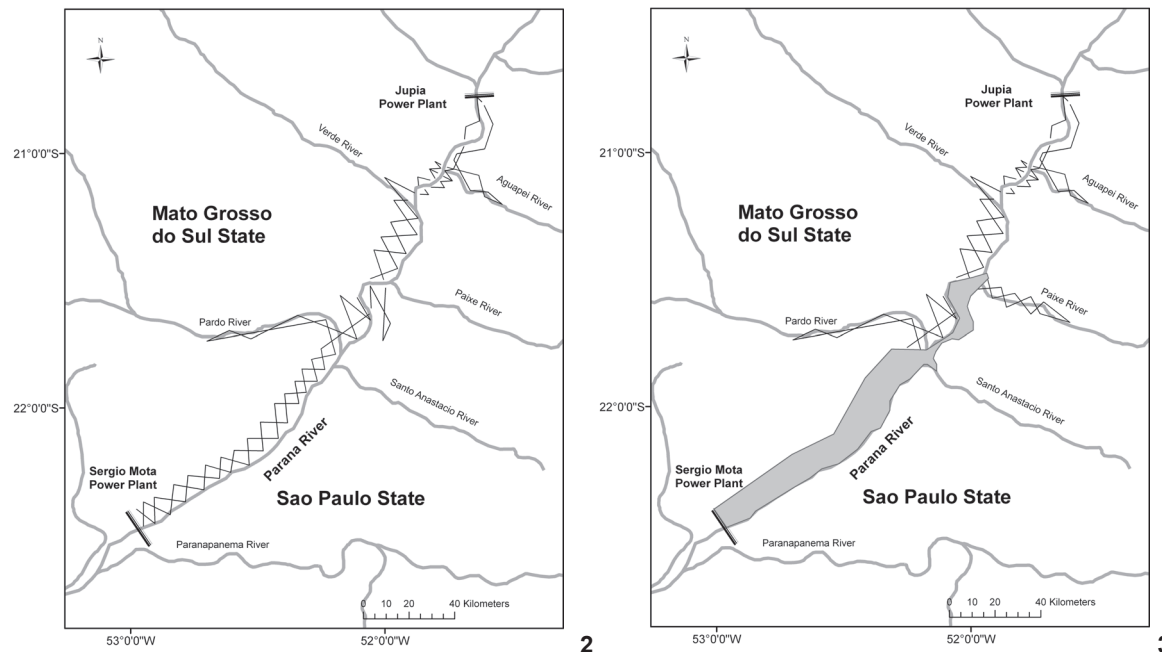
Figure 1. The study area was divided into seven blocks: Aguapeí, Peixe, Lagoa SP, Paraná N, Pardo, Paraná M and Paraná S.

The Paraná riverine plains are up to 10 km wide in Mato Grosso do Sul (Paraná N, Pardo, Paraná M and Paraná S), while the discontinuous floodplains of São Paulo (Aguapeí, Peixe, Lagoa SP) do not exceed 2 km in width. The plains immediately adjacent to the river are 50 m above sea level and are designated as 'lowland floodplains'. Farther from the river, a more ancient sedimentation plain is found a few meters higher. Lowland floodplains consist primarily of marshes and grasslands that are seasonally flooded in January and February. Several forest islands are interspersed in these marshlands above the flood line and parallel to the river. Bordering the marshlands are slopes that have been transformed into pasture lands (PINDER 1996).

The total floodplain area is approximately 200,000 ha. Initially, a total of 76 transects were designed to systematically cover the area along the basin. The transects were randomly and evenly distributed in the irregular area of the floodplain (Fig. 2).

The strategy of dam closure divided the lake filling process into two phases: the first began on November 28, 1998, filling approximately 75% (253 m cote) of the lake and the second began on February 2, 2001, completing the filling process (257 m cote).

The study was divided into a training phase (May 1998) and four subsequent surveys (September 1998, February 1999, November 1999, and December 2000). The training was performed between May 19 and 22, 1998 to train the pilot, observers and recorders and to familiarize the research crew with



Figures 2-3. Transects flown along the Paraná River basin in the aerial line-transect survey conducted before flooding (2) in September 1998, and after flooding (3) in February 1999, November 1999, and December 2000. The grey polygon represents the area affected by lake formation.

the aerial survey methodology. The crew flew over a total of 62 transects, encompassing 742 km in length, in which 55 animals were sighted. The data obtained during the training phase (May 1998) were not included in the analysis.

The first survey (September 1998) preceded the closing of the gates of the dam and was conducted between September 6 and 11, 1998 (Fig. 2). The second survey (February 1999) was performed between February 25 and 26, 1999, after the first lake-filling phase (November 1998). After the flooding, transects in the southern area (Paraná S and Paraná M) were no longer accessible. The remaining transects in the northern area were surveyed (Aguapeí, Lagoa SP, Paraná N and Pardo), as was a newly added area (Peixe block, Fig. 3).

The third survey (November 1999) occurred between 16 and 17 November 1999 and the fourth survey (December 2000) was carried out between 5 and 7 December, 2000. Table I presents the details of the survey regions and their sizes, as well as the area covered, the number of transects and the sampling effort (km) for each of the surveys used in the analysis.

A fixed-wing, flat window Cessna aircraft (model 182 in the training survey and model 172 in the subsequent surveys), equipped to carry four people, was used to survey the designated transects. The pilot and the data recorder occupied the front seats and the two observers occupied the back seats. The crew flew only on clear days, with cloud cover ranging from 0-40%, from 0900 to 1400 at 70 m in altitude and at an airspeed of 160 km/h.

Table I. Summary of the sample regions surveyed in different years, including area, number of transects (N), area covered and effort.

Sample date	Region	Area (km <sup>2</sup> )	N	Area covered (km <sup>2</sup> )	Effort (km)
Sep-98	Aguapeí	89.0	7	23.91	56.40
	Lagoa SP	144.7	4	26.08	61.50
	Paraná M	566.0	16	68.77	162.20
	Paraná N	699.3	28	132.08	311.50
	Paraná S	330.5	14	53.42	126.00
	Pardo	179.6	7	48.12	113.50
<b>Total</b>		<b>2,009.1</b>	<b>76</b>	<b>352.38</b>	<b>831.10</b>
Feb-99	Aguapeí	89.0	7	21.43	50.54
	Paraná N	699.3	27	101.32	238.96
	Pardo	179.6	9	40.85	96.35
	Peixe	88.2	15	25.18	59.39
<b>Total</b>		<b>1,056.1</b>	<b>58</b>	<b>188.78</b>	<b>445.24</b>
Nov-99	Aguapeí	89.0	8	25.59	60.36
	Paraná N	699.3	28	104.71	246.95
	Pardo	179.6	8	37.15	87.61
	Peixe	88.2	10	23.05	54.36
<b>Total</b>		<b>1,056.1</b>	<b>54</b>	<b>190.50</b>	<b>449.28</b>
Dec-00	Aguapeí	89.0	8	26.44	62.35
	Paraná N	699.3	25	93.82	221.28
	Pardo	179.6	7	36.18	85.32
	Peixe	88.2	10	21.43	50.55
<b>Total</b>		<b>1,056.1</b>	<b>50</b>	<b>177.87</b>	<b>419.5</b>
<b>Total over all sample periods</b>		<b>5,177.2</b>	<b>238</b>	<b>909.53</b>	<b>2,145.12</b>

The search effort was suspended at the end of each transect to allow the plane to circle before the start of the next transect. Visual confirmations of animal occurrences were recorded with the corresponding GPS coordinates and the vertical declination was registered using a handheld clinometer (where the horizon is set to 0 degrees and 90 degrees is directly below the aircraft) for subsequent distance calculations. There was a blind area below the plane due to the flat window. The sex of the animals and other information were recorded whenever possible. The recordings were made according to the distance sampling line transect method (BURNHAM *et al.* 1980, BUCKLAND *et al.* 1993), with the assumption that visibility decreases with distance.

Perpendicular distances were calculated from the aircraft's altitude and the declination angle to the sighting. The distances of animals from the transect line could not be accurately calculated for declinations greater than or equal to 60 degrees; thus, only observations of animals at declinations of 55 degrees or less were considered to ensure data quality (GUENZEL 1994, 1997). This value corresponded to 49 m on each side of the transect line and was subtracted from the calculated distance for each observation [ $g(49) = g(0) = 1$ ]. The declination was also subdivided into bands (e.g., ANDRIOLO *et al.* 2006, 2010b) of 10-degree cluster distances, which yielded intervals of 55, 45, 35, 25, 15 and 5 degrees, corresponding to distances of 0, 21, 51, 101, 212 and 751 m, respectively. The truncation was performed at 212 m.

The following strategies were considered in the study: 1) Modeled ungrouped perpendicular distance data truncated at 212 m; 2) Modeled grouped perpendicular distance data truncated at 212 m (ANDRIOLO *et al.* 2005).

The detection probability was assumed to be the same as the calculated probability of ANDRIOLO *et al.* (2005). The analysis incorporated  $g(0) = 0.71$  as a correction for missing animals and observer losses.

Abundance was estimated according to standard line-transect methods (BURNHAM *et al.* 1980, BUCKLAND *et al.* 1993). Data were analyzed using the R software (R CORE TEAM 2012) with the MCDS package (LAAKE *et al.* 2011).

The effects of covariates were incorporated into the detection function model by setting the scale parameter in the model to be an exponential function of the covariates. One advantage of this method is that it prevents sample size reduction due to stratification and therefore minimizes bias and a loss of precision. This method also allowed several covariates to be incorporated into the estimation procedure (MARQUES & BUCKLAND 2004).

The proposed models were the half normal and hazard rate models with the following covariate possibilities: 1) no covariate; 2) group size covariate – to investigate the effect of group size on the detection probability; 3) region covariate – to investigate the effect of differences in the region on the detection probability; and 4) period covariate – to investigate the effect of period (dry/wet) on the detection probability.

The only plausible models were options 1 and 4 above. Models incorporating the group size covariate were counter-intuitive (e.g., the detection probability decreased as distance from the line increased) and models with the region covariate needed to be evaluated in relation to the possible differences across areas.

The abundance variation was tested for significance using the Delta Method (SEBER 1982). The four surveys provided the following abundance estimates: Aa, Ab, Ac and Ad. The goal was to obtain A (the difference between the Aa estimate and the average of the Ab, Ac and Ad estimates) and  $\text{var}(A)$  (POWELL 2007).

## RESULTS

A total 134 sightings were performed and 153 individuals were observed on effort and considered in the analysis. Table II presents the model selections based on the Akaike Information Criterion (AIC) values. There was essentially no difference in the estimates using the half normal or hazard rate methods. The model parameter and abundance estimates were also similar for the two models. The number of sightings, the encounter rates and the number of individuals are summarized in Table III.

Table II. Summary of model selection statistics for the pooled marsh deer data.

Model	npar	AIC	DeltaAIC	Weight	p	CV
Hazard rate	2	371.2992	0	0.367963	0.50	0.15
Half normal	1	371.6936	0.394425	0.302104	0.53	0.08
Hazard rate + period covariate	3	372.5857	1.286507	0.193394	0.49	0.15
Half normal + period covariate	2	373.2819	1.982748	0.136539	0.53	0.08

Number of parameters (npar); Akaike Information Criterion (AIC); probability of detection (p); coefficient of variation (CV).

Figure 4 presents the distribution of perpendicular distances and the fit of the detection function. Based on the AIC values, the hazard-rate model best fit the perpendicular distance data. Table IV presents the abundance estimates for grouped (4 bins) data truncated at 212 m.

The overall abundance dropped from 974 individuals before the flooding to 395 (February 1999), 502 (November 1999) and 444 (December 2000) individuals after the flooding, with a reduction of 54% from the first, pre-flooding estimate to the last estimate. This reduction was statistically significant [ $A = 527$ ,  $CV = 0.45$ ,  $SE = 236$ ;  $\text{Var}(A) = 55,693.19$ ,  $LCL = 64$  and  $UCL = 989$ ]. Although there was a severe and statistically significant reduction in abundance, the density did not vary significantly.

Figures 5 and 6 provide a visual comparison of the distribution before and after the flooding.

Table III. Parameters estimated based on the surveys of marsh deer along the Paraná Basin in different years.

Sample date	Region	s	ERs	SE (ERs)	n	ERn	SE (ERn)
Sep-98	Aguapeí	13	0.23	0.09	16	0.28	0.13
	Lagoa SP	1	0.02	0.02	1	0.02	0.02
	Paraná M	10	0.06	0.03	10	0.06	0.03
	Paraná N	11	0.04	0.01	11	0.04	0.01
	Paraná S	17	0.13	0.04	18	0.14	0.04
	Pardo	6	0.05	0.02	7	0.06	0.03
Total		58			63		0.03
Feb-99	Aguapeí	2	0.04	0.03	2	0.04	0.02
	Paraná N	13	0.05	0.02	14	0.06	0.01
	Pardo	3	0.03	0.01	3	0.03	0.09
	Peixe	5	0.08	0.08	6	0.10	0.09
Total		23			25		0.02
Nov-99	Aguapeí	10	0.17	0.07	11	0.18	0.04
	Paraná N	14	0.06	0.02	15	0.06	0.03
	Pardo	4	0.05	0.04	4	0.05	0.04
	Peixe	4	0.07	0.03	5	0.09	0.02
Total		32			35		0.03
Dec-00	Aguapeí	3	0.05	0.02	6	0.10	0.08
	Paraná N	9	0.04	0.01	10	0.05	0.01
	Pardo	4	0.05	0.02	5	0.06	1.25
	Peixe	5	0.10	0.05	9	0.18	1.80
Total		21			30		

Number of sighting (s); encounter rate (ER); standard error (SE), and number of individuals (n).

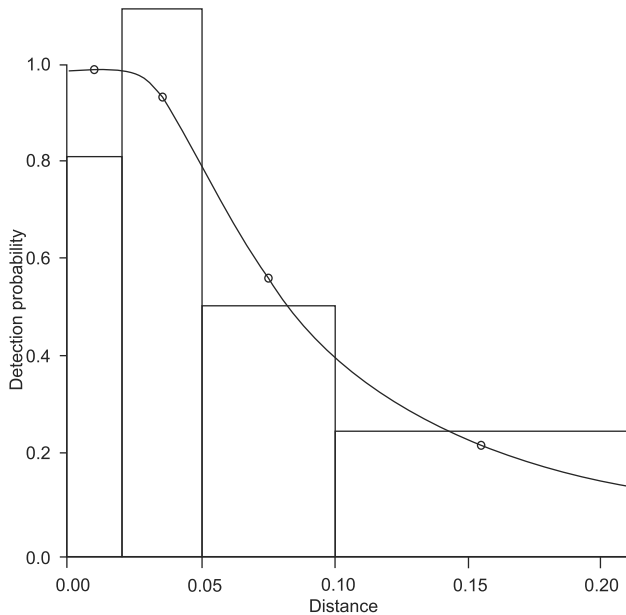


Figure 4. Frequency distribution of the sightings of marsh deer. The continuous curve represents the best fit function (hazard rate). The pooled distances are presented in kilometers.

Table IV. Marsh deer abundance estimates by year for the different survey areas.

Sample date	Region	D	A	CV	95%	CI
Sep-98	Aguapeí	1.90	169	0.48	69	416
	Lagoa SP	0.11	16	0.99	3	80
	Paraná M	0.41	234	0.51	90	604
	Paraná N	0.24	165	0.37	82	333
	Paraná S	0.96	316	0.37	157	635
	Pardo	0.41	74	0.47	31	179
Total		0.48	974	0.23	621	1530
Feb-99	Aguapeí	0.27	24	0.68	7	79
	Paraná N	0.39	274	0.38	134	560
	Pardo	0.21	37	0.45	16	87
	Peixe	0.68	60	0.88	14	262
Total		0.37	395	0.31	217	718
Nov-99	Aguapeí	1.22	109	0.52	41	286
	Paraná N	0.41	284	0.40	134	606
	Pardo	0.31	55	0.85	13	232
	Peixe	0.61	54	0.42	25	119
Total		0.47	502	0.29	287	879
Dec-00	Aguapeí	0.64	57	0.47	24	137
	Paraná N	0.30	212	0.40	100	449
	Pardo	0.39	70	0.47	29	169
	Peixe	1.19	105	0.47	44	253
Total		0.42	444	0.26	269	736

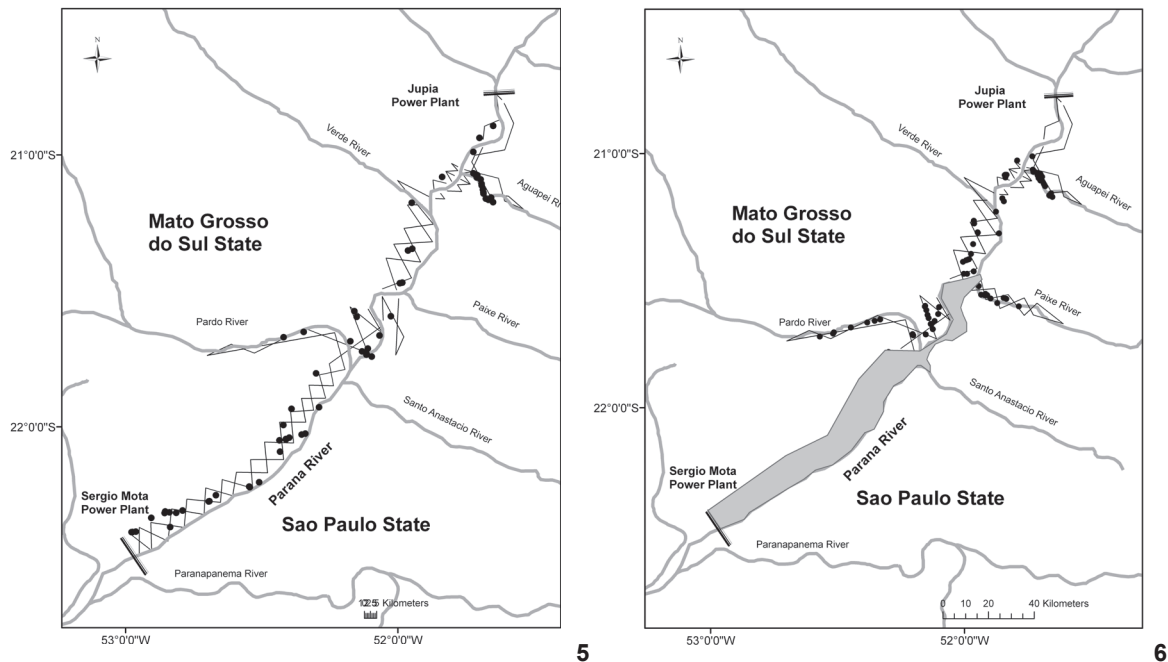
Density (D); abundance (A); coefficient of variation (CV); confidence interval (CI).

## DISCUSSION

### Distribution and abundance

The distribution of marsh deer was not uniform throughout the basin area. The northern block was more affected by human-related activities, which for the most part consisted of cattle ranching and clay exploitation. Hunting and diseases can also have an impact on marsh deer (SZABÓ *et al.* 2003, MACHADO *et al.* 2006, ARAÚJO JR *et al.* 2010, PIOVEZAN *et al.* 2010). The southern area was less impacted by human activities, most likely because the region is difficult for both humans and domestic animals to access due to the width of the marsh area (SZABÓ *et al.* 2003, 2007).

The estimated mean density prior to flooding was 0.48 deer/km<sup>2</sup> for the area around the Sergio Motta (Porto Primavera) Dam. According to a helicopter survey conducted in 1992, the marsh deer density was estimated to be 0.51 deer/km<sup>2</sup> (MOURÃO & CAMPOS 1995); in 1993, density was estimated to be 0.50-0.54 deer/km<sup>2</sup> for the same area (PINDER 1996). The gradual increase in water level jeopardized the availability of marsh deer habitat, constraining the animals to some small, fragmented patches of habitat. After flooding, there was a higher density of animals in some of the remaining areas of Paraná N, a fact that was expected given the population size and the reduction



Figures 5-6. Marsh deer distribution in the Paraná River basin before flooding (5) and after flooding (6); deer are represented by black dots. The grey polygon indicates the area affected by lake formation.

of habitat availability associated with the filling of the dam. However, the overall density per year was found to be stable, possibly because the habitat was at its maximum carrying capacity. A similar effect was documented for several species of birds that were confined to smaller territories with less food, which caused reduced nesting attempts and nesting success (FLESHMAN & KAUFMAN 1984).

Evasive movements and positive responses to the aircraft could have resulted in either the under- or overestimation of abundance, respectively. However, we did not notice a distinct response from the animals related to the aircraft and some sightings occurred when the animals were on the transect line.

The line-transect aerial survey methodology, performed in 1998 (ANDRIOLO *et al.* 2005), estimated the population to be 896 individuals (CV = 0.27). The abundance of marsh deer in the region surveyed for the present study prior to flooding was estimated to be 974 individuals (CV = 0.23). The difference in abundance can be accounted for by the standard error of both results. PINDER (1996), in 1992 and MOURÃO & CAMPOS (1995), in 1993, estimated population sizes of 950 and 650 individuals, respectively, for the same area; thus, this is the second largest population of this vulnerable species reported in Brazil.

### Population decline

The abundance results for the survey in February 1999 (0.37 deer/km<sup>2</sup>), after the flooding, was smaller than previous estimates, but not significantly so. A period parameter was incorporated into the analysis, but it did not have a significant

effect. The water level in the marsh area likely affected the visibility of the animals. Similarly, PINDER (1996) and ANDRIOLO *et al.* (2005) verified that the animals were more visible during the dry season than they were during the rainy season. MAURO *et al.* (1995) indicated that the deer have a marsh habitat preference related to the depth of the floodplain area. The depth of the water in lowland areas varies throughout the year and may affect the distribution of marsh deer. During the rainy season, the animals move closer to the higher, drier forest areas. Therefore, it becomes more difficult to see them from an observational aircraft.

During the phases occurring one (November 1999) and two (December 2000) years after flooding, the overall population density remained at the same level (0.47 and 0.42 deer/km<sup>2</sup>, respectively). The consistency of these density results from the years after flooding might indicate the carrying capacity of the environment. The density was expected to be higher in the remaining areas; however, our results indicate otherwise. An analysis of the direct impact on marsh deer by radio-tracking individuals in the flooded region has documented 20% of the direct deaths and 40% of the long-term deaths and animal emigration. Animals found outside the marsh area (DUARTE *et al.* 2003) were more vulnerable to hunters and in poor physical condition, likely because of a dearth of food and intraspecific agonistic interactions. Under these conditions, animals died primarily from diseases and starvation (DUARTE *et al.* 2003, SZABÓ *et al.* 2003, MACHADO *et al.* 2006, ARAÚJO JR. *et al.* 2010).

## Conservation

The overall abundance of marsh deer dropped from 974 to 444 individuals during the study period, a reduction of 54%, which can likely be attributed to the direct impact of the flooding process, exacerbated by indirect effects, such as increased disease, hunting and a reduction in food availability. Similar changes in abundance following habitat degradation have been observed in other species. Drastic reductions were documented for populations of the endangered southwestern Willow Flycatcher, *Empidonax traillii extimus* A.R. Phillips, 1948 (Passeriformes: Tyrannidae). The population of this bird species decreased 47%, from 209 territories in 2004 to 111 territories in 2006, in response to habitat changes (ELLIS *et al.* 2009).

The marsh deer case presented here reveals a sharp decline over the original area of occurrence, with habitat reduction being one of the greatest threats to the species in Brazil (WEMMER 1998). In this context, the construction of dams with large reservoirs for the production of electric power is undoubtedly a major contributor to the reduction of habitat because of the advancement of water onto the floodplains adjacent to the dammed watershed. Although habitat reduction is known to be the main threat to marsh deer, little is known about the ecology and behavior of this species in its natural environment. In addition, the ecological and behavioral changes following environmental changes from flooding caused by hydroelectric plants have not yet been assessed. The last large populations of marsh deer are found along the Araguaia, Tocantins and Guaporé Rivers, which are included in the Brazilian power matrix plan. The species has also been approaching extinction in the State of São Paulo (DUARTE & VOGLIOTTI 2009). Past conservation efforts have included the reintroduction of eight adult marsh deer in the Jataí Ecological Station, which were released back into the wild in December of 1998 (FIGUEIRA *et al.* 2005). All of the reintroduced animals came from the flooded areas of the Paraná River Basin.

The conservation outlook for the animals that survived the flooding does not look promising. The habitat remaining in the northern areas were considered to be too small to support a large population of marsh deer. Prior to flooding, the marsh deer population was distributed widely throughout the dam's catchment area; this habitat has now been almost completely destroyed from the lake filling process. The findings of this study highlight the need to implement management strategies that preserve the remaining fragmented populations. The distribution of marsh deer is more fragmented now than prior to the flooding and a non-equilibrium meta-population is expected for the marsh deer (VAN OORT *et al.* 2010). Further, the genetic information obtained from the current population reveals low genetic diversity (OLIVEIRA *et al.* 2005, 2009, MÁRQUEZ *et al.* 2006, LEITE *et al.* 2007), which may contribute to susceptibility to new selective pressures in the changing environment. Mountain caribou and other wide-ranging species that have been fragmented into subpopulations by human actions may appear to be part of

a meta-population, but unless they have the innate ability to disperse among subpopulations, the distribution is more likely akin to the geographic pattern of the extinction process. A meta-population strategy for the conservation of marsh deer must incorporate the remaining population of the Paraná River Basin and consider the genetic isolation of the remaining sub-populations. New research must be carried out to provide information on permit plan elaboration and decision strategies for marsh deer conservation and social/economic development.

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